

1 Diversity, resource utilization, and phenology of fruiting bodies of  
2 litter-decomposing macrofungi in subtropical, temperate, and subalpine forests

3

4 Takashi Osono

5

6 T. Osono

7 Center for Ecological Research, Kyoto University, Otsu, Shiga 520-2113 Japan

8 e-mail: tosono@ecology.kyoto-u.ac.jp

9

10 **Abstract**The diversity, vegetative and reproductive characteristics, and phenology  
11 of litter decomposing macrofungi (LDM) were compared between humus forms  
12 and climatic regions. Fruiting bodies of LDM were examined for the forest floor of  
13 subtropical (ST), cool temperate (CT), and subalpine (SA) forests in Japan. Field  
14 surveys during one growing season yielded 35, 32, and 18 species in ST, CT, and  
15 SA, respectively. Species richness was generally higher in mull than in moder  
16 humus and in warmer than in cooler climate. A total of 10 fungal families were  
17 observed, and species in the Mycenaceae dominated in the LDM assemblages at

18 all study sites. A larger number of species fruited on deeper F layers of the forest  
19 floor in SA than in ST, where 74% of species fruited directly on leaf litter. This  
20 observation was consistent with the analysis of radiocarbon content in fruiting  
21 bodies, implying that LDM tended to utilize older carbon accumulated at deeper  
22 layers of the forest floor in cooler climates. Seasonal changes in the fruiting  
23 frequency over a growing season exhibited similar two-peak patterns for all the  
24 study sites, coinciding with the periods of rainfall and increasing and decreasing  
25 air temperatures in early summer and autumn, respectively, but the fruiting  
26 period extended longer in warmer than in cooler climate.

27

28 **Keywords**      Climate · Decomposition · *Mycena* · Radiocarbon ·

29 Seasonal changes

30

31 **Introduction**

32

33 Litter-decomposing macrofungi (LDM) are major components of the diversity of  
34 soil organisms in terrestrial ecosystems and play central roles in the

35 decomposition of structural and soluble components in litter that often limit  
36 carbon and nutrient cycling in soil (Osono 2007; van der Wal et al. 2013). A suite of  
37 LDM with the ability to decompose lignin and other recalcitrant compounds are of  
38 particular importance because the colonization of litter materials by these fungi  
39 often stimulates the turnover of organic matter and nutrients in soil (Steffen et al.  
40 2007; Valášková et al. 2007; Osono et al. 2011a). Fruiting bodies of LDM provide  
41 reliable and useful information about their taxonomy, diversity, and reproduction  
42 and have been surveyed for their diversity (e.g. Schmit et al. 1999; Mueller et al.  
43 2007) and seasonal patterns (Murakami 1989; Yamashita and Hijii 2004) and for  
44 the effects on them of vegetation (Hansen and Tyler 1992; Lange 1993; Såstad  
45 1995), soil conditions (Tyler 1985; Rastin et al. 1990), and elevational gradient  
46 (Gómez-Hernández et al. 2012). Moreover, the observation of vegetative mycelia  
47 at the base of fruiting bodies can often yield insights into the substrate utilization  
48 and decomposing ability of LDM (Osono et al. 2011a). Currently, however, few  
49 studies have investigated the diversity, vegetative and reproductive  
50 characteristics, and phenology of LDM simultaneously and compared these  
51 between humus forms and climatic regions. It is hypothesized that the pattern of

52 diversity, substrate utilization, and phenology of fruiting bodies of LDM change  
53 along gradients of soil conditions and climate.

54           The purpose of the present study was to investigate fruiting bodies of  
55 LDM emerging from the forest floor of subtropical, cool temperate, and subalpine  
56 forests in Japan. Field sampling of fruiting bodies over growing seasons yielded  
57 information about the structure, diversity, and species composition of LDM  
58 assemblages and seasonal patterns of occurrence. Each LDM species was recorded  
59 for the soil layer from which its fruiting body emerged to examine the substrate  
60 its vegetative mycelia utilized. Radiocarbon ( $^{14}\text{C}$ ) contents of fruiting bodies were  
61 measured for major LDM species to estimate the age of carbon (i.e. time since  
62 death of plant litter) utilized by these species. The diameter of the pileus and the  
63 length of the stipe were measured for fruiting bodies found in the three forest soils,  
64 and possible roles of the size variation of fruiting bodies in the seasonal patterns  
65 of fruiting bodies were discussed.

