

1 Stoichiometry among bioactive trace metals in the Chukchi and Beaufort Seas

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10 Abstract

11 The distribution of Al, Mn, Fe, Co, Ni, Cu, Zn, Cd and Pb in seawater was
12 investigated in the Chukchi and Beaufort Seas of the western Arctic Ocean in September
13 2000. The unfiltered and filtered seawater samples were used for determination of total
14 dissolvable metal (TDM) and dissolved metal (DM), respectively. The concentration of
15 labile particulate metal (LPM) was estimated with the difference between that of TDM and
16 DM. The concentrations of TDAI, TDMn, TDFe, TDCo and TDPb varied substantially in
17 the study area. The high concentrations occurred at stations near the Bering Strait, in the
18 Mackenzie delta, and above reductive sediments on the shelf and slope. These elements
19 were mostly dominated by labile particulate species, such as Fe-Mn oxides and species
20 adsorbed on terrestrial clay. DCo was correlated with DMn over the study area ($r = 0.78$, n
21 $= 135$), and the slope of the regression line was 27 times higher at a pelagic station than at a
22 shelf station. TDNi, TDCu, TDZn and TDCd showed relatively small variations and were
23 generally dominated by dissolved species. There was a moderate correlation between DCD
24 and phosphate for all samples ($r = 0.79$), whereas there were no significant correlation
25 between the other DMs and nutrients. TDNi and TDCu showed a remarkable linearity for

1 most stations except those near the Bering Strait ($R^2 = 0.95$, $n = 126$). These results suggest
2 that biogeochemical cycling including uptake by phytoplankton and remineralization from
3 settling particles has only minor control over the distribution of trace metals in this area.
4 Using the present data, the annual input of bioactive trace metals from the Bering Strait and
5 the Mackenzie River was estimated. Also the trace metal compositions of major water
6 masses were evaluated. The dissolved elemental ratio was P : Al : Mn : Fe : Co : Ni : Cu :
7 Zn : Cd = 1 : 1.2×10^{-2} : 4.4×10^{-4} : 1.4×10^{-3} : 3.7×10^{-5} : 3.7×10^{-3} : 1.4×10^{-3} : 4.5×10^{-3} :
8 2.2×10^{-4} for Canada Basin deep water (CBDW). This ratio was significantly different from
9 that for Pacific deep water and Bering Sea water, suggesting substantial modification of the
10 trace metal compositions of seawater in the study area.

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12 Keywords: Arctic Ocean, Chukchi Sea, Beaufort Sea, seawater, bioactive trace metals, total
13 dissolvable species, dissolved species, labile particulate species, speciation, stoichiometry.

14

15 1. Introduction

16 Bioactive trace metals, such as Al, Mn, Fe, Co, Ni, Cu, Zn, Cd and Pb, are essential
17 to organisms and/or highly toxic at a high concentration. In recent decades, studies in
18 chemical oceanography have revealed that these elements are important as limiting factors of
19 biological productivity and as tracers of biogeochemical processes (SCOR Working Group
20 2007; Sohrin and Bruland 2011). However, our knowledge is still limited on their global
21 distributions and time-dependent dynamics. This is especially true to the Arctic Ocean that
22 is characterized by ice cover and harsh weather. Moore (1981; 1983) has reported the first
23 vertical profiles of dissolved (D) Al, Fe, Cu, Zn and Cd in the central Arctic Ocean (Makarov
24 Basin), suggesting that the high surface concentrations of Cu, Zn and Cd were related to
25 contributions from surface run-off and from the underlying nutrient-rich Bering Sea winter

1 water. Danielson and Westerlund (1983) observed the vertical profiles of total dissolvable
2 (TD) Fe, Ni, Cu, Zn and Cd in the eastern Arctic Ocean. Mart et al. (1984) investigated the
3 total dissolvable concentrations of Ni, Cu, Cd and Pb in surface and deep water collected
4 from the eastern Arctic Ocean. Yeats and Westerlund (Yeats 1988; Yeats and Westerlund
5 1991) reported the concentrations of TDMn, TDCo, TDNi, TDCu, TDCd and DMn near the
6 Canadian Ice Island. Measures (1999) determined the distribution of reactive Al and Fe in
7 surface water along a section from the Canada Basin to the Eurasian Basin, suggesting that
8 ice-rafted sediment may be an important source for Al and Fe. Substantial advancements in
9 the study of Al, Mn and Fe were attained during an expedition of R.V. Polarstern in 2007.
10 Ocean transects of DAl were determined in the Eurasian part of the Arctic Ocean, showing a
11 strong correlation between DAl and Si (Middag et al. 2009). The elevated DMn in the
12 surface layer was related to fresh water input, and a deep maximum of DMn in the Nansen
13 Basin was ascribed to hydrothermal input over the Gakkel Ridge (Middag et al. 2011). The
14 distribution of DFe in the Arctic shelf seas (Barents, Kara and Laptev Seas) and in the surface
15 waters of the central Arctic Ocean revealed the impact of river water and ice-melt (Klunder et
16 al. 2012a), and that in deep water of the Nansen, Amundsen and Makarov Basins showed
17 input from the shelf seas and slopes, effect of hydrothermal activity, and scavenging in the
18 deepest part (Klunder et al. 2012b). Thuróczy et al. (2011) investigated the size-fractionated
19 speciation and ligand characteristics of Fe. Lastly, Nakayama et al. (2011) reported the
20 vertical distribution of DFe and TDFe throughout the water column in the western Arctic
21 Ocean (Chukchi Sea and Canada Basin).

22 We participated in the MR00-K06 cruise of R/V Mirai and collected clean seawater
23 samples for trace metal study in the Bering, Chukchi and Beaufort Seas in September 2000.
24 In previous work, we reported the distribution of dissolved, total dissolvable and labile
25 particulate (LP) species for Al, Mn, Fe, Co, Ni, Cu, Zn, Cd and Pb in the Bering Sea (Cid et al.

1 2011). Here we present their distribution in the Chukchi and Beaufort Seas to provide a
2 comprehensive view on biogeochemistry of the nine elements in the western Arctic Ocean.

3

4 2. Sample Collections and Methods

5 2.1. Sampling locations

6 Seawater samples for this study were obtained during the MR00-K06 cruise of R/V
7 Mirai from 2 stations (A001 and A002) near the Bering Strait, 2 stations (A005 and A009) in
8 the Barrow Canyon, 2 stations (A018 and A020) in the Mackenzie Trough, 6 stations (A003,
9 A016, A022, A039, A042 and A054) on the slope of the Beaufort Sea, and 1 station (A023) in
10 the Canada Basin (Fig. 1).

11

12 2.2. Water sampling and analytical methods

13 The water sampling and analytical methods are almost identical with that in the
14 previous work (Cid et al. 2011). Seawater samples were collected with Niskin-X bottles that
15 have been thoroughly cleaned with detergent and HCl and mounted on a CTD-rosette water
16 sampling system (General Oceanics), the frame of which was finished with epoxy paint.
17 Upon retrieval of the CTD-rosette water sampling system, the seawater samples were
18 transferred to 500 ml pre-cleaned low-density polyethylene bottles (LDPE, Nalge Nunc) on
19 the deck using a silicon tube and filling bell (Nalge Nunc) to avoid contamination of airborne
20 particles. The samples were immediately brought into a clean room laboratory of the vessel.
21 An aliquot of seawater (250 ml) was filtered through a 0.2 μm Nuclepore filter (Coaster)
22 using a closed filtration system and HCl (TAMAPURE AA-10, Tama Chemicals) was added
23 resulting in a final concentration of 0.01 M and pH 2.2. This subsample was used for the
24 determination of DMs. The other aliquot of seawater (250 ml) was not filtered and a mixed
25 acid was added to achieve a final concentration of 0.01 M HCl and 0.002 M HF

1 (TAMAPURE AA-10), and used for the determination of TDMs. The seawater samples
2 were stored at room temperature in our laboratory more than 8 years until analysis. Here we
3 defined the labile particulate metal (LPM) as the difference between that of TDM and DM,
4 which would contain species such as those adsorbed on clay minerals, iron and manganese
5 oxyhydroxides, and those incorporated in organisms (Ezoe et al. 2004).

6 The details of the analytical method for bioactive trace metals have been reported
7 elsewhere (Sohrin et al. 2008). Preconcentration of the trace metals was performed using the
8 chelating resin on which ethylenediaminetriacetic and iminodiacetic acids were immobilized
9 (NOBIAS CHELATE-PA1, Hitachi High-Technologies). Finally, the trace metals were
10 eluted with 15 ml of 1 M HNO₃ (TAMAPURE AA-10). Concentrations of DMs and TDMs
11 in the eluate were determined using an inductively coupled plasma mass spectrometer (Elan
12 DRC II, Perkin Elmer) by a calibration curve method. The isotopes used for the
13 determination were ²⁷Al, ⁵⁵Mn, ⁵⁴Fe, ⁵⁹Co, ⁶⁰Ni, ⁶⁵Cu, ⁶⁸Zn, ¹¹⁴Cd and ²⁰⁸Pb. Other isotopes
14 were also measured for cross checking except mono-isotopic Al, Mn and Co.

15 The detection limits by ICP-MS after preconcentration with a concentration factor
16 of 8 were 0.1 nmol kg⁻¹ Al, 0.05 nmol kg⁻¹ Mn, 0.03 nmol kg⁻¹ Fe, 6 pmol kg⁻¹ Co, 0.1 nmol
17 kg⁻¹ Ni, 0.03 nmol kg⁻¹ Cu, 0.3 nmol kg⁻¹ Zn, 0.005 nmol kg⁻¹ Cd, and 0.2 pmol kg⁻¹ Pb.
18 The average percent recovery with sd was typically 100 ± 15% (*n* = 3) for each metal on the
19 experiments using seawater with and without metal-spike. The average concentrations with
20 sd in seawater reference material NASS-5 (National Research Council of Canada) by this
21 method were 3.59 ± 0.25 nmol kg⁻¹ for Al, 13.0 ± 0.23 nmol kg⁻¹ for Mn, 2.85 ± 0.18 nmol
22 kg⁻¹ for Fe, 145 ± 5 pmol kg⁻¹ for Co, 4.72 ± 0.08 nmol kg⁻¹ for Ni, 4.49 ± 0.15 nmol kg⁻¹ for
23 Cu, 1.66 ± 0.09 nmol kg⁻¹ for Zn, 0.176 ± 0.001 nmol kg⁻¹ for Cd, and 35 ± 1 pmol kg⁻¹ for
24 Pb (*n* = 3). The results agreed with the certified values, while there was no certified value
25 for Al. Our method was also validated using the SAFe reference materials of seawater

1 (Sohrin et al. 2008).

