

1 Post-fire forest regeneration under different restoration treatments in the
2 Greater Hinggan Mountain area of China

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8 **Abstract**

9 Forest fire is one of the dominant disturbance factors in boreal forests. Post-fire forest
10 regeneration is crucial to both ecological research and forest management. Three
11 different restoration treatments, namely natural regeneration, artificial regeneration,
12 and artificial promotion, were adopted in the Greater Hinggan Mountain area of China
13 after a serious forest fire occurred on May 6, 1987. Natural regeneration means
14 recovering naturally without any intervention, artificial regeneration comprises
15 salvage logging followed by complete planting, while artificial promotion refers to
16 regeneration by removing dead trees, weeding, and digging some pits to promote seed
17 germination.

18 The objectives of this study were to evaluate and compare the effects of the three
19 restoration treatments and determine which approach is the most suitable for local
20 forest recovery. A field survey was conducted to collect the attribute data, specifically
21 species composition, structural parameters, and Leaf Area Index (LAI), which were
22 analyzed through the analysis of variance and a post-hoc test. The broad-leaved
23 species occupied the main component of the forest under natural regeneration while
24 the coniferous species dominated those under the other two treatments. Tree height
25 and Diameter at Breast Height (DBH) were significantly highest for the forest under
26 artificial regeneration, but an insignificant difference was found for crown widths
27 among the three treatments. Significantly highest LAI was observed in forest under
28 natural regeneration. The results suggest artificial regeneration to be adopted in

29 post-fire recovery if the goal is timber production, while natural regeneration to be
30 utilized when focusing on canopy vertical density and species richness. The artificial
31 promotion treatment showed no advantage. This study demonstrated the advantages of
32 limited strategies that can be helpful for local post-fire forest management.

33 **Keywords**

34 Forest regeneration; Post-fire restoration; Restoration treatments; Leaf Area Index;
35 The Greater Hinggan Mountain area.

36 **1. Introduction**

37 Forests are integrated, multi-functional, and multi-value terrestrial ecosystems with
38 widely distributed coverage areas and complex composition, as well as high species
39 diversity (David et al., 2008; Chen et al., 2010). Forest disturbance and recovery have
40 been regarded as a primary mechanism for transferring carbon between the land
41 surface and atmosphere, thus further play a crucial role in both regional and global
42 carbon cycles as well as in forest monitoring and management (Healey et al., 2005;
43 Soja et al., 2007; Masek et al., 2008; Cao et al., 2011; Dainou et al., 2011; Huang et
44 al., 2012; Camac et al., 2013). Forest fire is one of the main disturbances involved in
45 carbon transfer, especially in boreal forests (Asselin et al., 2001; Gromtsev, 2002;
46 Forkel et al., 2012). Fires are the primary processes which organize the physical and
47 biological attributes of the boreal biome over most of its range and influence the
48 energy flow and biogeochemical cycles (D íz-Delgado and Pons, 2001; Shorohova et
49 al., 2009; Wotton et al., 2010; Chen et al., 2011). Consequently it is necessary to
50 investigate the effects of fire on forest ecosystems and monitor the patterns of
51 post-fire forest regeneration as it pertains to both post-fire ecological research and
52 forest management (Bonan, 1989; Schulze et al., 2005; Stueve et al., 2009; Marzano
53 et al., 2013; Otoda et al., 2013).

54 Although there are already a large number of studies about forest fires and post-fire
55 forest recovery (Bergeron, 2000; Schulze et al., 2005; Shorohova et al., 2009), most of
56 them focus on species composition and forest succession (Uemura et al., 1997;
57 Schulze et al., 2005; Gauthier et al., 2010; Otoda et al., 2013), as well as on the

58 structural parameters and plant cover/tree density (Pausas et al., 1999; Johnstone et al.,
59 2004; Moreira et al., 2009, 2013; Ascoli et al., 2013a; Senici et al., 2013). However
60 few studies have considered the forest ecosystem as a whole one, including all life
61 forms (such as arbor species, shrubs, and grasses) (David et al., 2008), especially
62 using quantitative indicators (Healey et al., 2005). Here, we introduce a
63 three-dimensional scaling parameter, Leaf Area Index (LAI), which has been widely
64 developed and validated in the interdisciplinary field of forest remote sensing
65 combining forestry with remote sensing (McMichael et al., 2004; Boer et al., 2008;
66 Chen et al., 2010).

67 The leaf forms the main surface for matter and energy exchange between the plant
68 canopy and atmosphere, and LAI, defined as half the total leaf area per unit horizontal
69 ground surface area, is proposed as a key variable in the research about terrestrial
70 ecosystems and their development (Chen and Black, 1992; Arias et al., 2007; Schleppi
71 et al., 2011). LAI can be used to characterize the canopy-atmosphere interface of an
72 ecosystem, and is related to precipitation, atmospheric nutrient deposition and
73 interception, canopy microclimate, radiation extinction, as well as water, carbon, and
74 energy exchanges with the atmosphere (Chen et al., 2010). It can be retrieved from
75 remote sensing data, and measured using advanced canopy analyzer devices (Arias et
76 al., 2007).

77 The Greater Hinggan Mountains area, located in the northeast of China, is one of
78 the most important forestry bases with dense virgin forests. This region is rich in
79 forest resources, but also suffers a high incidence of forest fires that are highly
80 determined by fire suppression (Chang et al., 2007; Chen et al., 2011). Among all the
81 fires occurred in this area, that broke out on May 6th, 1987 (abbreviated as “5.6 Fire”
82 hereafter) has become the most noteworthy one, as it was the most serious forest fire
83 since the founding of the People’s Republic of China. It resulted in a burned area of
84 1.7×10^6 ha and a burned forest area of 1.01×10^6 ha (Zhao et al., 1994). After this
85 fire, the local forest bureaus have taken a series of measures to recover the forests;
86 many favorable conditions have been created for speeding up recovery and for
87 ecological forest construction. During the restoration process, three different

88 restoration treatments were adopted for the regeneration of the forests in the burned
89 area. The dynamics of forest regeneration were different depending on the treatment
90 determined for regeneration.

91 In this study, taking the “5.6 Fire” as an example, we investigated the post-fire
92 forest recovery. A field survey was designed and a detailed statistical analysis on the
93 collected data of structural parameters and LAI was performed. The objectives were
94 to evaluate and compare the effects of different restoration treatments, based on which
95 to determine which approach is most suitable for successional regeneration of local
96 forest ecosystems. On the basis of the results, effective suggestions about post-fire
97 forest recovery could be provided for reference in local forest utilization and
98 management.

99 **2. Methods**

100 **2.1 Study area**

101 The study area of the Greater Hinggan Mountain area, located in the northern part
102 of Heilongjiang Province and Inner Mongolia Autonomous Region, is the watershed
103 of the Mongolian Plateau and the flat Songliao Plain, with geographic coordinates
104 ranging from 50°10' to 53°33'N in latitude and from 121°12' to 127°00'E in longitude
105 (Fig. 1). This region has a total length of over 1200 km, a width of 200–300 km, and
106 an average altitude of 1200–1300 m.