66

67 **Materials and methods**

68

69 Study site

70

71 Samples were collected from three sites in Japan: a subtropical forest (ST), a cool  
72 temperate forest (CT), and a subalpine forest (SA). ST was located in Okinawa,  
73 southern Japan. CT was located in Kyoto, Japan. In CT, two study plots were  
74 established on the upper and lower parts of a northwest-facing slope  
75 (approximately 200 m long). SA was located on Mt. Ontake, Gifu, Japan. Details  
76 of the location, climatic conditions, and vegetation are given in Osono (submitted).  
77 In summary, the three sites differed in mean annual temperature (22°C, 9°C, and  
78 2°C in ST, CT, and SA, respectively), seasonal patterns of change in air  
79 temperature, and the duration of the growing season, but they received similar  
80 amounts of precipitation annually. The study sites experience a rainy season from  
81 May to June in ST and from June to July in CT and SA. Snow covers the forest  
82 floor of CT from December to April and that of SA from mid-November to early  
83 June. Table 1 shows properties of the forest floor of the study sites. The  
84 accumulation of forest floor material, in terms of the depth and the mass, was in  
85 the order: ST, CT (lower) < CT (upper) < SA, whereas the order was generally

86 reversed for the leaf fall mass. Consequently, the turnover time of the forest floor  
87 was lower in SA and CT (lower) (less than two years) than in CT (upper) (10.4  
88 years) and in SA (29.1 years).

89

90 Study plot and field survey

91

92 A study plot of  $50 \times 10$  m ( $500 \text{ m}^2$ ) was laid out in each of ST, CT (upper), CT  
93 (lower), and SA sites and was divided into 125 grids of  $2 \times 2$  m. The study area of  
94  $500 \text{ m}^2$  was found to be large enough to describe species richness of macrofungi in  
95 CT sites, according to Okabe (1986).

96 Fruiting bodies of LDM were collected in the study plots, seven times at  
97 1- to 2-month intervals from March 2007 to January 2008 in ST, nine times at 2-  
98 to 4-week intervals from May to November 2001 in CT, and five times at 1-month  
99 intervals from June to October 2008 in SA. On each sampling occasion, all fruiting  
100 bodies encountered on the surface of the forest floor were recorded, excepting  
101 obviously immature or rotting ones. Records were kept of taxa and of grid number  
102 and soil horizons (L layer, the border between L and F layers, F layer, or A layer)

103 from which the fruiting bodies emerged (see Table S1 in Electronic  
104 Supplementary Material). Fruiting bodies occurring on logs, twigs, or roots that  
105 were fallen or buried were not recorded. Ascomycetes were omitted, but the  
106 Xylariaceae on leaf litter were included because of their ligninolytic activity  
107 (Osono et al. 2011b). In October and November 2011, fruiting bodies were  
108 measured for the diameter of their pileus and length of their stipe at the three  
109 sites.

110 Identification was primarily made macroscopically after Imazeki et al.  
111 (1988), Imazeki and Hongo (1987, 1989), and Hongo (1994). Small fruiting bodies  
112 of *Mycena* and *Marasmius* that were difficult to distinguish and identify at the  
113 species level in the field were classified at the genus or section level, which was  
114 referred to as species in the present study for the sake of simplicity (but see  
115 Discussion). Tissues of some fruiting bodies were further analyzed for the DNA  
116 sequence of amplicons of rDNA ITS region obtained using primers ITS5 and ITS4  
117 (White et al. 1990) and of the 28S rRNA gene D1/D2 region using primers D1  
118 (Peterson 2000) and NL4 (O'Donnell 1993), according to the method of Hirose and  
119 Osono (2006). The sequences determined were compared with the available rDNA

120 sequences in the GenBank database by means of BLAST+ (Camacho et al. 2009)  
121 and assigned taxonomically. The data of molecular analyses will be given in a  
122 future paper.