2 Temperature (T) was measured with the CTD. Salinity (S) was determined by
3 conductivity on board the vessel (Participants of MR00-K06 2000). Dissolved oxygen was
4 determined by the Winkler method. Macronutrients were measured on board with an
5 AutoAnalyzer (TRAACSS). Chlorophyll *a* (Chl. *a*) was determined by fluorometry.
6 Ocean Data View (Schlitzer, R., <http://odv.awi.de>, 2012) was utilized for data analysis and
7 preparation some figures.

8

9 3. Results

10 3.1. Hydrography

11 Schematics of water masses and currents are shown in Fig. 1 (Grebmeier et al.
12 2006). Pacific water inflow from the Bering Sea consists of Alaska Coastal Water (ACW)
13 on the eastern side, Anadyr Water (AW) on the western side, and Bering Shelf Water (BSW)
14 lying between ACW and AW. As these waters flow north through the Bering Strait, AW
15 and BSW mix to form a modified Bering Shelf-Anadyr Water (BSAW). An annual mean
16 transport of the Pacific water inflow is assumed to be 0.8 Sv with salinity of 32.5 (Woodgate
17 and Aagaard 2005). This is an important source of heat, freshwater and nutrients into the
18 Arctic Ocean. Shelf transformation processes modify the physical and biogeochemical
19 properties of Pacific waters as they cross the Chukchi shelf. For the Chukchi and Beaufort
20 regions, current transport from the shelves to slope and deep basins occurs primarily through
21 the Barrow and Herald Canyons, with subsequent transport into the basin via currents flowing
22 eastward along the continental slope, eddies generated along the slope, and/or by the effects
23 of surface wind-forcing. The eastward flow consists of modified BSAW, ACW and Atlantic
24 Intermediate Water (AIW). The Mackenzie River is the largest river that flows to the
25 Beaufort Sea with a water discharge of 249-333 km³ y⁻¹ (Dittmar and Kattner 2003). In the

1 Canadian Basin, there are anticyclonic (westward) surface circulation and cyclonic deep
2 circulation (Jones et al. 1995). The Arctic Ocean is relatively well ventilated above the ridge
3 depth of the Lomonosov Ridge (Talley et al. 2011). Deep water in the Canadian Basin is
4 called Canadian Basin Deep Water (CBDW), which has a mean age of 450 years below the
5 ridge depth.

6 Figure 2 shows a potential temperature-salinity diagram for the stations except
7 A018 and A020. Salinity of surface water decreased to 9.8 at A018 and 17.8 at A020,
8 indicating input of freshwater from the Mackenzie River. According to Fig. 2, several water
9 masses can be identified (Aagaard et al. 1981; Codispoti et al. 2005; Wang et al. 2006); (1)
10 surface water (SW), extending from the surface to ~50 m, with salinities of 26-32 and
11 temperatures of -1 to 7°C, (2) cold and salty water from ~50 to ~200 m, named as the upper
12 halocline layer (UHL), with salinities of 32.5-33.6 and temperatures of -1.7 to -1.3°C, (3)
13 warmer and more saline water from ~150 to ~250 m, named as the lower halocline layer
14 (LHL), with salinities of 34.0-34.4 and temperatures of -0.9 to -0.5°C, (4) AIW from ~200 to
15 ~775 m, with salinities of 34.6-34.9 and temperatures of 0 to 0.9°C, and (5) CBDW, below
16 1000 m depth, with salinities of > 34.9 and temperatures of < 0°C.

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18 3.2. Distribution of bioactive trace metals near the Bering Strait

19 The data of trace metals are summarized in Appendix Table 1. Obviously
20 contaminated and/or questionable values were removed from the following discussion, taking
21 account of oceanographic consistency. The concentration ranges for TDM and DM in each
22 region are given in Table 1. The vertical profiles of TDMs at A001 and A002 are compared
23 with the profile of station BR013 (167°W, 64.0°N) in the northern Bering Sea (Cid et al.
24 2011) in Fig. 3. The vertical profiles of DMs and oceanographic parameters are presented in
25 Appendix Fig. 1. The concentrations of labile Al at 168.8°W, 67.8°N were 16-22 nM

1 (Measures 1999), which were slightly higher than those of our DA1. At stations A001 and
2 A002, labile particulate species occupied >94% for TDA1, >60% for TDMn, >99% for TDFe,
3 >44% for TDCo, and >45% for TDPb. This was similar to the situation observed on the
4 Bering Sea shelf (Cid et al. 2011). Most of TDMs and DMs gave higher concentrations at
5 A002 than those at BR013 and A001 as well as phosphate and dissolved inorganic nitrogen
6 (DIN; the sum of nitrate, nitrite and ammonium) that was dominated by ammonium (Fig. 3),
7 suggesting additional input from the underlying sediments in the Chukchi shelf.

8

9 3.3. Distribution of bioactive trace metals in the Barrow Canyon

10 Stations A009 and A005 in the Barrow Canyon were characterized by high
11 concentrations of Mn and Co (Table 1). LPMn accounted for 18-70% of TDMn, and LPCo
12 accounted for 7-81% of TDCo. TDA1 and TDFe were also dominated by labile particulate
13 species; LPA1 accounted for 56-99% of TDA1, and LPFe accounted for more than 93% of
14 TDFe. In contrast, TDNi, TDCu, TDZn and TDCd were dominated by dissolved species.
15 Figure 4 shows the sectional distributions of TDMs and potential temperature from the
16 Barrow Canyon to continental slope (stations A009, A005 and A003). The sectional
17 distributions of DMs and oceanographic parameters are presented in Appendix Fig. 2. TDA1,
18 TDMn, TDFe and TDCo were rich at shelf stations and rapidly decreased between A005 and
19 A003 (Fig. 4). TDA1 and TDFe showed a strong correlation ($r = 0.90$, $n = 34$), suggesting
20 dominance of terrestrial mineral particles. TDCo showed moderate correlations with TDMn
21 ($r = 0.80$) and TDFe ($r = 0.81$). DMn, DCo and ammonium showed moderate correlations
22 with each other ($r = 0.73-0.86$). This is ascribed to the effect of denitrification and
23 manganese reduction in sediments of the shelf region. High concentrations of DNi, DCu,
24 DZn and DCd were observed at 50-200 m depth at A003 in accordance with the maximums
25 of silicate, nitrate and phosphate. DNi and DCd showed moderate correlations with silicate

1 ($r = 0.71$) and phosphate ($r = 0.69$). Also DNi showed moderate to strong correlations with
2 DCd ($r = 0.87$) and DCu ($r = 0.77$). DZn did not show such a correlation. TDPb was
3 dominated by the labile particulate fraction and did not show a significant correlation with the
4 other parameters.

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6 3.4. Distribution of bioactive trace metals in the Mackenzie Trough

7 Al, Co, Ni, Cu, and Zn showed the highest concentrations at station A018 in the
8 Mackenzie Trough (Table 1). Figure 5 shows the sectional distributions of TDMs and
9 salinity from the Mackenzie Trough to continental slope (stations A018, A020 and A022).
10 The sectional distributions of DMs and oceanographic parameters are presented in Appendix
11 Fig. 3. The concentrations of TDMs except TDCd were elevated at A018, giving significant
12 negative correlations with S ($r = -0.63$ - -0.84 , $n = 25$) and positive correlations with each
13 other ($r = 0.71$ - 0.99). TDAI, TDMn, TDFe and TDCo were dominated by labile particulate
14 species, and their LPMs showed strong correlations ($r = 0.95$ - 0.98). This suggests that these
15 metals are mostly supplied as particulate forms from the Mackenzie River. DMn, DFe, DCo,
16 nitrite and ammonium also showed moderate correlations with each other ($r = 0.71$ - 0.88),
17 whereas there were no significant correlations with DAI. This means substantial amounts of
18 DMn, DFe and DCo are supplied through reductive dissolution from the sediments. TDNi,
19 TDCu, TDZn and TDCd were dominated by dissolved species. It is likely that these metals
20 are mainly supplied as dissolved species from the Mackenzie River. DNi and DCu showed
21 moderate correlations with DMn, DCo, nitrite and ammonium ($r = 0.74$ - 0.93), suggesting
22 DNi and DCu are also released from the sediments through manganese reduction. Only
23 DCd showed strong correlations with silicate and phosphate ($r = 0.82$ - 0.84). DZn showed
24 weak correlations with DCo, DNi and DCu ($r = 0.62$ - 0.69). There was a moderate
25 correlation between DPb and DFe ($r = 0.80$).

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3.5. Distribution of bioactive trace metals on the slope of the Beaufort Sea

At the stations on the slope of the Beaufort Sea, the concentrations for Al, Mn, Fe, Co and Zn were generally between those at shelf stations and those at a station in the Canada Basin (Table 1). The concentrations for Ni, Cu and Cd above the slope of the Beaufort Sea were comparable with those in the Canada Basin. The vertical profiles of TDMs at A054, A003, A016 and A022 are compared in Fig. 6 to examine east-west differences. The vertical profiles of DMs and oceanographic parameters are presented in Appendix Fig. 4. TDAI, TDMn, TDFe and TDCo showed maximums at ~225 m depth at station A054, which was within the UHL and just below the maximums of silicate, nitrate and phosphate at ~200 m depth. Nakayama et al. (Nakayama et al. 2011) reported the vertical profiles of TDFe and DFe at a slope station B2 (162°W, 74.5°N) in September 2008, which was located northwest of A054. The concentrations were 3.0-48 nM for TDFe and 0.50-1.3 nM for DFe with maximums at ~175 m depth, which were also located in the UHL and just below the maximums of silicate and phosphate at ~125 m depth. Thus the vertical profiles of TDFe and DFe were similar between station A054 in 2000 and station B2 in 2008, whereas the concentrations were significantly lower for B2. TDAI, TDMn, TDFe and TDCo showed three maximums at ~25, 200 and 500 m depths at A003. The middle depth maximums were located in the LHL. The middle depth maximums appeared at 120~140 m depths corresponding to the UHL at A016. Broad maximums of TDAI, TDMn, TDFe and TDCo occurred in the upper water column above 180 m depth at A022. These results indicate that the distribution of TDAI, TDMn, TDFe and TDCo is highly influenced by local inputs. Correlation was studied for 6 stations (the above 4 stations plus A042 and A039). At these stations strong correlations were observed between TDAI and TDFe ($r = 0.82, n = 67$) and between DMn and DCo ($r = 0.81$). The concentrations of DMn and TDCo were relatively

1 low below 180 m at A022 and comparable to those observed at the Canadian Ice Island in
2 June 1985 (Yeats and Westerlund 1991).