107 This area is an important and crucial climatic zone. It has a typical cold temperate
108 continental monsoon climate with warm summers and cold winters. The annual
109 average temperature in this region is -2.8 °C, with the lowest temperature being
110 -52.3 °C, and an average annual precipitation of 746 mm. It is China's northernmost
111 and largest modern state-owned forest area, with a total ground area of $8.46 \times 10^4 \text{ km}^2$
112 and a forest-covered area of $6.46 \times 10^4 \text{ km}^2$. The forest coverage rate amounts to
113 76.4% and the total stand volume is approximately $5.01 \times 10^8 \text{ m}^3$ which accounts for
114 7.8% of the national total stand volume. This mountain area has more than 400
115 species of wild animals and over 1000 varieties of wild plants, and hence is a great
116 reserve of biological resources. It is a mixed forest area with the dominant coniferous

117 species of Mongolian pine (*Pinus sylvestris* L.) and Larch (*Larix gmelini* R.), and the
118 broad-leaved species of Birch (*Betula platyphylla* S.) and Aspen (*Populus davidiana*
119 *D.*).

120 The annual burned forest area of this region ranks first in China, making it the most
121 serious forest fire hazard area. These fires cause great impacts on local forest
122 ecosystems. According to the records of fire in a period of 20 years from 1987 to 2006
123 (Tian et al., 2011), 1059 fires occurred, with a burned area of 2.81×10^6 ha, including
124 1.36×10^6 ha of forest area. The Greater Hinggan Mountain area has been regarded as
125 a key focus region for forest fire prevention and post-fire forest recovery since the
126 most serious forest fire in the history of P. R. China occurred in this region (Sun et al.,
127 2011).

128 **2.2 Sampling design and field data collection**

129 After the “5.6 Fire”, three totally different restoration treatments, namely artificial
130 regeneration, natural regeneration, and artificial promotion, were conducted for forest
131 regeneration. As the name suggests, artificial regeneration indicates an active role of
132 humans in the recovery process through removing all dead or damaged trees
133 belonging to the pre-fire stand from the burned area, followed by complete replanting.
134 In these sites, the coniferous species of *P. sylvestris* and *L. gmelini* were selected and a
135 regular plant spacing (1.5×1.5 m or 1.5×2 m) was adopted according to field
136 conditions. Natural regeneration means allowing the forests to regenerate completely
137 naturally without any human intervention. Thus no any salvage logging or harvesting
138 was conducted, allowing natural restoration to take place. The third approach,
139 artificial promotion, indicates essentially natural regeneration with a select number of
140 artificial aids, which include removing all dead trees and snags, clearing the burned
141 area, weeding, and digging some pits to promote seed germination and growth
142 naturally. In a year with adequate seeds (as the case of forest regeneration in this
143 study), only pits with regular spacing were dug to promote seed rooting, while in a
144 year lacking in seeds, on the basis of the digging, some supplementary measures of
145 artificial seeding was also taken. In any case, no plantation by means of

146 transplantation was performed.

147 In order to investigate the forest restoration dynamics and compare the effects of
148 different restoration strategies, we designed and performed a field forest survey within
149 the burned area of the “5.6 Fire” during July 12–18, 2012, which was the 25th year
150 after the fire. In this survey, we examined three forestry bureaus, Xilinji, Tuqiang, and
151 Amuer, which covered over 85% of the burned forest area of the “5.6 Fire”. We
152 selected and surveyed three plots of 10 m × 10 m in each forest bureau under each
153 type of restoration treatment. With respect to the plot size, in all plots under artificial
154 regeneration and artificial promotion treatments, the trees are evenly distributed
155 within the plots and their surrounding areas, making the size of the plot effectively
156 irrelevant. However, it was not such a case for the plots under natural regeneration.
157 Thus, in future surveys, the size of sampling plots should be increased and the number
158 of observations in each combination of treatment and region is also in need of
159 augmentation.

160 The coordinates of the four corners and the centre of each plot were measured using
161 a differential global positioning system (DGPS). Within each regeneration plot, the
162 species of each individual tree was recorded and the corresponding structural
163 parameters including tree height, Diameter at Breast Height (DBH), and crown width
164 (in the direction of both North-South and West-East) were measured using such
165 devices as an altimeter rod, a tape measure, and a NIKON Forestry Pro 550. The
166 young trees with DBH < 2 cm were recognized as seedlings and being counted;
167 however, their structural parameters (tree height, DBH, and crown widths) were not
168 measured and incorporated in the subsequent statistics. The dominant shrub species
169 were also identified, but not measured. The LAI of each plot was measured using the
170 LAI-2200 canopy analyzer with a height of just above the shrub under stable weather
171 conditions (Chen and Cao, 2012). For all plots in the three forest bureaus, forest
172 regeneration began during the 1987–1989 period, regardless of the restoration
173 treatments, and never suffered any disturbance thereafter. The spatial distribution of
174 these plots was shown in Fig. 2.

175 Fire severity is a crucial element to understanding and interpreting post-fire forest

176 dynamics, as it is one of the main factors affecting post-fire forest regeneration
177 (Schimmel and Granstrom, 1996; Keeley, 2009; Hollingsworth et al., 2013; Marzano
178 et al., 2013). For example, Hollingsworth et al. (2013) proposed that patterns of forest
179 community composition were primarily related to gradients in fire severity, which
180 could thus determine early patterns of community assembly because it has a greater
181 influence than do environmental constraints. In this study, since all the dead and
182 damaged trees had been removed from the burned sites under the artificial
183 regeneration, it was difficult to measure the fire severity in all sites using field survey
184 data. As an alternative, we proposed estimating the disturbance index (DI) (Healey et
185 al., 2005), calculated from remote sensing data, as a proxy of fire severity. The DI
186 image for the entire burned area of the “5.6 Fire” was acquired (Chen et al., 2013) and
187 the values of all surveyed plots were extracted and analyzed. The results indicated that
188 there was no significant difference in fire severity among the three forestry bureaus,
189 and among the areas restored using the three restoration treatments. Therefore, in this
190 study, we did not consider this factor for subsequent analyses and comparisons.

191 **2.3 Statistical analysis**

192 The species composition, structural parameters (tree height, DBH, and crown
193 widths), and LAI of the forests in each forestry bureau under each type of restoration
194 treatment were statistically analyzed. Firstly, the basic statistics (mean and standard
195 deviation) of these attributes were calculated and compared. Then the quantitative
196 Analysis Of Variance (ANOVA) was performed to further demonstrate the
197 comparison. In the ANOVA setting, it was hypothesized that no significant difference
198 existed among different regions and restoration treatments. As the interaction effect of
199 “treatment” by “region” was tested and proven insignificant in all these forest
200 attributes, it was incorporated in the “error effect” to emphasize the “main effect”
201 (“treatment effect” and “region effect”) in the final model. Here the factor of
202 restoration treatment was taken as a fixed factor and the region effect was tested as a
203 random factor. Before ANOVA, a test of normal distribution was performed by using
204 Kolmogorov-Smirnov and Shapiro-Wilk statistics, and the equivalence of variances

205 was tested through Levene's statistic. When the assumptions for the parametric
206 statistical tests of normality and homoscedasticity on original data were not met,
207 logarithmic transformation was carried out to meet ANOVA requirements. The p value
208 (significant probability) was used to determine whether the difference was significant
209 or not. When a significant effect was found (the null hypothesis was rejected),
210 multiple comparisons (post-hoc test) among the three restoration treatments for all
211 attributes were performed using the algorithm of Tukey's Honestly Significant
212 Difference (HSD) and the results were further analyzed to explore the relative
213 relationship between any two of the restoration treatments.