123           The frequency of occurrence of LDM was calculated as a percentage of  
124 incidences based on the number of grids in which the fruiting body was  
125 encountered relative to the total number of grids (125) at each study site. Relative  
126 frequency of an individual species was calculated as the percentage of its  
127 frequency of occurrence with respect to the grand sum of the frequency of  
128 occurrence of all species at each study site. Data of fruiting bodies of mycorrhizal  
129 fungi were excluded from the following analyses.

130

131 Radiocarbon analysis

132

133 Samples of fruiting bodies were ground in a laboratory mill to make particles that  
134 would pass through a 0.5-mm screen and sent to the Institute of Accelerator  
135 Analysis Ltd, Kanagawa, Japan, for accelerator mass spectrometry  
136 measurements of radiocarbon. The methods are described in Hyodo et al. (2006).



137 Radiocarbon values were reported as  $\Delta^{14}\text{C}$  (‰), which is the part per-thousand  
138 deviation from the activity of nineteenth century wood, and corrected for the  
139 fractionation using stable carbon (C) isotope ratios of the samples.

140           The method to estimate the carbon age of fruiting bodies of LDM followed  
141 Hyodo et al. (2006). The carbon age of fruiting bodies of fungi was defined as the  
142 time elapsed since C in their substrates was fixed from atmospheric  $\text{CO}_2$  by  
143 primary producers.  $\Delta^{14}\text{C}$  values of samples were compared with those of  
144 atmospheric  $\text{CO}_2$  recorded at Schauinsland, Germany, for 1976-97 (Levin and  
145 Kromer 1997). I estimated the  $\Delta^{14}\text{C}$  values of atmospheric  $\text{CO}_2$  after 1997 by  
146 extrapolation of the exponential function:  $\Delta^{14}\text{C}(t) = 417 \times \exp(-t/16.0)$ , where t is  
147 the year after 1974 (Levin and Kromer 1997). This method yielded two estimates  
148 of the year of C fixation for the measured  $\Delta^{14}\text{C}$  values of fruiting bodies, one before  
149 and another after the peak of bomb- $\Delta^{14}\text{C}$  in mid-1960s, and hence two carbon  
150 ages (Hyodo et al. 2006). I adopted the carbon ages estimated from the year of C  
151 fixation after the peak bomb- $\Delta^{14}\text{C}$ , because these estimated carbon ages were  
152 compatible with the turnover rates of the forest floor (1.5 to 29.1 years in the  
153 study sites, Table 1).

154

155 Statistical analysis

156

157 The observed number of LDM species at each study site was denoted as  $S_{obs}$ . It  
158 was possible that  $S_{obs}$  be underestimated when the abundance of fruiting bodies of  
159 LDM encountered (i.e., the sampling effort) was low at any study site, compared  
160 to other sites. To avoid this, I used an individual-based Coleman rarefaction curve  
161 (Colwell and Coddington 1994) to depict the cumulative number of species versus  
162 the observation of fruiting bodies (Fig. S2 in Electronic Supplementary Material).  
163 In the present study, the number of observation of fruiting bodies was variable  
164 among the study sites, ranging from 77 at CT (upper) to 620 observations at ST.  
165 Thus, the study sites were compared for the estimated numbers of LDM species at  
166 77 observations (denoted as  $S_{est}$ ). Calculations were performed with R version  
167 3.0.2 for Mac (R Development Core Team 2009) and its vegan package (Oksanen  
168 2013).

169 Simpson's diversity index (D) and equitability (E) were calculated in the  
170 following equations (Osono et al. 2002):  $D = 1 / \sum P_i^2$ ,  $E = D / S_{obs}$ , where  $P_i$  was a

171 proportion of the frequency of occurrence of  $i$ th species to the sum of frequency of  
172 all species.

173 Generalized linear models (GLMs) were used with a Gaussian  
174 distribution to compare the size of fruiting bodies of LDM between the study sites.  
175 The GLMs were performed with the *glm* function and with the *glht* function of the  
176 R multcomp package for multiple comparisons with Tukey's test.