3 TDCd was dominated by DCd, which was strongly correlated with silicate and
4 phosphate ($r = 0.91-0.90$). The maximum concentration occurred at 70-225 m depths
5 corresponding to the UHL. The maximum concentrations were higher at western stations.
6 TDNi, TDCu and TDZn were also mostly dominated by dissolved species. DNi, DCu and
7 DCd showed moderate to strong correlations with each other ($r = 0.64-0.89$). However, the
8 surface depletion was generally weak for DNi and DCu compared to DCd and nutrients, and
9 the maximum concentrations of DNi and DCu were higher at eastern stations. The latter is
10 probably due to input from the Mackenzie River. Although Zn shows a nutrient type profile
11 in the Atlantic and Pacific Oceans as well as Ni and Cd, the profiles of Zn in this area were
12 substantially different from those of nutrients, Ni, Cu and Cd. The concentrations of TDNi,
13 TDCu, TDZn and TDCd were relatively low below 200 m at A022 and comparable to those
14 observed at the Canadian Ice Island in June 1985 (Yeats and Westerlund 1991).

15 TDPb did not show significant correlations with any parameters. TDPb was
16 generally dominated by dissolved species, whereas considerably high concentrations of LPPb
17 were observed at some depths in the upper water column (< 300 m).

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19 3.6. Distribution of bioactive trace metals in the Canada Basin

20 Generally, the bioactive trace metals showed the lowest concentrations at station
21 A023 in the Canada Basin (Table 1). The vertical profiles of TDMs and DMs at A023 are
22 plotted in Fig. 7. The vertical profiles of oceanographic parameters are presented in
23 Appendix Fig. 5. DAI increased with depth at A023. The profile was consistent with that
24 of DAI observed at LOREX station (Moore 1981) and in the Nansen, Amundsen and
25 Makarov Basins (Middag et al. 2009). However, there was a break on the DAI profile at

1 2000 m depth at A023, where there was a break on the profile of potential temperature.
2 Middag et al. (Middag et al. 2009) reported strong correlations between DAl and silicate for
3 intermediate (~150-2000 m) and deep (> 2000 m) waters. DAl at A023 is plotted against
4 silicate in Fig. 8. The upper 300 m water column contained high concentrations of silicate
5 and there was no correlation between DAl and silicate. The correlations between DAl and
6 silicate were as follows:

7 for intermediate (500-1750 m) water,

$$8 \quad [\text{DAl nmol kg}^{-1}] = 2.8 [\text{Si } \mu\text{mol kg}^{-1}] - 18.8 \quad R^2 = 0.96, n = 6$$

9 for deep (> 2000 m) water

$$10 \quad [\text{DAl nmol kg}^{-1}] = 3.3 [\text{Si } \mu\text{mol kg}^{-1}] - 30.7 \quad R^2 = 0.96, n = 5$$

11 The slope for intermediate water was higher than in the other basins (1.9-2.2) and that for
12 deep water was considerably lower than in the other basins (7.4-13.5) (Middag et al. 2009).
13 These results may be related to deep water circulation. The Canada basin is separated from
14 the Makarov Basin by the Mendeleev-Alpha Ridge by a ~2000 m sill depth. It is reported
15 that the Makarov Basin is well ventilated by dense shelf water originating in the Barents,
16 Kara and Laptev Seas, whereas the Canada basin is relatively more isolated from this
17 ventilation source (Swift et al. 1997). TDAI became systematically lower than DAl below
18 1500 m depth at A023. Such abnormal inversion was significant only at this station. It is
19 likely that TDAI in deep water samples was underestimated. For most TD samples, added
20 HF was consumed to dissolve silicate particles and to form hexafluorosilicate. At this
21 station, terrigenous silicate particles occurred at a low concentration. It may be possible that
22 the remaining fluoride ions formed the complex with Al and interfered with the
23 preconcentration.

24 TDMn, TDFe and TDCo showed maximums at ~175 m depth of the UHL and
25 decreased with depth. TDMn and TDFe were dominated by a labile particulate fraction

1 (>51% for Mn and >78% for Fe), whereas TDCo consisted of comparable amounts of
2 dissolved and labile particulate species. TDMn showed strong correlations not only with
3 TDFe and TDCo but also with TDNi, TDCu and TDCd ($r = 0.81-0.93$, $n = 23$). LPMn,
4 LPFe and LPCo also showed strong correlations each other ($r = 0.86-0.95$). DMn gave a
5 surface maximum at 25 m depth and decreased with depth. The profile was consistent with
6 observations in the Nansen, Amundsen and Makarov Basins (Thuróczy et al. 2011). The
7 average TDFe concentration with sd was 14 ± 2 nmol kg⁻¹ in CBDW at A023. This value
8 was higher than the TDFe concentration of ~ 8 nmol kg⁻¹ in CBDW in the western Canada
9 Basin (Nakayama et al. 2011) and 1.5-5.5 nmol kg⁻¹ in the Nansen, Amundsen and Makarov
10 Basins (Middag et al. 2011). It is likely that the long storage of acidified samples with a
11 small amount of HF until preconcentration resulted in more complete dissolution of LPFe in
12 this work. DFe showed a relatively uniform profile at A023, and the baseline concentration
13 was similar to that in the western Canada Basin (Nakayama et al. 2011) and in the other
14 basins of the Arctic Ocean (Klunder et al. 2012b; Thuróczy et al. 2011). The profile of DCo
15 was mostly similar to that of DMn ($r = 0.83$). DCo also showed strong correlations with
16 DNi and DCu ($r = 0.87-0.92$). The break at 2000 m depth on the profile was observed for
17 DMn, DFe and DCo as well as DA1. A broad maximum of DMn occurred below 2000 m.
18 DFe and DCo gave a maximum above 2000 m. Another maximum of DFe was observed at
19 ~ 2500 m.

20 TDNi, TDCu and TDCd showed similar profiles with phosphate and silicate.
21 Although phosphate and silicate increased with depth in deep water, the concentrations of
22 TDNi, TDCu and TDCd were more constant. TDNi, TDCu and TDCd were dominated by
23 dissolved species (>89% for Ni, >82% for Cu, and >79% for Cd). DCd showed strong
24 correlations with phosphate and silicate ($r = 0.88-0.89$). The correlations for DNi and DCu
25 with phosphate and silicate were moderate ($r = 0.66-0.77$). DNi, DCu and DCd were

1 strongly correlated with each other ($r = 0.85-0.90$). The concentration of TDNi in deep
2 water was similar to that observed in the eastern Arctic Ocean ($3.7-3.9 \text{ nmol kg}^{-1}$) (Mart et al.
3 1984). The profiles of DCu and DCd accorded with those observed at LOREX station
4 (Moore 1981). Although TDZn was also dominated by dissolved species ($>75\%$), it showed
5 a vertical profile distinct from the other trace metals. It was characterized by broad
6 maximums in shallow ($< 150 \text{ m}$) and middle ($1250-1500 \text{ m}$) depths. The reason for the
7 decoupling between Zn and nutrients is not clear now. The concentration of DZn at A023
8 was similar to that at LOREX station (Moore 1981). TDPb and DPb were generally low
9 throughout the water column. The relatively high variations might have been caused by
10 problems in sampling and/or analysis. The concentrations of TDPb were within the range of
11 $15-72 \text{ pmol kg}^{-1}$ reported for the eastern Arctic Ocean (Mart et al. 1984).

12

13 4. Discussion

14 4.1. Regional variation of bioactive trace metals

15 The ranges of the TDM concentration at each station are plotted in Fig. 9. The
16 average TDM/DM ratio at each station is plotted in Fig. 10. It is obvious that there are
17 considerable variations in the concentrations of TDMs in the study area. Particularly, TDAl,
18 TDMn, TDFe, TDCo and TDPb showed wide variations. The average TDM/DM ratios of
19 these metals were generally high and also showed considerable spatial variations: 1.6-414 for
20 Al, 1.8-28 for Mn, 29-873 for Fe, 1.4-2.2 for Co and 1.1-4.8 for Pb. These results indicate
21 that substantial amounts of these metals are supplied from localized sources and that species
22 of the input are mostly in a labile particulate fraction, such as terrigenous clay minerals,
23 and/or transformed promptly into labile particulate species, such as authigenic Fe-Mn oxides.
24 The major local sources are the Pacific water that flows through the Bering Strait (A001, Fig.
25 3), emission from reductive shelf sediments (A009, Fig. 4), and the Mackenzie River water

1 (A018, Fig. 5), as described in the results section. Most of the input is removed from
2 seawater within the shelf and slope regions, resulting in low concentrations at a pelagic
3 station A023 (Fig. 7). Such removal is consistent with that described in literature for Fe
4 (Klunder et al. 2012a; Nakayama et al. 2011; Thuróczy et al. 2011), for Mn (Middag et al.
5 2011) and for ^{210}Pb (Lepore et al. 2009). TDNi, TDCu and TDZn also gave relatively high
6 concentrations at A001, A002 and A018, suggesting the importance of input from the Bering
7 Strait and the Mackenzie River (Figs. 3, 5 and 9). The effect of local sources was not clear
8 for TDCd. It is probable that the distribution of Cd is primarily controlled by uptake and
9 remineralization through the biogeochemical cycle. The correlation between Cd and
10 nutrients will be discussed in the section 4.3. The TDM/DM ratio for Ni, Cu, Zn and Cd
11 were close to 1 except at A001, A002 and A018 (Fig. 10). Thus, Ni, Cu, Zn and Cd are
12 dominated by dissolved species in this area. Especially, DCd represented 95% of TDCd in
13 average. This result is clearly different from the observations on the Bering Sea shelf, where
14 the labile particulate fraction accounts for 20-80% of TDNi, TDCu, TDZn and TDCd (Cid et
15 al. 2011). This is probably due to the high concentrations of TDFe (3500 nmol/kg in
16 average) and TDMn (80 nmol/kg in average) and the adsorption of trace metals on Fe-Mn
17 oxides in the Bering Sea.

18

19 4.2. Input of bioactive trace metals from the Bering Strait and the Mackenzie River

20 Among the three sources mentioned above, the Pacific inflow water and the
21 Mackenzie River water carry new import of bioactive trace metals to the study area. Here
22 we estimate amounts of their annual input. The concentration of TDMs in the Pacific inflow
23 water can be expressed by the average concentration at A001. The water inflow is estimated
24 to be $25,000 \text{ km}^3 \text{ y}^{-1}$ (Woodgate and Aagaard 2005). To evaluate the concentration of
25 TDMs in the Mackenzie River water, we can use linear regression between TDM and S at

1 A018. Good linearity was observed for each TDM ($R^2 > 0.90$) except TDCd ($R^2 = 0.18$).
2 Annual water discharge is assumed to be $290 \text{ km}^3 \text{ y}^{-1}$ for the Mackenzie River (Dittmar and
3 Kattner 2003). The resulting concentration and annual input of TDMs for the Pacific inflow
4 water and the Mackenzie River water are summarized in Table 2. The TDM concentration is
5 higher in the Mackenzie River water than in the Pacific inflow water by a factor of 1.2-7.3.
6 However, the annual input from the Mackenzie River is only 1-8% of that from the Bering
7 Strait. The figures in Table 2 are just a first approximation, since it is based on only single
8 observation in summer. In the Lena River (Arctic Siberia), significantly high concentrations
9 of DMs were observed during the spring high flow (Hölemann et al. 2005). It seems likely
10 that there would be a similar seasonal variation in the Mackenzie River. In addition, we do
11 not know the seasonal change in concentration of trace metals in the Pacific inflow water.
12 We think, however, these figures may be useful, since they are the first estimation of annual
13 input of TDMs by the Pacific inflow water and the Mackenzie River water.