214 **3. Results**

215 **3.1 Species composition**

216 There were only four arbor species in the surveyed plots, which were the coniferous
217 species of *P. sylvestris* and *L. gmelini*, as well as the broad-leaved species of *B.*
218 *platyphylla* and *P. davidiana*. The species composition of regenerated forests in the
219 burned area under the three restoration treatments was quite different due to the
220 different effects of the three restoration treatments (Table 1).

221 It was found there were only coniferous species of *P. sylvestris* and *L. gmelini* in the
222 forest area under artificial regeneration. This result was determined by the species
223 selection in the planting process. The species selection was made according to the
224 local climatic and topographic conditions and considering the recommendations of
225 local forestry engineers and technicians. Additionally, *P. sylvestris* and *L. gmelini*
226 were relatively more suitable for human production, including the paper industry and
227 construction activities, when compared with the broad-leaved species. In the forest
228 area under natural regeneration, there were both the coniferous species (Mongolian
229 pine and Larch) and broad-leaved species (Birch and Aspen), and the latter accounted
230 for the main part. The dominant species in the completely naturally regenerated forest
231 area was *P. davidiana* in Xilinji (78.02%), as well as *B. platyphylla* in Tuqiang
232 (86.24%) and Amuer (84.61%). This suggested that the broad-leaved species had
233 stronger resilience than coniferous species in the naturally regenerated forest area.

234 Besides, from field records, we found more coniferous seedlings and shrub species in
235 forest plots under natural regeneration. For the comparison studies, their structural
236 parameters were not measured and included in the statistical analysis. In the forest
237 area under artificial promotion, the coniferous species became dominant, which was
238 the same as that in the artificially regenerated area.

239 **3.2 Tree Height**

240 The statistics of tree height in the three forestry bureaus, and all regions under the
241 three restoration treatments, was performed and compared (Fig. 3). The results
242 indicated that the average tree heights of forests under artificial regeneration was
243 higher than those under natural regeneration and artificial promotion, with that of
244 artificial promotion being the lowest. The ANOVA result (Table 2a) suggested that the
245 difference in tree height was significant among the three restoration treatments ($p =$
246 $0.031 < 0.05$), but not among different regions ($p = 0.705 > 0.05$). Based on multiple
247 comparisons by Tukey's HSD (letters of a and b on Fig. 3), we were able to conclude
248 that there was no significant difference in tree height between the restoration
249 treatments of artificial promotion and natural regeneration, while they were both
250 significantly lower than that of forest under artificial regeneration.

251 **3.3 Diameter at Breast Height**

252 The statistics (Fig. 4) indicated that the recovered forest under artificial
253 regeneration had a higher average DBH than those under the other two restoration
254 treatments. As shown in Table 2(b), the difference in DBH among the three restoration
255 treatments was extremely significant ($p = 0.004 < 0.01$). However, there was no
256 significant difference in DBH among different regions ($p = 0.432 > 0.05$). The
257 multiple comparison (Fig. 4) gave the result that there was no significant difference in
258 DBH of forests under the treatments of artificial promotion and natural regeneration,
259 however, they were both significantly lower than that under artificial regeneration. It
260 was similar to the analysis result of tree height.

261 **3.4 Crown Width**

262 **3.4.1 Crown Width in North-South direction**

263 The statistics of crown width in the N-S direction (Fig. 5) suggested that the
264 parameter differed in the three forestry bureaus. By qualitative comparison, crown
265 width was larger for trees under artificial regeneration treatment than under the other
266 two treatments. But ANOVA results (Table 2c) indicated that the difference among the
267 three restoration treatments was marginally insignificant ($p = 0.075 > 0.05$) and that
268 among the three regions was also insignificant ($p = 0.160 > 0.05$). Consequently, here
269 further multiple comparison was not necessary. It was quantitatively different from the
270 results of tree height and DBH. Since these sites were located in high latitudes, the
271 subtle difference of solar radiation in the N-S direction may produce an impact on the
272 dynamics of crown width.

273 3.4.2 Crown Width in West-East direction

274 The statistics of crown width in the W-E direction were also analyzed (Fig. 6), from
275 which we came to similar conclusion to that in N-S (Fig. 5). However, ANOVA results
276 (Table 2d) indicated that the crown width in W-E had a marginally significant
277 difference among the three treatments ($p = 0.043 < 0.05$), but not for the three regions
278 ($p = 0.306 > 0.05$). By multiple comparison (Fig. 6), we could conclude that there was
279 a significant difference in crown width in W-E between the restoration treatments of
280 artificial regeneration and artificial promotion. The differences between the two and
281 natural regeneration were both insignificant.

282 3.5 Leaf Area Index

283 The statistical result (Fig. 7) suggested that the LAI of forest under natural
284 regeneration was the highest. ANOVA results (Table 2e) indicated that there was no
285 significant difference in LAI among different regions ($p = 0.544 > 0.05$), however, an
286 extremely significant difference in LAI was observed among the three restoration
287 treatments ($p = 0.007 < 0.01$). Multiple comparisons (Fig. 7) further indicated that a
288 significantly higher LAI was achieved in forest under natural regeneration than under
289 the other two restoration treatments. The average LAI of forests under artificial
290 regeneration was a little higher than that under artificial promotion. It was completely
291 different from the results of structural parameters (tree height, DBH, and crown

292 widths). The structural parameters characterized the states of individual trees, while
293 LAI was a characterization of the overall condition of forest within specific range of
294 regions.

295 **4. Discussion**

296 Forest fire is a common disturbance regime, and post-fire forest recovery following
297 these disturbances plays a crucial role in a variety of fields, including climate change,
298 forest utilization and management which have been widely studied (Pausas et al.,
299 1999; David et al., 2008; Huang et al., 2012; Hollingsworth et al., 2013; Senici et al.,
300 2013). Extensive studies have been conducted to investigate and compare the effects
301 of different strategies on post-fire forest recovery (Moreira et al., 2009, 2013; Beghin
302 et al., 2010; Ascoli et al., 2013b). For example, Moreira et al. (2009) compared two
303 restoration treatments, direct planting and natural regeneration through resprouting, by
304 focusing on the survival and size (height and basal diameter) of *Fraxinus angustifolia*
305 and *Quercus faginea* in Central Portugal. The results suggested that, using natural
306 resprouting may be a cheaper and more effective technique than direct planting to
307 restore burned forests. Beghin et al. (2010) evaluated the impacts of five post-fire
308 management options (no intervention; salvage logging; broadleaved plantation; *L.*
309 *decidua* plantation; and *P. sylvestris* or *Pseudotsuga menziesii* plantation) on natural
310 regeneration structure and composition, and found that density, size, and structural
311 diversity of natural regeneration were higher in the area with no intervention. Aiming
312 at our study area of the Greater Hinggan Mountains, there were three different
313 post-fire forest restoration strategies, artificial regeneration, natural regeneration, and
314 artificial promotion, which had never been comparatively studied in previous research.
315 Here, we proposed to evaluate and compare the effects of the three restoration
316 treatments on forest ecosystem using several forest attributes of structural parameters
317 and LAI simultaneously.