177

## 178 **Results**

179

### 180 Species richness and taxonomic composition

181

182 A total of 35, 32, and 18 species of LDM were observed ( $S_{\text{obs}}$ ) in ST, CT, and SA,  
183 respectively; and in CT, 25 and 11 species were observed at lower and upper slopes,  
184 respectively (see ESM; summarized in Table 2). The number of singleton species  
185 (i.e. species encountered in only one grid) accounted for 17% to 49% of the total  
186 number of species, in the order: ST > CT (lower) > CT (upper) > SA (Table 2).  
187 Simpson's diversity index was in the order: ST > SA > CT (lower) > CT (upper),

188 and equitability was in the order: CT (upper) > SA > ST > CT (lower) (Table 2).  
189 Rarefaction analysis showed that the estimated number of LDM species ( $S_{est}$ )  
190 were higher in mull [ST and CT (lower)] than in moder humus [CT (upper) and  
191 SA] (Table 2).

192 A total of 10 fungal families were observed: five, seven, five, and five  
193 families in ST, CT (lower), CT (upper), and SB, respectively (Table 2). *Mycena*  
194 species in the Mycenaceae dominated in the LDM assemblages at each study site  
195 in terms of the number of species (27% to 50% of the total number of species; Table  
196 2) and the relative frequency (Fig. 1). The frequencies of occurrence of two major  
197 *Mycena* species reached more than 90% (i.e. the fruiting bodies of these species  
198 occurred in more than 90% of the 125 grids) in ST and between 13.6% to 60.8% in  
199 CT and SB (see ESM). These major *Mycena* species were followed by species in  
200 Marasmiaceae in ST and CT (upper), by species in Agaricaceae in CT (lower), and  
201 by species in Hymenogasteraceae in SA (Table 2; Fig. 1).

202

203 Soil layer from which macrofungi fruited

204

205 Soil layer from which LDM fruited differed among the study sites: more number of  
206 LDM species that fruited from deeper layers of the forest floor at cooler climate.  
207 That is, 74% (26/35) of species in ST fruited on the surface L layer (i.e. emerging  
208 directly on leaf litter), whereas 73% to 92% from the border between L and F  
209 layers in CT, and 78% from the F layer in SA (Table 2) did so. Those that fruited  
210 on L layer were 'component-restricted' sensu Osono (2007) in that individual  
211 mycelia were limited in extent by the physical boundaries of the litter component  
212 they occupied. Conversely, those that fruited on the L-F border and F layer were  
213 'component-non-restricted' in that the entire forest floor, rather than an  
214 individual litter component, provided a habitat for the fungi.

215

216 Radiocarbon content

217

218 The mean  $\Delta^{14}\text{C}$  values of fruiting bodies ranged between 48.2‰ and 139.7‰ (Table  
219 3), indicative of the fungal uptake of bomb- $\Delta^{14}\text{C}$  that was primarily derived from  
220 the litter that these LDM utilized (Hyodo et al. 2006). The carbon age of fruiting  
221 bodies from ST ranged from 2.8 to 9.0 years. These values suggested that these

222 LDM utilized leaves that died at least 1.8 to 8.0 years previously, because tree  
223 leaves in ST were mostly evergreen and had leaf longevity of more than one year.  
224 In contrast, the carbon age of fruiting bodies from CT ranged from 3.6 to 11.4  
225 years, suggestive of the utilization of deciduous leaves that died as long as 11.4  
226 years before. The carbon age of fruiting bodies from SA reached as old as 20.3  
227 years, suggestive of the utilization of evergreen leaves (maximum ages of 6 to 11  
228 years, Mori and Takeda 2004) that had died at least 10 years before. These results,  
229 together with the results of direct observation of fruiting bodies, suggested that  
230 LDM tended to utilize older carbon accumulated at deeper layers of the forest  
231 floor in cooler climates.

232

233 Size and phenology of fruiting bodies

234

235 The mean size of the fruiting bodies, measured as the diameter of pileus and  
236 length of stipe, was significantly different among the study sites (ANOVA,  $p < 0.05$ ),  
237 in the order: CT > SA > ST (Table 4). This was accounted for by the difference in  
238 size of the major *Mycena* species in these study sites (Table 4).

239           The frequency of occurrence of fruiting bodies was generally higher in ST  
240 than in CT or SA (Fig. 2). Seasonal changes in the frequency over a growing  
241 season exhibited similar two-peak patterns for all the study sites: a peak during  
242 the rainy season in early summer and another in autumn (Fig. 2). That is, the  
243 peaks were found in June and in September to January at ST, in June and in  
244 September to November at CT, and in July and in September-October at SA. The  
245 number of LDM species followed similar seasonal patterns as the frequency of  
246 occurrence of fruiting bodies, except that there was a rapid increase in the  
247 number of species in June in ST.