14

15 4.3. Stoichiometry of bioactive trace metals

16 Here we examine the relationship among bioactive trace metals and oceanographic
17 parameters over the study area. Correlation matrixes for TDMs, DMs and LPMs for all
18 samples are given in Appendix Tables 2-4. TDNi and TDCu showed the highest correlation
19 coefficient among TDMs ($r = 0.93$, $n = 136$). DNi and DCu also showed the highest
20 correlation coefficient among DMs for all samples ($r = 0.84$). TDNi, TDCu, TDZn and
21 TDAI showed moderate to strong correlations with each other ($r = 0.69$ - 0.93). LPNi and
22 LPZn also showed moderate correlations with LPAI ($r = 0.68$ - 0.79). Ni, Cu and Zn did not
23 show significant correlations with nutrients and Chl. *a*. A remarkably high linearity was
24 observed between TDCu and TDNi for stations except those near the Bering Strait, where
25 TDCu was relatively elevated (Fig 11a):

1
$$[\text{TDCu nmol kg}^{-1}] = 0.68 [\text{TDNi nmol kg}^{-1}] - 0.84 \quad R^2 = 0.95, n = 126$$

2 A similar high linearity was observed on the Bering Sea shelf with a higher slope (Cid et al.
3 2011):

4
$$[\text{TDCu nmol kg}^{-1}] = 0.92 [\text{TDNi nmol kg}^{-1}] - 2.19 \quad R^2 = 0.97, n = 44$$

5 These results suggest dominant effect of physical transport and chemical reactions, such as
6 redox and scavenging, on the distribution of Ni, Cu and Zn in these areas (Figs. 3-7). It is
7 likely that differentiation between Ni and Cu was weak, because biogeochemical cycling was
8 less important in controlling the distribution.

9 In contrast, DCd showed moderate correlations with phosphate ($r = 0.79$) and
10 silicate ($r = 0.77$). A regression line between DCd and phosphate was calculated for all
11 samples as follows (Fig 11b):

12
$$[\text{DCd nmol kg}^{-1}] = 0.33 [\text{PO}_4 \mu\text{mol kg}^{-1}] + 0.050 \quad R^2 = 0.62, n = 134$$

13 The slope was significantly higher than that of the regression line observed on the Bering Sea
14 shelf (Cid et al. 2011):

15
$$[\text{DCd nmol kg}^{-1}] = 0.19 [\text{PO}_4 \mu\text{mol kg}^{-1}] + 0.185 \quad R^2 = 0.74, n = 43$$

16 Cullen (2006) proposed that preferential uptake of DCd in Fe-limited surface water would
17 cause a kink in the DCd-P relationship. Our results are consistent with this model: these
18 areas were Fe-sufficient and there were no kinks in the DCd-P relationship. DIN is plotted
19 against DCd in Fig. 11c. There are two different trends in this figure for low salinity SW
20 and for high salinity AIW and CBDW. These characteristics are similar to those observed
21 between nitrate and phosphate (Jones et al. 1998) or between DIN and phosphate
22 (Yamamoto-Kawai et al. 2008). The two trends between nitrogen vs. phosphorus were
23 attributed to different origins, namely Pacific and Atlantic waters. The two trends between
24 DIN and DCd may be explained similarly. The regression line is as follows for SW within
25 the range of $30 < S < 33$:

1
$$[\text{DIN } \mu\text{mol kg}^{-1}] = 25 [\text{DCd nmol kg}^{-1}] - 5.7 \quad R^2 = 0.50, n = 63$$

2 Although it is difficult to establish a regression line for AIW and CBDW because of
3 insufficient data, the slope looks like higher than observed in SW, similar to the
4 nitrogen-phosphorus relationship (Jones et al. 1998; Yamamoto-Kawai et al. 2008). The
5 following regression was observed on the Bering Sea shelf (Cid et al. 2011):

6
$$[\text{DIN } \mu\text{mol kg}^{-1}] = 57 [\text{DCd nmol kg}^{-1}] - 18.5 \quad R^2 = 0.69, n = 43$$

7 These results suggest that DCd behaves similarly with phosphate in this area as well as in the
8 other oceanic regimes (Cullen 2006; de Baar et al. 1994). It is likely that biogeochemical
9 cycling is an important factor controlling the distribution of Cd. The stoichiometry between
10 Cd and nutrients, however, is distinct between the Arctic Ocean and the Bering Sea.

11 Moderate correlations occurred among TDFe, TDCo and TDAI ($r = 0.70-0.85$).
12 Moderate correlations were also found among DMn, DFe, DCo and ammonium ($r =$
13 $0.63-0.78$). Although DFe shows positive correlations with nutrients in the oceans (Johnson
14 et al. 1997; Turner and Hunter 2001), there were no significant correlations between DFe and
15 nutrients in this area. DCo is plotted against DMn in Fig. 11d. In this case, there is
16 difference in stoichiometry between the Chukchi shelf and pelagic stations. The regression
17 line for A009 is as follows:

18
$$[\text{DCo pmol kg}^{-1}] = 3.4 [\text{DMn nmol kg}^{-1}] + 340 \quad R^2 = 0.97, n = 11$$

19 The regression line for A023 is as follows:

20
$$[\text{DCo pmol kg}^{-1}] = 91 [\text{DMn nmol kg}^{-1}] + 28 \quad R^2 = 0.69, n = 23$$

21 In the other oceans, DCo takes different distribution from DMn due to biological uptake and
22 remineralization from settling particles in depth in addition to scavenging (Noble et al. 2008).
23 It is likely that biological uptake and remineralization are less significant for Co in the Arctic
24 Ocean, resulting in the strong correlations with Mn. The former regression should be
25 attributed to reductive dissolution-oxidative removal in the shelf and slope area (Appendix

1 Figs. 2 and 4), and the latter regression should be attributed to scavenging in CBDW (Fig. 7),
2 which has a mean age of about 450 years (Talley et al. 2011). Similarly the biogeochemical
3 cycling has only minor effect on the distribution of Fe.

4 For all samples, LPPb showed weak correlations with LPFe ($r = 0.66$) and LPCo (r
5 $= 0.60$), while TDPb and DPb did not show a significant correlation with the other parameters.
6 This may be ascribed to the effect of aeolian supply of anthropogenic Pb.

7 The average composition of four water masses, UHL, LHL, AIW and CBDW, is
8 summarized in Table 3. TDAI, TDPb, and DPb in CBDW were removed from the following
9 discussion due to the above-mentioned problems on data quality. SW was not included here,
10 because its salinity and concentrations of constituents varied substantially. Using these data,
11 we have calculated the DM/P and nutrient/P ratio for each water mass and compared the
12 results with that for North Pacific deep water and Bering Sea water (Fig. 12). The ratio for
13 the North Pacific deep water was calculated using our unpublished data. The ratio for the
14 Bering Sea water was calculated using data for BR013 (Cid et al. 2011), which was almost
15 identical with that for A001. It is apparent that the variation of DM/P is distinct from Si/P
16 and DIN/P. The Bering Sea water is enriched with trace metals especially with Al, Mn, Fe,
17 Co and Pb compared to the Pacific water. Generally DAI/P, DMn/P, DCo/P and DFe/P in
18 the Arctic waters are between the ratios for the Bering and Pacific waters. DAI/P and
19 DMn/P vary substantially among the water masses in the Arctic Ocean. DNi/P, DCu/P,
20 DZn/P and DCd/P in the Arctic waters are comparable to those in the Pacific water, whereas
21 DPb/P in the Arctic waters is close to that in the Bering water. The dissolved elemental ratio
22 is $P : Al : Mn : Fe : Co : Ni : Cu : Zn : Cd = 1 : 1.2 \times 10^{-2} : 4.4 \times 10^{-4} : 1.4 \times 10^{-3} : 3.7 \times 10^{-5} :$
23 $3.7 \times 10^{-3} : 1.4 \times 10^{-3} : 4.5 \times 10^{-3} : 2.2 \times 10^{-4}$ for CBDW. This ratio is significantly different
24 from that for the Pacific and Bering waters. Bruland et al. (1991) proposed an approximate
25 Redfield-type elemental composition of plankton organic tissue of $P : Fe : Zn : Cu, Mn, Ni,$

1 Cd = 1 : 0.005 : 0.002: 0.0004. Compared to this requirement, CBDW is poor in DFe and
2 DCd. According to Twining et al. (2011), Fe : Co ratio was 10-100 : 1 in plankton. Thus,
3 DCo could be also insufficient, when macro- and micro-nutrients in CBDW are utilized by
4 phytoplankton according to the reported stoichiometry.

5

6 5. Conclusions

7 The trace metal composition of Pacific inflow water through the Bering Strait is
8 considerably different from that of Pacific deep water. The composition is further modified
9 in the Chukchi and Beaufort Seas. Reductive sediments on the shelf or slope and the
10 Mackenzie River supply a large amount of Al, Mn, Fe and Co, resulting in high loads of
11 labile particulate species in seawater. Although correlation between DCd and phosphate is
12 consistent with the other oceans, this area is unique in weak or insignificant correlations
13 between the other DMs and nutrients. Some behaviors of trace metals are also unique in this
14 area: remarkable correlations between TDNi and TDCu or between DCo and DMn, and
15 decoupling of DZn from nutrients. These results suggest that biogeochemical cycling is not
16 a dominant factor controlling the distribution of trace metals in this area. Probably physical
17 transport and chemical reactions, such as redox and scavenging, are more important. Further
18 study is necessary to quantify the mechanism controlling the distribution of trace metals.

19

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4

5

1 Figure captions

2 Figure 1. Sampling locations in the western Arctic Ocean. Schematics of water masses
3 and currents are also shown using arrows: dark blue, Alaska Coastal Water (ACW); purple,
4 Bering Shelf-Anadyr Water (BSAW); pale blue, Siberian Coastal Current; yellow, Atlantic
5 Intermediate Water (AIW, subsurface); red, Beaufort Gyre (surface).

6 Figure 2. Potential temperature-salinity diagram: ○, A001; ●, A002; □, A003; ■, A005;
7 ◇, A009; ◆, A016; △, A022; ▲, A023; ⊕, A039; ×, A042; ▲, A054.