318 On the basis of the field collected data, we concluded that there was significant
319 difference in the dominant species of the forests under different restoration treatments.
320 In the burned area under the treatment of artificial regeneration, the regenerated forest

321 was dominated by the species selected for the planting. However, in the burned forest
322 area, which was allowed to recover completely by natural regeneration, especially in
323 boreal forests, broad-leaved species dominated in the initial stage after the disturbance
324 event. This conclusion has been demonstrated by many previous studies (Uemura et
325 al., 1997; Johnstone et al., 2004; Schulze et al., 2005).

326 Further analysis of the structural parameters (tree height, DBH, and crown widths)
327 suggested that forest under the restoration treatment of artificial regeneration
328 recovered significantly faster than those under the other two treatments. This probably
329 resulted from the different species composition in forests under the three restoration
330 treatments, which in turn, would affect the ecological functions and values of the
331 ecosystems (Johnstone et al., 2004; Gauthier et al., 2010). However, the fact that the
332 forest under natural regeneration recovered marginally faster than that under artificial
333 promotion should be given attention as it indicated that the effect of the restoration
334 treatment of artificial promotion is insignificant and hence its application as a
335 restoration approach need to be reconsidered.

336 As LAI reflects the “layers” of leaves within a certain area in various ecosystems
337 and can characterize the canopy-atmosphere interface effectively, it can be used to
338 indicate the vertical density of vegetation canopy as well as species richness in forest
339 ecosystems (Arias et al., 2007; Schleppei et al., 2011). Besides, LAI is not influenced
340 by the species composition which benefits the comparison among different forest
341 ecosystems. By the analysis of measured LAI of forests under the three different
342 restoration treatments, we found that the completely naturally recovered forest had a
343 higher canopy vertical density and relatively abundant forest species. The fact that the
344 forest under artificial regeneration achieved a larger LAI than that under artificial
345 promotion also raises a question on the need for the latter.

346 By the combined analysis of structural parameters and LAI, we concluded that the
347 artificial regeneration treatment could be adopted in the actual forest management of
348 post-fire recovery if the goal is timber production since the planted coniferous species
349 of *P. sylvestris* and *L. gmelini* are more suitable for human utilization, including the
350 paper industry and construction activities, than the broad-leaved species of *B.*

351 *platyphylla* and *P. davidiana*. Additionally, a variety of other forest species can also be
352 selected in consideration of local climate and soil conditions. However, when the aim
353 of forest restoration is to promote species richness in local forest ecosystem, the
354 burned forest area should be allowed to recover completely under natural regeneration
355 process without any anthropogenic interference. The restoration treatment of artificial
356 promotion, which did not make sense and demonstrate its importance or necessity,
357 should be carefully evaluated and reconsidered. It could probably serve as a means of
358 balance in forest management.

359 In previous studies, it has been found that there is a transition in the dominant
360 species of naturally regenerated forests, usually from dominance of broad-leaved
361 pioneers (such as *Betula* and *Populus*) in the early period to that of conifers (including
362 *Picea* and *Pinus*) in the late-successional stage (Gromtsev, 2002; Johnstone et al.,
363 2004; Gauthier et al., 2010; Otoda et al., 2013), particularly 60–180 years after the fire.
364 Therefore, we need to continue the field observations on forest regeneration in the
365 future, as it has been only 26 years since the serious “5.6 Fire” occurred. This
366 observation will be accomplished with the assistance of local forestry bureaus at a
367 predesignated interval of five years. Moreover, we would design to investigate more
368 sampling sites and increase the plot size in the future field studies. Furthermore, the
369 time-series monitoring of forest disturbance and recovery would be focused on in
370 combination with remote sensing data (McMichael et al., 2004; Masek et al., 2008).
371 In this synergy application, scale matching between the size of the field plots and
372 spatial resolution of remote sensing images should be carefully studied (Mitri and
373 Gitas, 2013).

374 **Acknowledgments**

375 The research in this manuscript was supported by the Japanese Government MEXT
376 Scholarship. The authors are grateful to Professor Shilei Lu from the Institute of
377 Remote Sensing and Digital Earth, Chinese Academy of Sciences, for the guidance
378 and suggestions in determining the technical solution of the sampling design. We
379 express our sincere gratitude to the three doctoral students, Zhou Fang, Haibing Xiang,

380 and Mingren Huang, as well as the forestry technicians and workers from local
381 forestry bureaus, for their assistance in the field work, especially in the measurements
382 of forest structural parameters.

383 **References**

384 Arias, D., Calvo-Alvarado, J., Dohrenbusch, A., 2007. Calibration of LAI-2000 to
385 estimate leaf area index (LAI) and assessment of its relationship with stand
386 productivity in six native and introduced tree species in Costa Rica. *For. Ecol.
387 Manage.* 247, 185-193.

388 Ascoli, D., Castagneri, D., Valsecchi, C., Conedera, M., Bovio, G., 2013a. Post-fire
389 restoration of beech stands in the Southern Alps by natural regeneration. *Ecol. Eng.*
390 54, 210-217.

391 Ascoli, D., Lonati, M., Marzano, R., Bovio, G., Cavallero, A., Lombardi, G., 2013b.
392 Prescribed burning and browsing to control tree encroachment in southern
393 European heathlands. *For. Ecol. Manage.* 289, 69-77.

394 Asselin, H., Fortin, M.J., Bergeron, Y., 2001. Spatial distribution of late-successional
395 coniferous species regeneration following disturbance in southwestern Quebec
396 boreal forest. *For. Ecol. Manage.* 140, 29-37.

397 Beghin, R., Lingua, E., Garbarino, M., Lonati, M., Bovio, G., Motta, R., Marzano, R.,
398 2010. *Pinus sylvestris* forest regeneration under different post-fire restoration
399 practices in the northwestern Italian Alps. *Ecol. Eng.* 36, 1365-1372.

400 Bergeron, Y., 2000. Species and stand dynamics in the mixed woods of Quebec's
401 southern boreal forest. *Ecology* 81, 1500-1516.

402 Boer, M.M., Macfarlane, C., Norris, J., Sadler, R.J., Wallace, J., Grierson, P.F., 2008.
403 Mapping burned areas and burn severity patterns in SW Australian eucalypt forest
404 using remotely-sensed changes in leaf area index. *Remote Sens. Environ.* 112,
405 4358-4369.

406 Bonan, G.B., 1989. Environmental factors and ecological processes controlling
407 vegetation patterns in boreal forests. *Landsc. Ecol.* 3, 111-130.