248           The major LDM species differed in the seasonal patterns of their  
249 frequency of occurrence over a growing season (Fig. 3). In ST, fruiting bodies of  
250 some major species, such as *Mycena* sp.ST1 and *Marasmius* sp.ST1, occurred  
251 relatively constantly over the growing season, whereas *Mycena* sp.ST2, *Xylaria*  
252 sp.ST1, and *Crinipellis* sp.ST1 displayed fruiting peaks in June and/or in  
253 September to January. In CT and SA, the frequencies of major species increased  
254 once in autumn (*My. polygramma* in CT and *My. eipterygia*, *G. atkinsoniana*,  
255 and *Mycena* sp.SA1 in SA) or twice (in early summer and in autumn) (*My. filopes*

256 in CT and *My. aurantiidisca* in SA) over the growing season.

257

## 258 **Discussion**

259

260 The numbers of species of LDM observed in CT and SA (Table 2) were within the  
261 range reported previously in temperate and boreal forests (Tyler 1985; Hintikka  
262 1988; Villeneuve et al. 1989; Brunner et al. 1992; Miyamoto et al. 2000;  
263 Outerbridge 2002; Richard et al. 2004; Gates et al. 2011; O'Hanlon and  
264 Harrington 2012), despite the short survey period (one growing season) at each  
265 study site. The dominance of *Mycena* in terms of the number of species and the  
266 frequency of occurrence is consistent with these previous reports. López-Quintero  
267 et al. (2012) also observed the occurrence of *Mycena* and *Marasmius* species in  
268 Amazon tropical rainforests, but comparative studies on the diversity of LDM in  
269 tropical and subtropical forests have been relatively scarce, especially in Asian  
270 tropical regions (Mueller et al. 2007).

271 The number of observed and estimated species and Simpson's diversity  
272 index of LDM were generally higher in mull than in moder humus (Table 2). This



273 was evident in CT, where the richness of LDM was higher at the lower (mull) than  
274 the upper (moder) slope. The two most frequent species were common to the two  
275 sites, so that infrequent species accounted for the low similarity of species  
276 composition (Table 2, Fig. 1). Rastin et al. (1990) also compared LDM between the  
277 lower and upper slope of a spruce forest in Germany and reported that the species  
278 composition was generally similar between those parts, in contrast to the results  
279 of the present study. This discrepancy may be partly due to the slope length [30 m  
280 in Rastin et al. (1990) versus 200 m in the present study]. The causal factors for  
281 the higher LDM richness in mull of CT remain unclear, but the relatively  
282 favorable moisture condition at the lower slope could possibly favor the fruiting  
283 and co-occurrence of more LDM species on the forest floor.

284           The LDM diversity was generally higher in warmer than in cooler  
285 climates, suggesting a climatic gradient of diversity. Similar climatic gradients of  
286 fungi have been found for litter decomposing microfungi (Osono 2011) and foliar  
287 endophytic fungi (Arnold and Lutzoni 2007; Ikeda et al. 2014). At least two  
288 explanations are possible for the putative higher diversity of fruiting bodies of  
289 LDM in warmer locations. First, the warmer condition throughout the year and

290 lack of snow cover period in winter of ST can favor the fruiting (and possibly, the  
291 co-existence) of multiple LDM. This is illustrated in the fruiting phenology of  
292 major LDM species (Fig. 3): fruiting bodies occurred throughout the year or with  
293 multiple peaks in ST, whereas in CT and SA they peaked once or twice over the  
294 growing season. Such differences may be partly due to suitability of the climatic  
295 conditions for establishment, growth, and fruiting of more LDM species in ST.