8 Figure 3. Vertical distributions of TDMs at A001 (●), A002 (◆) and BR013 (167°W,
9 64.0°N; △) in the northern Bering Sea (Cid et al. 2011).

10 Figure 4. Sectional distributions of TDMs and potential temperature in the Barrow Canyon.

11 Figure 5. Sectional distributions of TDMs and salinity in the Mackenzie Trough.

12 Figure 6. Vertical profiles of TDMs on the slope of the Beaufort Sea: ○, A054; ■, A003;
13 ◇, A016; ▲, A022.

14 Figure 7. Vertical profiles of TDMs (○) and DMs (●) at A023.

15 Figure 8. Plot of DAI versus silicate at A023: ○, upper water (< 300 m); ●, intermediate
16 water (500-1750 m); □, deep water (> 2000 m). The formulas of regression lines are
17 presented in the text.

18 Figure 9. Concentration ranges of TDMs at each station. The top and bottom bars
19 represent the maximum and minimum values, respectively.

20 Figure 10. Variations in the average TDM/DM ratio at each station: ○, Al; □, Mn; ◇, Fe;
21 △, Co; ●, Zn; ■, Cd; ◆, Pb.

22 Figure 11. (a) TDCu vs. TDNi. The red line presents the regression line for stations except
23 A01 and A02. (b) DCd vs. phosphate. The red line presents the regression line for all
24 samples. (c) DIN vs. DCd. The red line presents the regression line for SW (30 < S < 33)
25 and the blue broken line presents a speculated relationship for Atlantic water. (d) DCo vs.

1 DMn. The red and blue lines present the regression lines for A009 and A023, respectively.
2 The color of the dots represents salinity. The formulas of regression lines are presented in the
3 text.

4 Figure 12. The average DM/P and nutrient/P ratio of each water mass: ○, UHL; ●, LHL;
5 □, AIW; ■, CBDW; △, Bering Sea; ▲, North Pacific Deep Water.

6

7

8 Appendix Figure captions

9 Appendix Figure 1. Vertical distributions of DMs and oceanographic parameters at A001
10 (●), A002 (◆) and BR013 (167°W, 64.0°N; △) in the northern Bering Sea (Cid et al. 2011).

11 Appendix Figure 2. Sectional distributions of DMs and oceanographic parameters in the
12 Barrow Canyon.

13 Appendix Figure 3. Sectional distributions of DMs and oceanographic parameters in the
14 Mackenzie Trough.

15 Appendix Figure 4. Vertical profiles of DMs and oceanographic parameters on the slope of
16 the Beaufort Sea: ○, A054; ■, A003; ◇, A016; ▲, A022.