408 Camac, J.S., Williams, R.J., Wahren, C.H., Morris, W.K., Morgan, J.W., 2013.

409 Post-fire regeneration in alpine heathland: Does fire severity matter? *Austral Ecol.*
410 38, 199-207.

411 Cao, C.X., Chen, W., Li, G.H., Jia, H.C., Ji, W., Xu, M., Gao, M.X., Ni, X.L., Zhao, J.,
412 Zheng, S., Tian R., Liu C., Li, S., 2011. The retrieval of shrub fractional cover
413 based on a geometric-optical model in combination with linear spectral mixture
414 analysis. *Can. J. Remote Sens.* 37, 348-358.

415 Chang, Y., He, H.S., Bishop, I., Hu, Y.M., Bu, R.C., Xu, C.G., Li, X.Z., 2007.
416 Long-term forest landscape responses to fire exclusion in the Great Xing'an
417 Mountains, China. *Int. J. Wildland Fire* 16, 34-44.

418 Chen, H.W., Hu, Y.M., Chang, Y., Bu, R.C., Li, Y.H., Liu, M., 2011. Simulating
419 impact of larch caterpillar (*Dendrolimus superans*) on fire regime and forest
420 landscape in Da Hinggan Mountains, Northeast China. *Chin. Geogr. Sci.* 21,
421 575-586.

422 Chen, J.M., Black, T.A., 1992. Defining leaf-area index for non-flat leaves. *Plant Cell*
423 *Environ.* 15, 421-429.

424 Chen, W., Cao, C.X., 2012. Topographic correction-based retrieval of leaf area index
425 in mountain areas. *J. Mt. Sci.* 9, 166-174.

426 Chen, W., Cao, C.X., He, Q.S., Guo, H.D., Zhang, H., Li, R.Q., Zheng, S., Xu, M.,
427 Gao, M.X., Zhao, J., Li, S., Ni, X.L., Jia, H.C., Ji, W., Tian, R., Liu, C., Zhao, Y.X.,
428 Li, J.L., 2010. Quantitative estimation of the shrub canopy LAI from
429 atmosphere-corrected HJ-1 CCD data in Mu Us Sandland. *Sci. China Ser. D Earth*
430 *Sci.* 53, 26-33.

431 Chen, W., Sakai, T., Moriya, K., Koyama, L., Cao, C.X., 2013. Extraction of burned
432 forest area in the Greater Hinggan Mountain of China based on Landsat TM data.
433 In proceedings of 2013 IEEE International Geoscience and Remote Sensing
434 Symposium (IGARSS). IEEE Geoscience and Remote Sensing Society, NY, USA.

435 Dainou, K., Bauduin, A., Bourland, N., Gillet, J.F., Feteke, F., Doucet, J.L., 2011.
436 Soil seed bank characteristics in Cameroonian rainforests and implications for
437 post-logging forest recovery. *Ecol. Eng.* 37, 1499-1506.

438 David, A.P., Ram, O., Stephen, C.H., 2008. *Forest Ecosystems*, second ed. The Johns

439 Hopkins University Press, Baltimore, Maryland, USA.

440 D áz-Delgado, R., Pons, X., 2001. Spatial patterns of forest fires in Catalonia (NE of
441 Spain) along the period 1975-1995 analysis of vegetation recovery after fire. *For.*
442 *Ecol. Manage.* 147, 67-74.

443 Forkel, M., Thonicke, K., Beer, C., Cramer, W., Bartalev, S., Schullius, C., 2012.
444 Extreme fire events are related to previous-year surface moisture conditions in
445 permafrost-underlain larch forests of Siberia. *Environ. Res. Lett.* 7, 044021.

446 Gauthier, S., Boucher, D., Morissette, J., De Grandpre, L., 2010. Fifty-seven years of
447 composition change in the eastern boreal forest of Canada. *J. Veg. Sci.* 21, 772-785.

448 Gromtsev, A., 2002. Natural disturbance dynamics in the boreal forests of European
449 Russia: a review. *Silva. Fenn.* 36, 41-55.

450 Healey, S.P., Cohen, W.B., Yang, Z.Q., Krankina, O.N., 2005. Comparison of Tasseled
451 Cap-based Landsat data structures for use in forest disturbance detection. *Remote*
452 *Sens. Environ.* 97, 301-310.

453 Hollingsworth, T.N., Johnstone, J.F., Bernhardt, E.L., Chapin, F.S., 2013. Fire
454 Severity Filters Regeneration Traits to Shape Community Assembly in Alaska's
455 Boreal Forest. *PLoS One* 8, e56033.

456 Huang, L., Shao, Q.Q., Liu, J.Y., 2012. Forest restoration to achieve both ecological
457 and economic progress, Poyang Lake basin, China. *Ecol. Eng.* 44, 53-60.

458 Johnstone, J.F., Chapin, F.S., Foote, J., Kemmett, S., Price, K., Viereck, L., 2004.
459 Decadal observations of tree regeneration following fire in boreal forests. *Can. J.*
460 *For. Res.-Rev. Can. Rech. For.* 34, 267-273.

461 Keeley, J.E., 2009. Fire intensity, fire severity and burn severity: a brief review and
462 suggested usage. *Int. J. Wildland Fire* 18, 116-126.

463 Marzano, R., Garbarino, M., Marcolin, E., Pividori, M., Lingua, E., 2013. Deadwood
464 anisotropic facilitation on seedling establishment after a stand-replacing wildfire in
465 Aosta Valley (NW Italy). *Ecol. Eng.* 51, 117-122.

466 Masek, J.G., Huang, C.Q., Wolfe, R., Cohen, W., Hall, F., Kutler, J., Nelson, P., 2008.
467 North American forest disturbance mapped from a decadal Landsat record. *Remote*
468 *Sens. Environ.* 112, 2914-2926.

469 McMichael, C.E., Hope, A.S., Roberts, D.A., Anaya, M.R., 2004. Post-fire recovery
470 of leaf area index in California chaparral: a remote sensing-chronosequence
471 approach. *Int. J. Remote Sens.* 25, 4743-4760.

472 Mitri, G.H., Gitas, I.Z., 2013. Mapping post-fire forest regeneration and vegetation
473 recovery using a combination of very high spatial resolution and hyperspectral
474 satellite imagery. *Int. J. Appl. Earth Obs. Geoinf.* 20, 60-66.

475 Moreira, F., Catry, F., Lopes, T., Bugalho, M.N., Rego, F., 2009. Comparing survival
476 and size of resprouts and planted trees for post-fire forest restoration in central
477 Portugal. *Ecol. Eng.* 35, 870-873.

478 Moreira, F., Ferreira, A., Abrantes, N., Catry, F., Fernandes, P., Roxo, L., Keizer, J.J.,
479 Silva, J., 2013. Occurrence of native and exotic invasive trees in burned pine and
480 eucalypt plantations: Implications for post-fire forest conversion. *Ecol. Eng.* 58,
481 296-302.

482 Otoda, T., Doi, T., Sakamoto, K., Hirobe, M., Nachin, B., Yoshikawa, K., 2013.
483 Frequent fires may alter the future composition of the boreal forest in northern
484 Mongolia. *J. For. Res.* 18, 246-255.