17 Appendix Figure 5. Vertical profiles of oceanographic parameters at A023.

18

19

20

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Figure

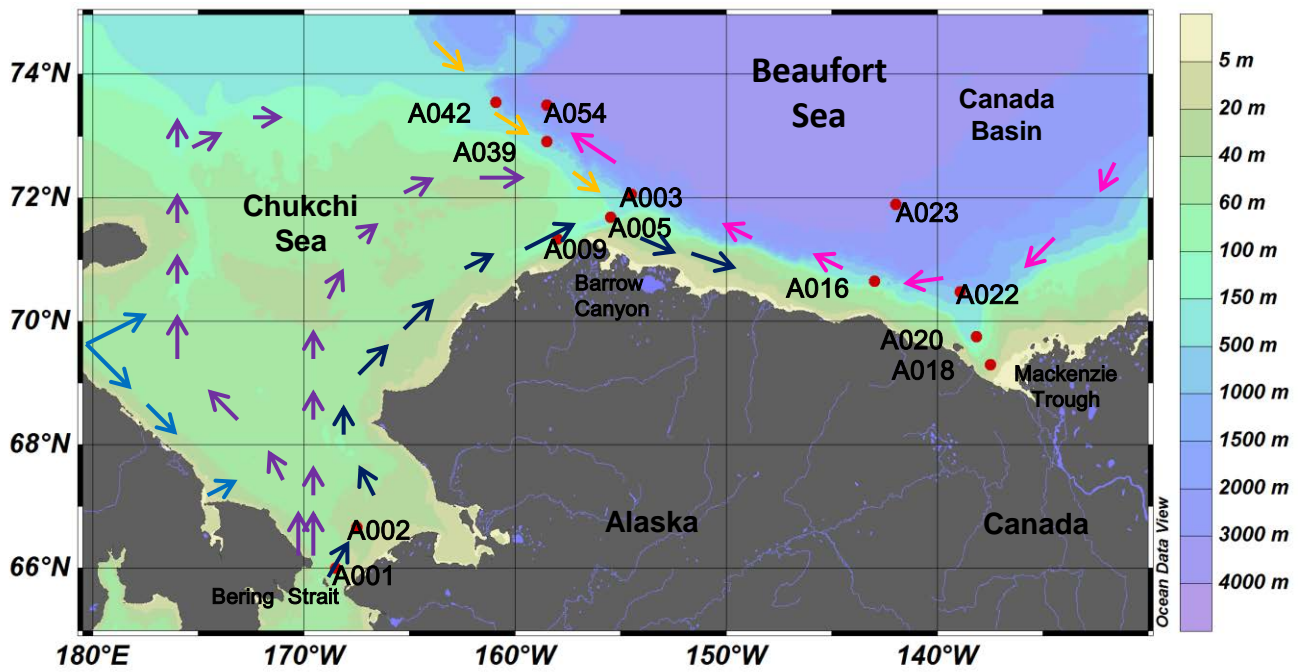


Figure 1

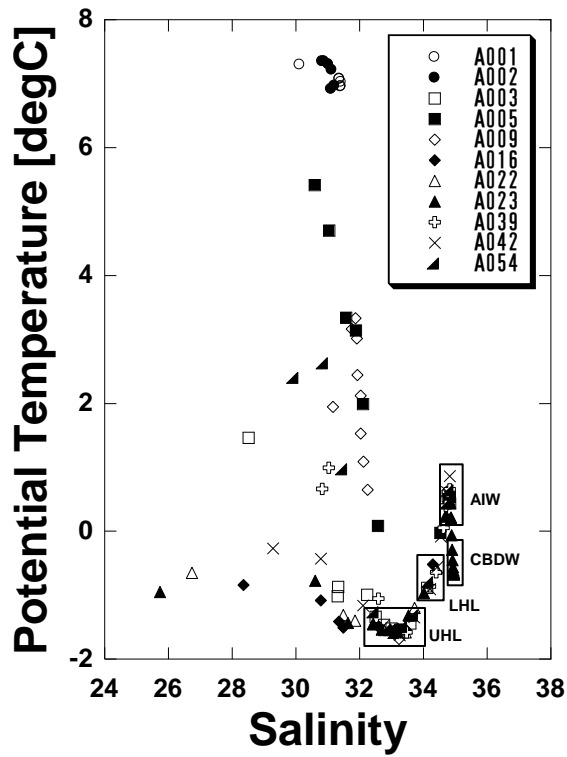


Figure 2

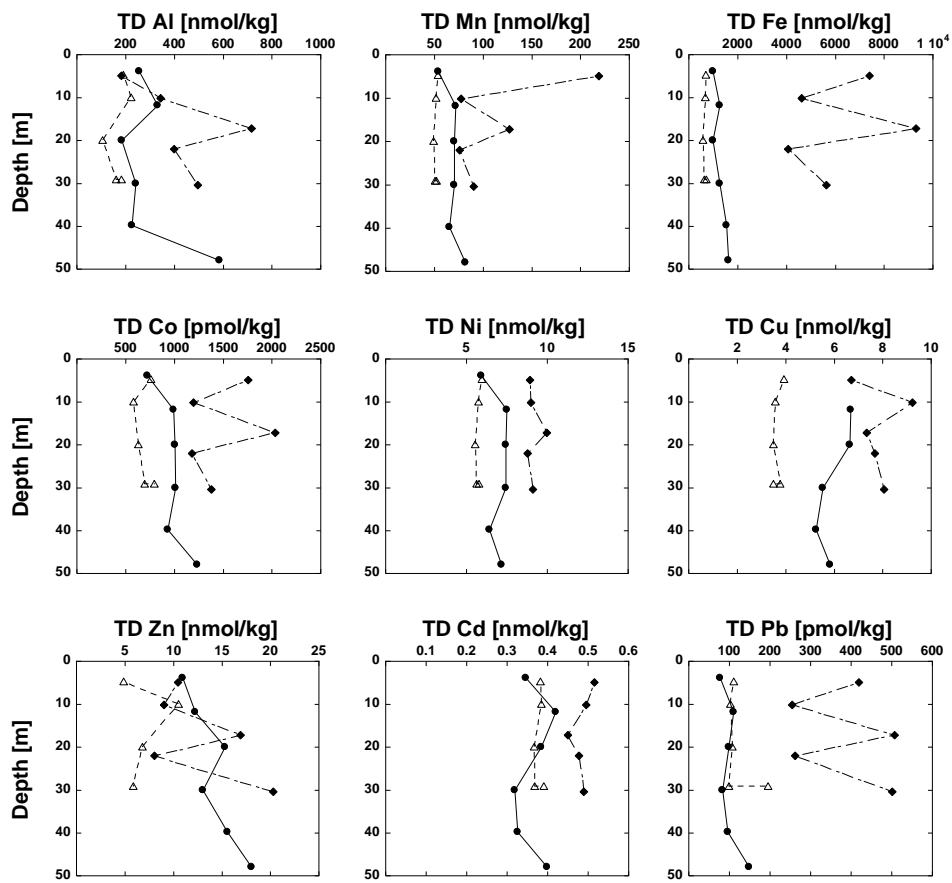


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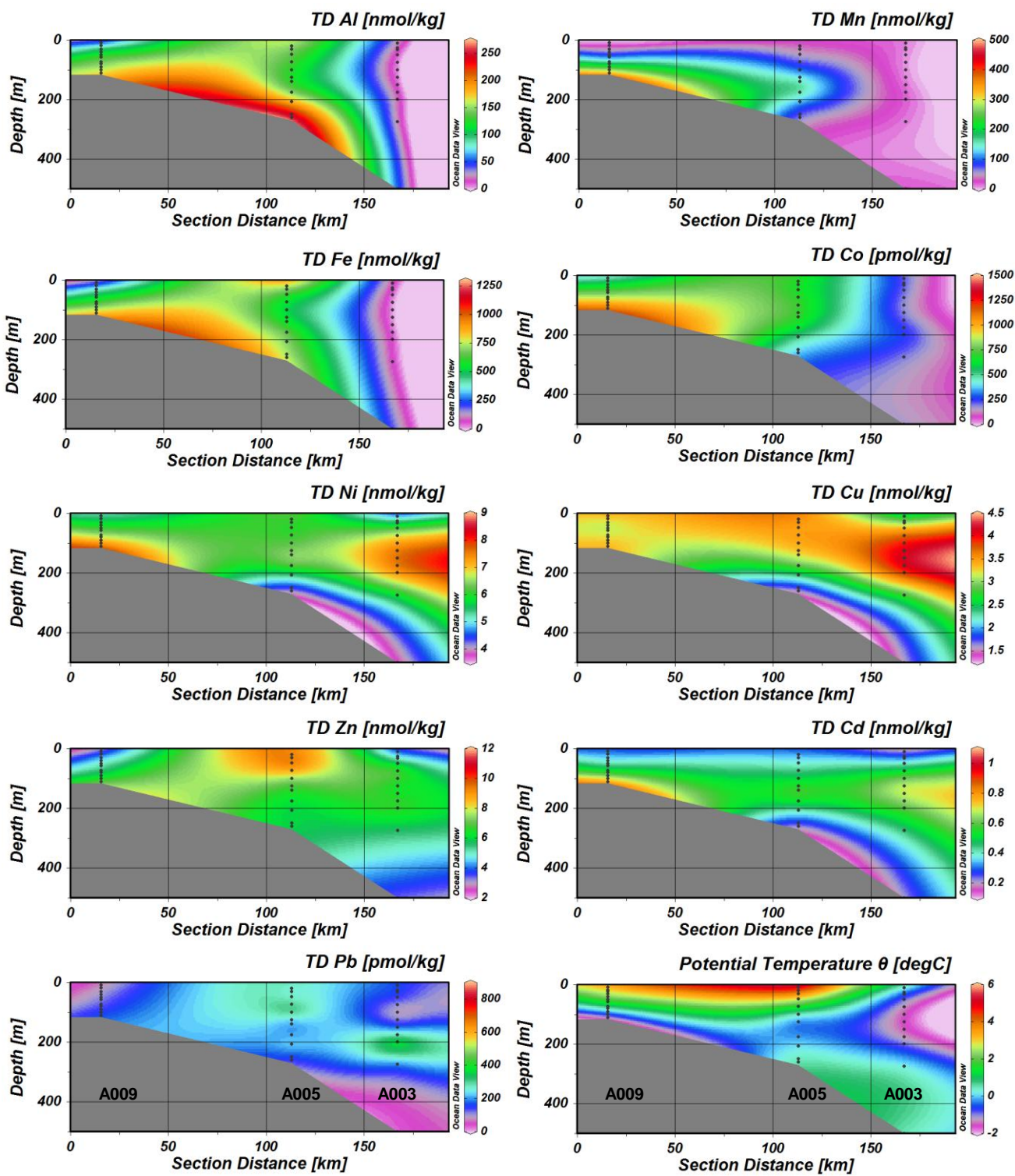


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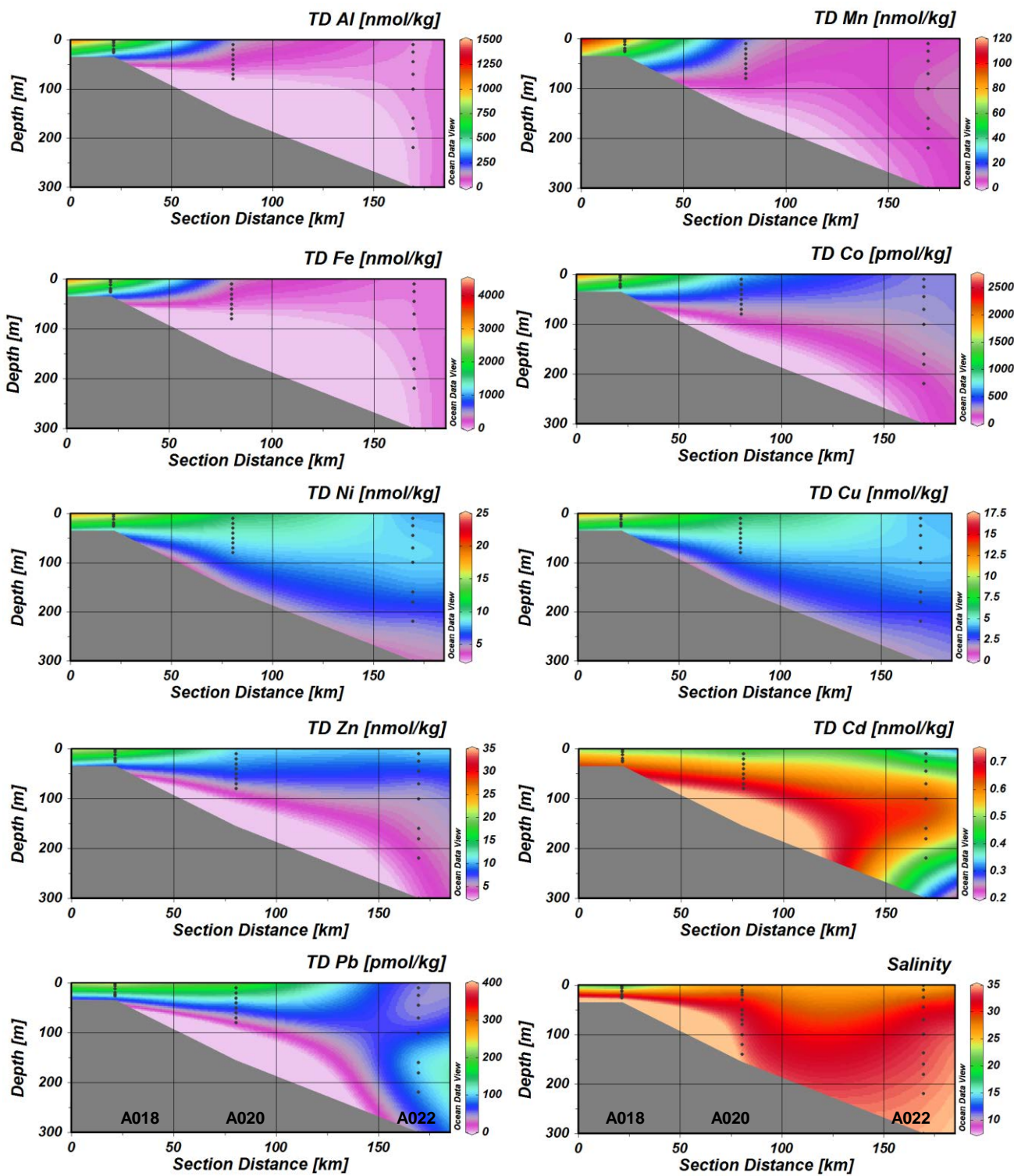


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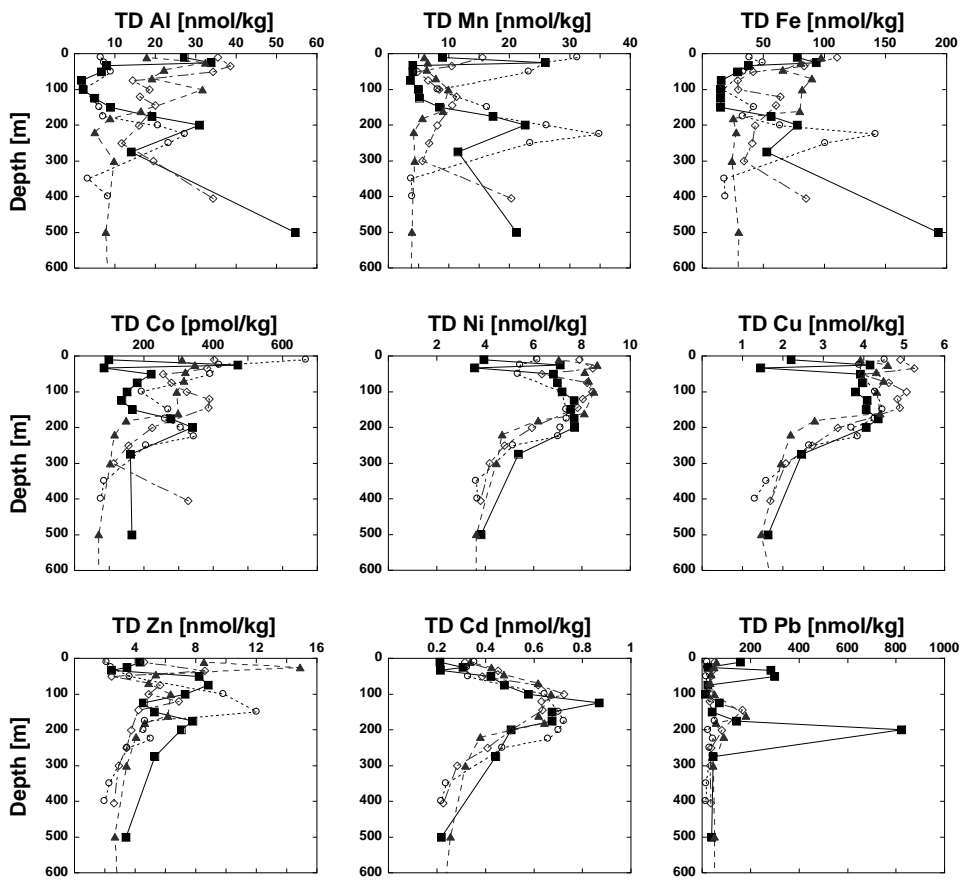


Figure 6

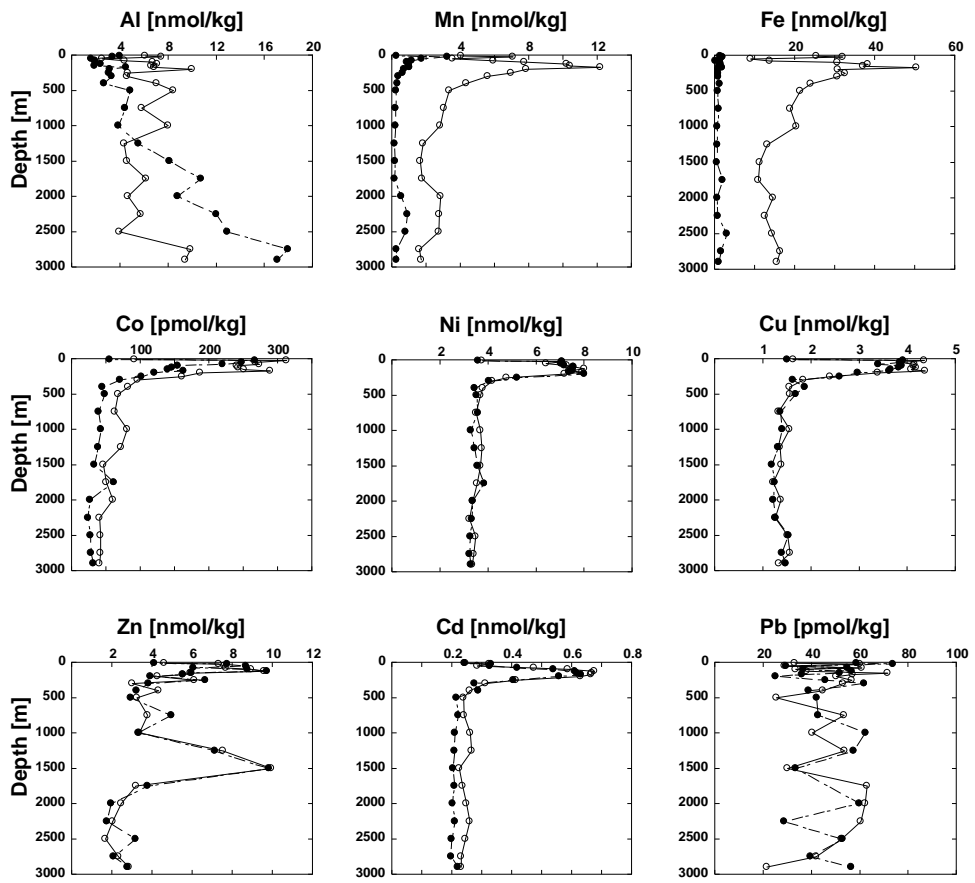


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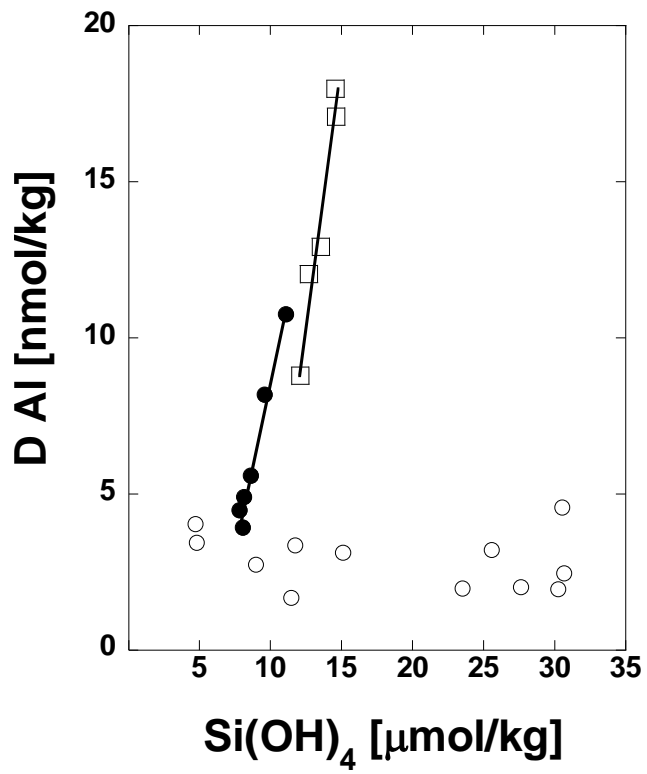