485 Pausas, J.G., Carbó E., Neus, C.R., Gil, J.M., Vallejo, R., 1999. Post-fire regeneration
486 patterns in the eastern Iberian Peninsula. *Acta Oecol.* 20, 499-508.

487 Schimmel, J., Granstrom, A., 1996. Fire severity and vegetation response in the boreal
488 Swedish forest. *Ecology* 77, 1436-1450.

489 Schleppi, P., Thimonier, A., Walthert, L., 2011. Estimating leaf area index of mature
490 temperate forests using regressions on site and vegetation data. *For. Ecol. Manage.*
491 261, 601-610.

492 Schulze, E.D., Wirth, C., Mollicone, D., Ziegler, W., 2005. Succession after stand
493 replacing disturbances by fire, wind throw, and insects in the dark Taiga of Central
494 Siberia. *Oecologia* 146, 77-88.

495 Senici, D., Lucas, A., Chen, H.Y.H., Bergeron, Y., Larouche, A., Brossier, B.,
496 Blarquez, O., Ali, A.A., 2013. Multi-millennial fire frequency and tree abundance
497 differ between xeric and mesic boreal forests in central Canada. *J. Ecol.* 101,
498 356-367.

499 Shorohova, E., Kuuluvainen, T., Kangur, A., Jogiste, K., 2009. Natural stand
500 structures, disturbance regimes and successional dynamics in the Eurasian boreal
501 forests: a review with special reference to Russian studies. *Ann. For. Sci.* 66, 1-20.

502 Soja, A.J., Tchebakova, N.M., French, N.H.F., Flannigan, M.D., Shugart, H.H., Stocks,
503 B.J., Sukhinin, A.I., Parfenova, E.I., Chapin, F.S., Stackhouse, P.W., 2007.
504 Climate-induced boreal forest change: predictions versus current observations.
505 *Glob. Planet. Change* 56, 274-296.

506 Stueve, K.M., Cerney, D.L., Rochefort, R.M., Kurth, L.L., 2009. Post-fire tree
507 establishment patterns at the alpine treeline ecotone: Mount Rainier National Park,
508 Washington, USA. *J. Veg. Sci.* 20, 107-120.

509 Sun, L., Hu, H.Q., Guo, Q.X., Lv, X.S., 2011. Estimating carbon emissions from
510 forest fires during 1980 to 1999 in Daxing'an Mountain, China. *Afr. J. Biotechnol.*
511 10, 8046-8053.

512 Tian, X.R., McRae, D.J., Jin, J.Z., Shu, L.F., Zhao, F.J., Wang, M.Y., 2011. Wildfires
513 and the Canadian Forest Fire Weather Index system for the Daxing'anling region of
514 China. *Int. J. Wildland Fire* 20, 963-973.

515 Uemura, S., Kanda, F., Isaev, A.P., Tsujii, T., 1997. Forest structure and succession in
516 southeastern Siberia. *Veg. Sci.* 14, 119-127.

517 Wotton, B.M., Nock, C.A., Flannigan, M.D., 2010. Forest fire occurrence and climate
518 change in Canada. *Int. J. Wildland Fire* 19, 253-271.

519 Zhao, K.Y., Zhang, W.F., Zhou, Y.W., 1994. The impact of Da xing'an ling forest fires
520 on environment and its countermeasures. Science Press, Beijing, China.

521

522 **Tables**

523 **Table 1.** The species composition of forests in the forestry bureaus of Xilinji, Tuqiang,
 524 and Amuer under three different restoration treatments of Artificial Regeneration
 525 (AR), Natural Regeneration (NR), and Artificial Promotion (AP) (unit: %)

Restoration treatments	Coniferous species		Broad-leaved species	
	<i>Pinus sylvestris</i>	<i>Larix gmelini</i>	<i>Betula platyphylla</i>	<i>Populus davidiana</i>
(a) Xilinji				
AR	66.67	33.33	0.00	0.00
NR	6.59	10.99	4.40	78.02
AP	76.79	16.07	7.14	0.00
(b) Tuqiang				
AR	91.89	8.11	0.00	0.00
NR	3.67	9.17	86.24	0.92
AP	59.18	6.13	34.69	0.00
(c) Amuer				
AR	38.05	61.95	0.00	0.00
NR	3.13	10.94	84.61	1.32
AP	16.67	77.78	5.55	0.00

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529 **Table 2.** Tests of between-subjects effects (ANOVA) in the structural parameters
 530 (Tree height, DBH, and Crown width) and LAI. The “Treatment” and “Region”
 531 denote three forestry bureaus and three restoration treatments

Source	Type III Sum of Squares	df	Mean Square	F statistics	Significant difference (<i>p</i>)
(a) Tree Height					
Treatment	29.233	2	14.617	9.294	0.031*
Region	1.202	2	0.601	0.382	0.705
Error	6.291	4	1.573		
(b) DBH					
Treatment	85.203	2	42.601	29.488	0.004**
Region	3.018	2	1.509	1.044	0.432
Error	5.779	4	1.445		
(c) Crown Width in N-S					
Treatment	3.528	2	1.764	5.288	0.075
Region	2.004	2	1.002	3.003	0.160
Error	1.334	4	0.334		
(d) Crown Width in W-E					

Treatment	3.351	2	1.676	7.603	0.043*
Region	0.712	2	0.356	1.616	0.306
Error	0.882	4	0.220		
(e) LAI					
Treatment	3.551	2	1.775	22.571	0.007**
Region	0.112	2	0.056	0.711	0.544
Error	0.315	4	0.079		

532 * $0.01 < p \leq 0.05$; ** $p \leq 0.01$.

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534

535 **Figure captions**

536 **Fig. 1.** The location of the study area in the Greater Hinggan Mountain area. The
537 background is the mosaic of two Landsat TM scenes representing the burned area of
538 the “5.6 Fire” in dark color.

539 **Fig. 2.** The spatial distribution of the experimental sites and survey plots within the
540 burned forest area of the “5.6 Fire”.

541 **Fig. 3.** The statistics (mean \pm S.D.) and comparison of tree height in different regions
542 under different restoration treatments. The letters a and b indicate the results of
543 multiple comparison.

544 **Fig. 4.** The statistics (mean \pm S.D.) and comparison of Diameter at Breast Height
545 (DBH) in different regions under different restoration treatments. The letters a and b
546 indicate the results of multiple comparison.

547 **Fig. 5.** The statistics (mean \pm S.D.) and comparison of crown width in N-S direction
548 in different regions under different restoration treatments.

549 **Fig. 6.** The statistics (mean \pm S.D.) and comparison of crown width in W-E direction
550 in different regions under different restoration treatments. The letters a and b indicate
551 the results of multiple comparison.

552 **Fig. 7.** The statistics (mean \pm S.D.) and comparison of Leaf Area Index (LAI) in
553 different regions under different restoration treatments. The letters a and b indicate the
554 results of multiple comparison.