Figure 8

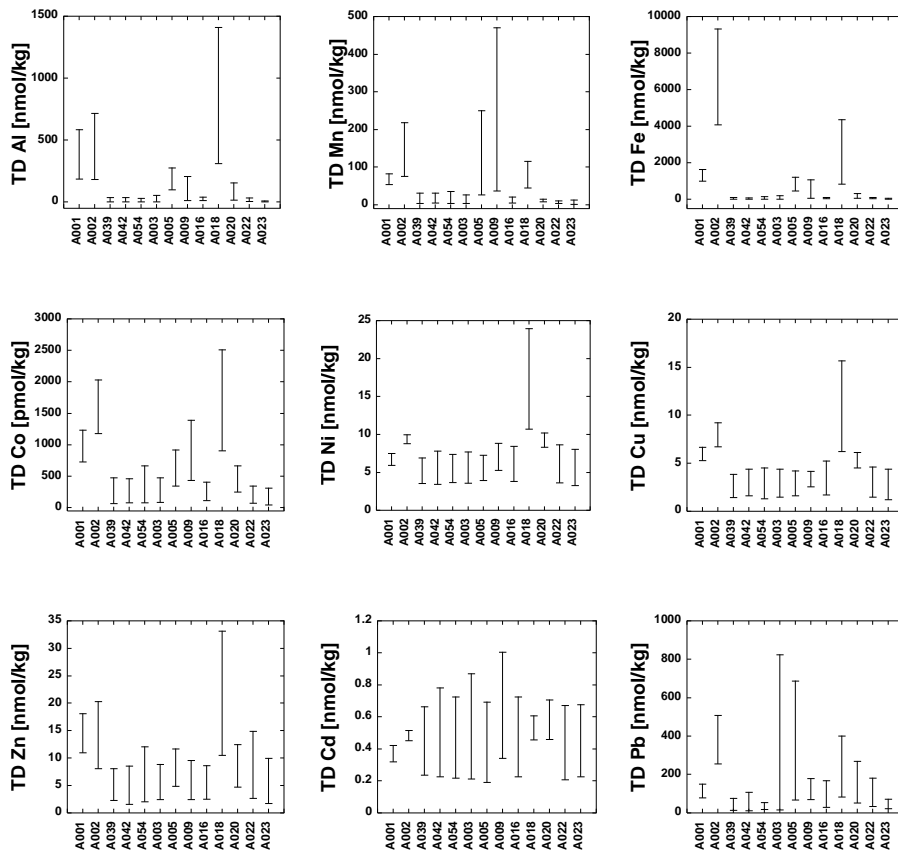


Figure 9

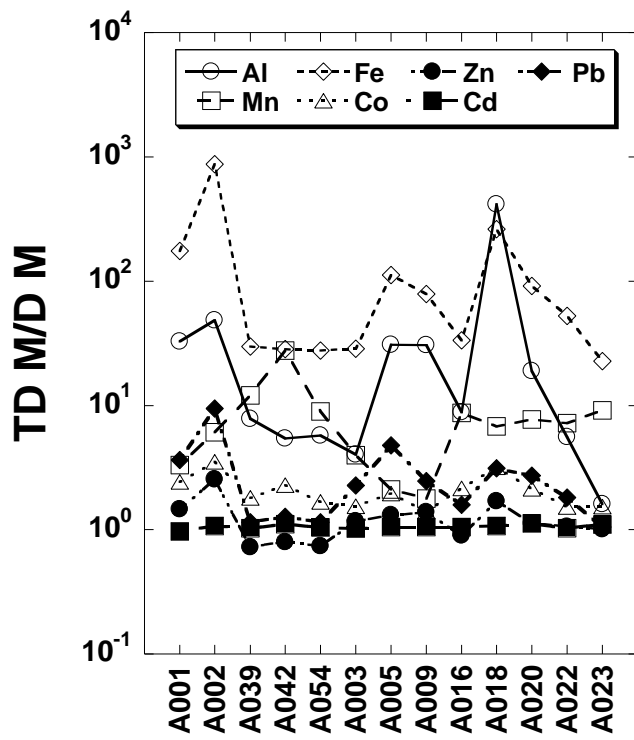


Figure 10

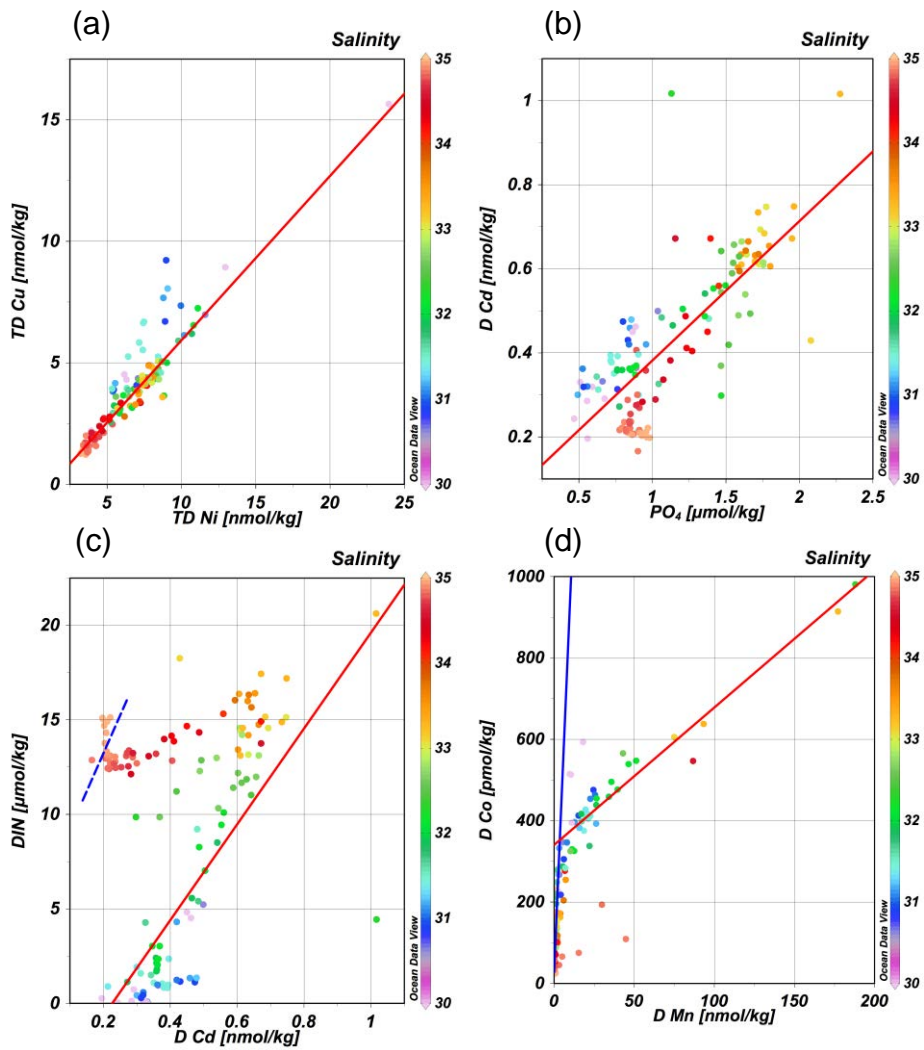


Figure 11

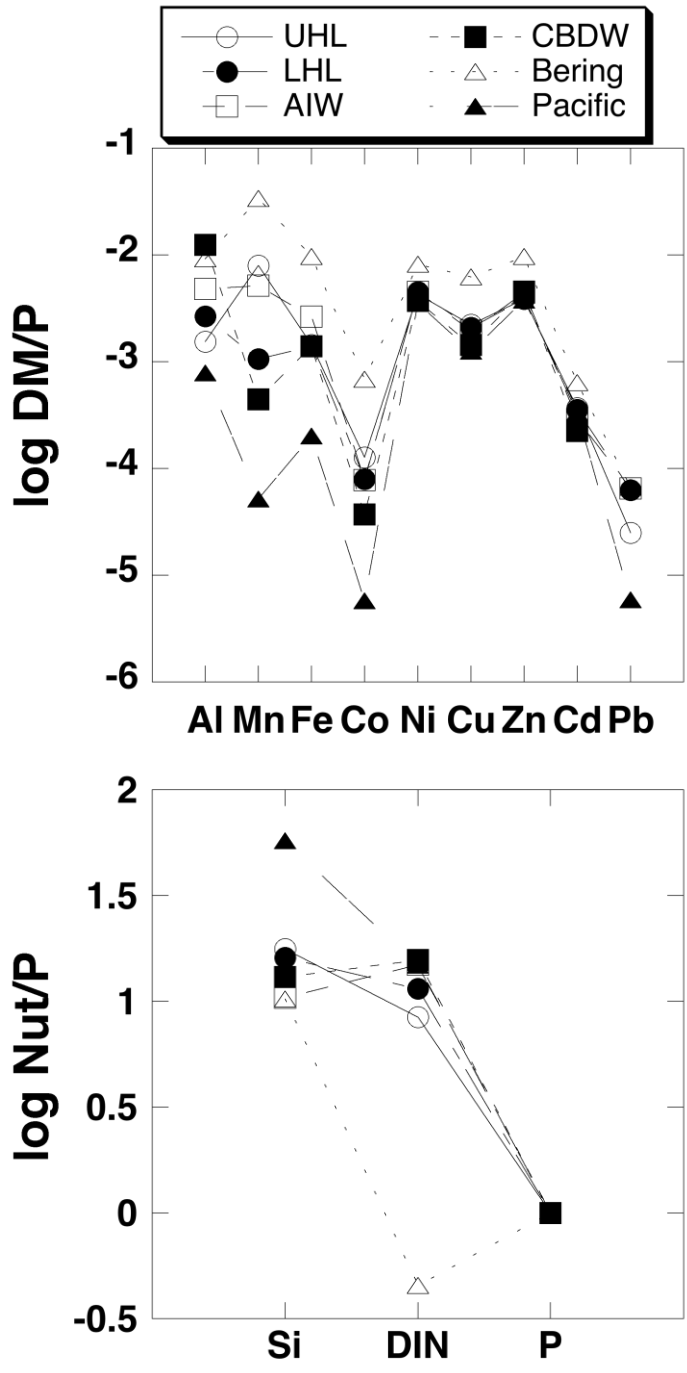


Figure 12

Table 1 Concentrations of bioactive trace metals in each region

		Bering Strait ^a	Barrow Canyon ^b	Mackenzie Trough ^c	Beaufort Sea Slope ^d	Canada Basin ^e
		min - max	min - max	min - max	min - max	min - max
Al [nmol/kg]	TD	182 - 717	12 - 275	17 - 1410	1.7 - 54.6	2.53 - 9.99 ^f
	D	7.2 - 12.0	1.8 - 52.4	1.0 - 12.4	0.9 - 10.4	1.6 - 18.0
Mn [nmol/kg]	TD	54 - 218	26 - 470	8 - 116	3.1 - 34.9	1.6 - 12.2
	D	12.9 - 25.6	15 - 188	0.9 - 18.1	0.2 - 19.7	0.15 - 3.25
Fe [nmol/kg]	TD	998 - 9311	52 - 1212	60 - 4355	10 - 193	9.1 - 50.4
	D	4.9 - 10.4	1.4 - 33.1	1.0 - 24.9	0.57 - 4.56	0.19 - 3.14
Co [pmol/kg]	TD	725 - 2033	347 - 1390	252 - 2505	66 - 668	41 - 313
	D	376 - 475	75 - 980	122 - 593	33 - 515	25 - 267
Ni [nmol/kg]	TD	5.92 - 9.97	3.91 - 8.83	8.3 - 24.0	3.41 - 8.62	3.24 - 8.02
	D	4.88 - 7.59	4.03 - 8.09	6.2 - 15.1	3.18 - 8.07	3.23 - 8.02
Cu [nmol/kg]	TD	5.25 - 9.22	1.58 - 4.18	4.5 - 15.6	1.31 - 5.25	1.22 - 4.38
	D	3.99 - 7.68	1.76 - 4.60	3.2 - 13.0	1.00 - 4.63	1.19 - 3.91
Zn [nmol/kg]	TD	8.0 - 20.3	2.4 - 11.7	4.7 - 33.1	1.5 - 14.9	1.68 - 9.94
	D	2.5 - 13.7	1.74 - 8.54	3.3 - 12.0	2.4 - 13.7	1.76 - 9.82
Cd [nmol/kg]	TD	0.319 - 0.515	0.19 - 1.00	0.456 - 0.707	0.208 - 0.868	0.225 - 0.675
	D	0.351 - 0.479	0.17 - 1.02	0.420 - 0.657	0.196 - 0.747	0.198 - 0.632
Pb [pmol/kg]	TD	77 - 507	66 - 686	51 - 400	10 - 823	22 - 72 ^g
	D	19 - 55	21 - 345	24 - 111	13 - 478	25 - 74 ^g

^aStations A001 and A002; ^bstations A005 and A009; ^cstations A018 and A020; ^dstations A003, A016, A022, A039, A042 and A054; ^estation A023.

^fThe concentrations of TDAI become significantly lower than those of DAI only at this station. See text.

^gThe vertical profiles of TDPb and DPb at this station are very erratic probably due to contamination or faults in determination.

Table 2 Input of bioactive trace metals by the Pacific inflow water and the Mackenzie River water

	Pacific inflow water		Mackenzie River water	
	conc (mol/kg)	input (mol/y)	conc (mol/kg)	input (mol/y)
Al	3.0E-07	7.6E+09	2.2E-06	6.4E+08
Mn	6.9E-08	1.7E+09	1.7E-07	4.9E+07
Fe	1.3E-06	3.3E+10	6.9E-06	2.0E+09
Co	9.8E-10	2.5E+07	3.7E-09	1.1E+06
Ni	7.0E-09	1.8E+08	3.5E-08	1.0E+07
Cu	6.0E-09	1.5E+08	2.3E-08	6.7E+06
Zn	1.4E-08	3.5E+08	5.0E-08	1.5E+07
Cd	3.7E-10	9.3E+06	4.5E-10	1.3E+05
Pb	1.0E-10	2.5E+06	6.3E-10	1.8E+05

Table 3 Compositions of water masses in the Chukchi and Beaufort Seas

	UHL (n = 29)	LHL (n = 9)	AIW (n = 23)	CBDW (n = 8)
	ave \pm sd	ave \pm sd	ave \pm sd	ave \pm sd
Temperature	-1.51 \pm 0.08	-0.74 \pm 0.17	0.46 \pm 0.19	-0.46 \pm 0.10
Salinity	33.03 \pm 0.34	34.25 \pm 0.14	34.78 \pm 0.06	34.94 \pm 0.02
O ₂ [μ mol/kg]	291 \pm 15	268 \pm 6	291 \pm 7	291 \pm 6
Si(OH) ₄ [μ mol/kg]	30.6 \pm 4.4	20.0 \pm 4.2	9.0 \pm 1.5	12.1 \pm 2.2
NO ₃ [μ mol/kg]	13.8 \pm 1.5	14.0 \pm 1.0	12.8 \pm 0.3	14.5 \pm 0.7
NO ₂ [μ mol/kg]	0.04 \pm 0.06	0.01 \pm 0.02	0.00 \pm 0.01	0.00 \pm 0.00
NH ₄ [μ mol/kg]	0.71 \pm 1.67	0.04 \pm 0.03	0.07 \pm 0.09	0.00 \pm 0.00
DIN [μ mol/kg]	14.6 \pm 2.1	14.0 \pm 1.0	12.8 \pm 0.3	14.5 \pm 0.7
PO ₄ [μ mol/kg]	1.72 \pm 0.17	1.23 \pm 0.17	0.86 \pm 0.05	0.93 \pm 0.05
Chl. a [μ g/kg]	0.10 \pm 0.14	0.01 \pm 0.01	0.04 \pm 0.05	
TDAI [nmol/kg]	33.3 \pm 45.8	18.7 \pm 11.4	41.3 \pm 78.4	
DAI [nmol/kg]	2.6 \pm 1.3	3.3 \pm 2.7	4.1 \pm 1.3	11.7 \pm 4.3
TDMn [nmol/kg]	40.9 \pm 95.1	14.1 \pm 8.2	13.2 \pm 18.3	2.1 \pm 0.6
DMn [nmol/kg]	16.0 \pm 39.6	1.4 \pm 2.0	4.6 \pm 11.0	0.4 \pm 0.3
TDFe [nmol/kg]	131 \pm 220	53 \pm 28	129 \pm 240	14 \pm 2
DFe [nmol/kg]	2.7 \pm 3.4	1.7 \pm 1.1	2.3 \pm 2.5	1.3 \pm 0.9
TDCo [pmol/kg]	360 \pm 238	180 \pm 73	147 \pm 107	49 \pm 11
DCo [pmol/kg]	228 \pm 195	102 \pm 72	68 \pm 35	34 \pm 12
TDNi [nmol/kg]	7.7 \pm 0.7	5.4 \pm 1.3	4.0 \pm 0.5	3.5 \pm 0.2
DNi [nmol/kg]	7.3 \pm 0.6	5.6 \pm 1.3	3.9 \pm 0.4	3.4 \pm 0.2
TDCu [nmol/kg]	4.2 \pm 0.5	2.8 \pm 0.7	1.7 \pm 0.4	1.4 \pm 0.1
DCu [nmol/kg]	3.8 \pm 0.4	2.6 \pm 0.6	1.7 \pm 0.4	1.3 \pm 0.1
TDZn [nmol/kg]	6.7 \pm 2.1	4.2 \pm 1.3	3.4 \pm 1.3	4.0 \pm 3.0
DZn [nmol/kg]	6.6 \pm 2.5	4.7 \pm 0.5	3.8 \pm 1.0	4.1 \pm 2.9
TDCd [nmol/kg]	0.67 \pm 0.10	0.45 \pm 0.10	0.26 \pm 0.06	0.24 \pm 0.01
DCd [nmol/kg]	0.63 \pm 0.10	0.44 \pm 0.12	0.25 \pm 0.05	0.21 \pm 0.01
TDPb [pmol/kg]	68 \pm 50	144 \pm 257	60 \pm 79	
DPb [pmol/kg]	43 \pm 22	83 \pm 148	57 \pm 71	