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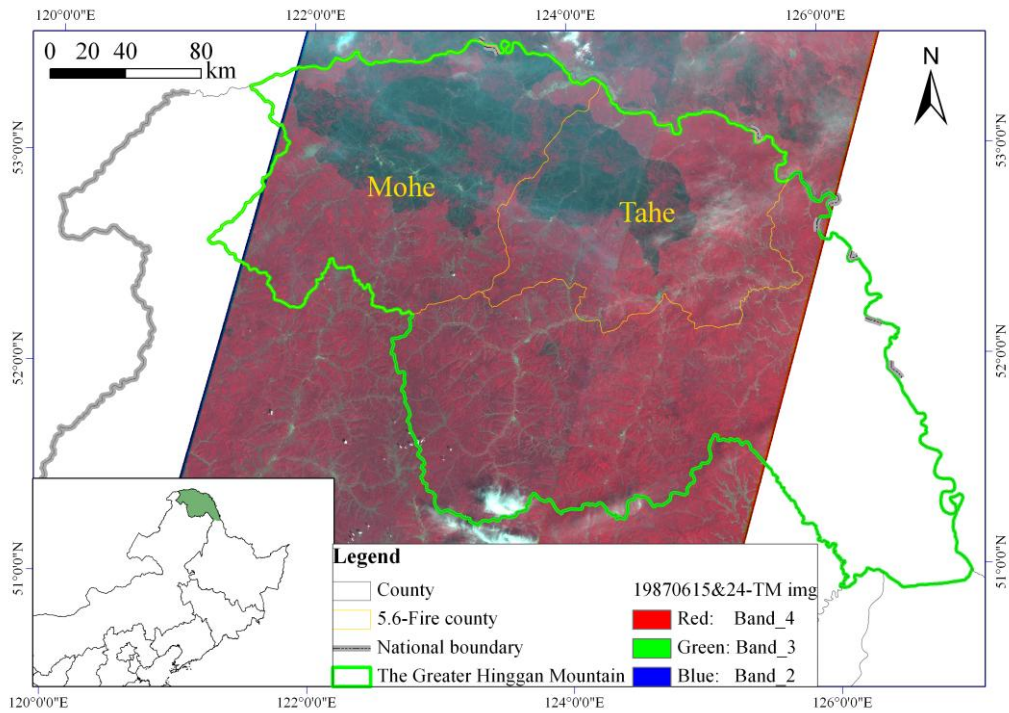


Fig. 1. The location of the study area in the Greater Hinggan Mountain area. The background is the mosaic of two Landsat TM scenes representing the burned area of the “5.6 Fire” in dark color.

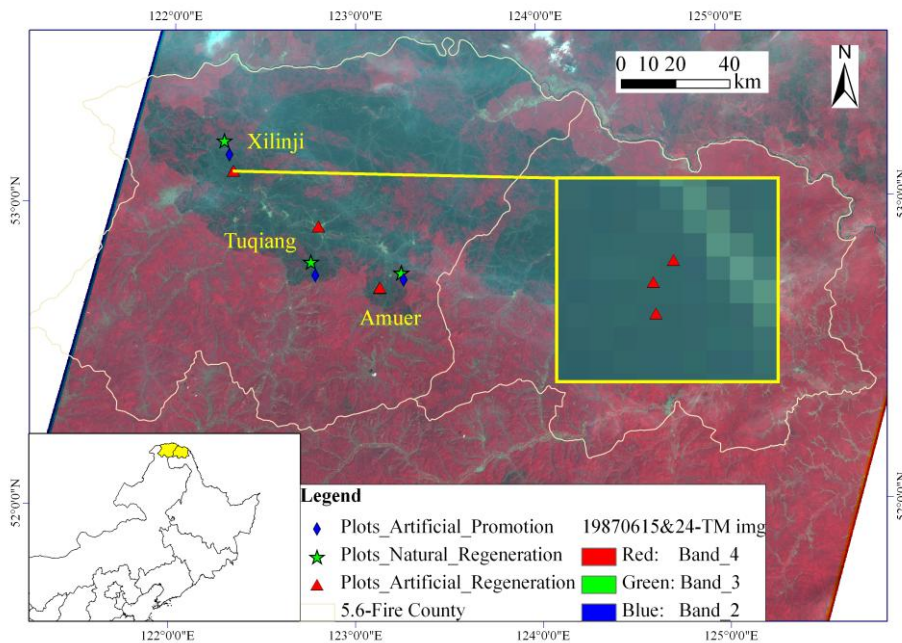


Fig. 2. The spatial distribution of the experimental sites and survey plots within the burned forest area of the “5.6 Fire”.

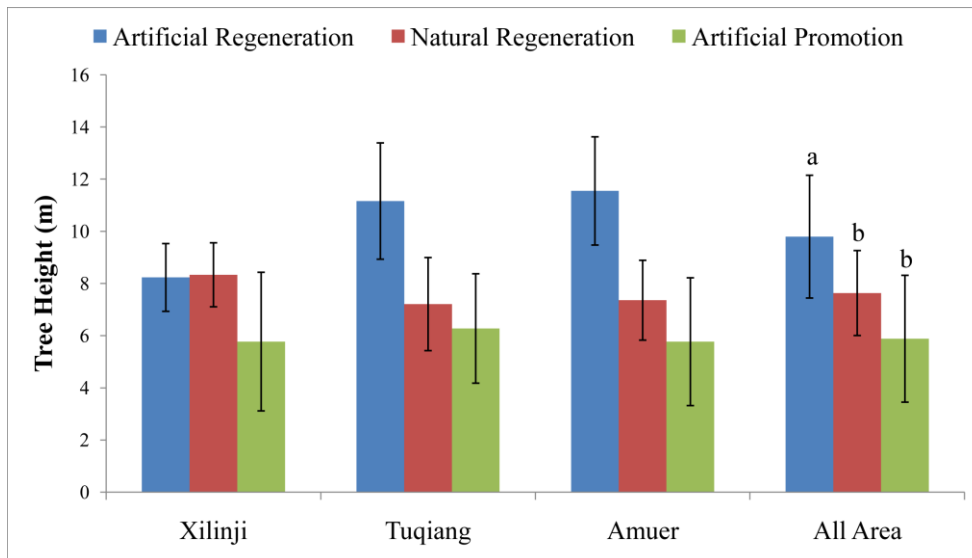


Fig. 3. The statistics (mean \pm S.D.) and comparison of tree height in different regions under different restoration treatments. The letters a and b indicate the results of multiple comparison.

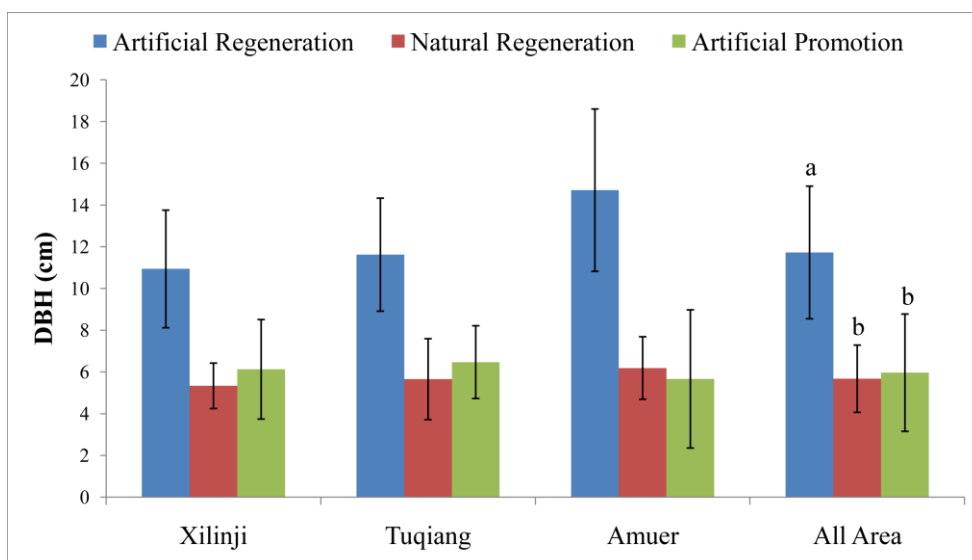


Fig. 4. The statistics (mean \pm S.D.) and comparison of Diameter at Breast Height (DBH) in different regions under different restoration treatments. The letters a and b indicate the results of multiple comparison.

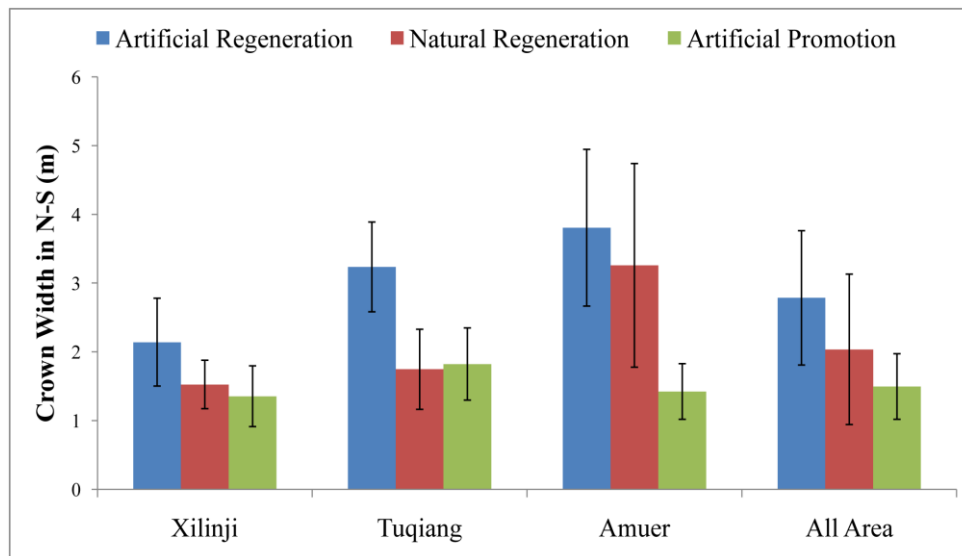


Fig. 5. The statistics (mean \pm S.D.) and comparison of crown width in N-S direction in different regions under different restoration treatments.

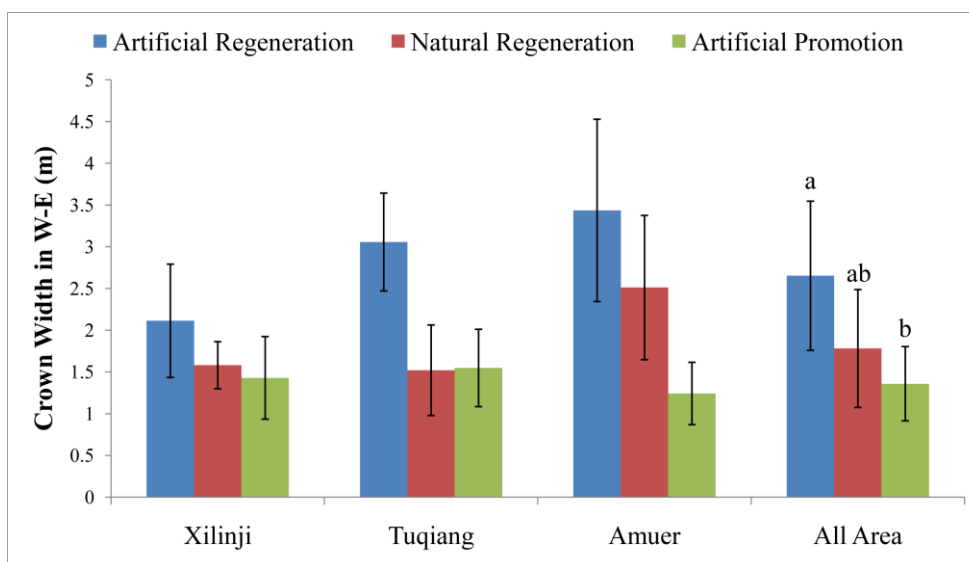


Fig. 6. The statistics (mean \pm S.D.) and comparison of crown width in W-E direction in different regions under different restoration treatments. The letters a and b indicate the results of multiple comparison.

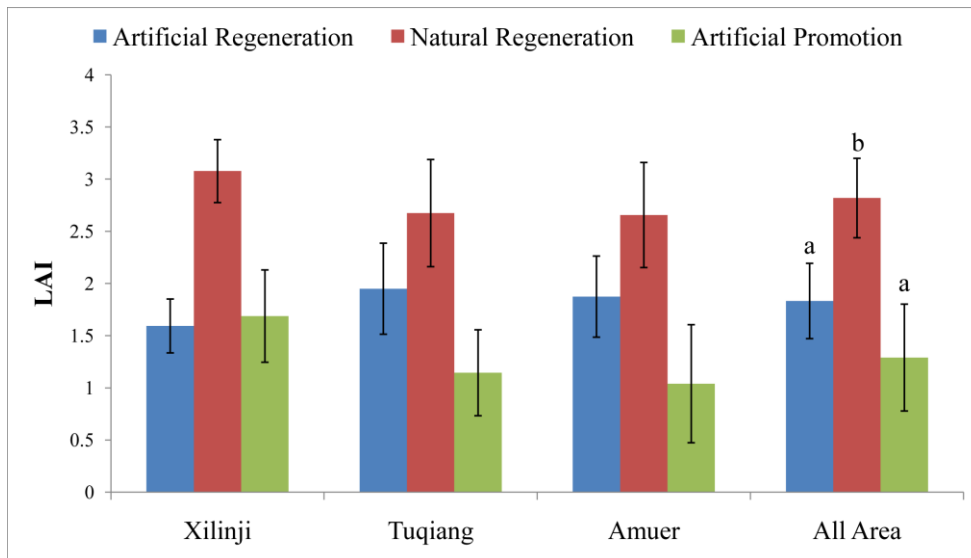


Fig. 7. The statistics (mean \pm S.D.) and comparison of Leaf Area Index (LAI) in different regions under different restoration treatments. The letters a and b indicate the results of multiple comparison.