296         Secondly, differences in the quality of resources can also affect the  
297 diversity of LDM. I found that LDM from a cooler climate produced fruiting bodies  
298 that originated from deeper soil layers than those from a warmer climate (Table 2).  
299 In accordance with this, LDM from a cooler climate appeared to utilize more aged  
300 dead carbon than those from a warmer climate (Table 3). Given that more  
301 decomposed materials in deeper layers contain less readily available organic  
302 carbon sources, such as non-lignified holocellulose and soluble carbohydrates  
303 (Berg 1986; Osono et al. 2003), the utilization of resources at deeper layers in a  
304 cooler climate may be unfavorable for the growth and fruiting of LDM. Osono  
305 (2011) demonstrated that non-ligninolytic microfungi are major components of  
306 fungal assemblages on recently fallen litter in a cooler climate, suggesting that

307 ligninolytic LDM are less competitive for readily available resources in the  
308 surface litter in a cooler climate (Osono 2007). The dominance in SA of conifers,  
309 whose needles are rich in secondary compounds that inhibit the growth of LDM  
310 (Bađci and Diđrak 1996), can also affect the colonization of the L layer by LDM.

311           Two peaks were found for the occurrence of fruiting bodies over a growing  
312 season at all three climates, but the fruiting period extended longer at warmer  
313 than at cooler climates (Fig. 2). The two peaks coincided with the period of rainfall  
314 and with the increasing and decreasing air temperatures in early summer and  
315 autumn, respectively, at the three forest sites. Similar one- or two-peak patterns  
316 of fruiting of LDM have commonly been found in temperate forests (Okabe 1983;  
317 Straatsma et al. 2001; Yamashita and Hijii 2004; Gates et al. 2011).

318           The fruiting phenology of major *Mycena* species may also be associated  
319 with the size of fruiting bodies. For example, *Mycena* species with smaller fruiting  
320 bodies fruited more frequently over a growing season than those with larger ones;  
321 *M. filopes* in CT and *M. aurantiidisca* in SA with smaller fruiting bodies showed  
322 two peaks, whereas *M. polygramma* in CT and *M. eipterygia* in SA with larger  
323 fruiting bodies showed only one peak in autumn (Fig. 3, Table 4), and even

324 smaller *Mycena* sp. ST1 and ST2 in ST fruited throughout the growing season.  
325 Because the production of larger fruiting bodies should require more resources to  
326 be utilized, the size of fruiting bodies can set a limit on the reproduction. This  
327 discussion is obviously speculative, however, as few data have been available  
328 regarding the population structure and reproductive biology of individual LDM  
329 species. More studies are needed to examine the life history strategy of LDM,  
330 especially in tropical and subtropical regions.

331

332 **Acknowledgments**            I thank Dr. D. Hirose and Dr. T. Miyamoto for help with  
333 the identification of macrofungi; Dr. F. Hyodo for useful discussion of radiocarbon  
334 analysis; Dr. A. Takashima and staff at the Yona Experimental Forest, University  
335 of the Ryukyus for help with fieldwork at ST; Dr. Y. Fukasawa and staff at Ashiu  
336 Experimental Forest, Kyoto University for help with fieldwork at CT; and Dr.  
337 Elizabeth Nakajima for critical reading of the manuscript. This study received  
338 partial financial support from the Ministry of Education, Culture, Sports, Science,  
339 and Technology of Japan (MEXT) (No. 19780114), The Sumitomo Foundation,  
340 Nissan Global Foundation, Nippon Life Inst. Foundation, and the Grants for

341 Excellent Graduate Schools, MEXT, Japan (12-01) to Kyoto University.

342

343 **References**

344

345 Arnold AE, Lutzoni F (2007) Diversity and host range of foliar endophytes: are  
346 tropical leaves biodiversity hotspots? *Ecology* 88:541-549

347 Bağci E, Diğrak M (1996) Antimicrobial activity of essential oils of some *Abies*  
348 (Fir) species from Turkey. *Flavour Fragr J* 11:251-256

349 Berg B (1986) Nutrient release from litter and humus in coniferous forest soils - a  
350 mini review. *Scand J For Res* 1:359-369

351 Brunner I, Brunner F, Laursen GA (1992) Characterization and comparison of  
352 macrofungal communities in an *Alnus tenuifolia* and an *Alnus crispa*  
353 forest in Alaska. *Can J Bot* 70:1247-1258

354 Camacho C, Coulouris G, Avagyan V, Ma N, Papadopoulos J, Bealer K, Madden  
355 TL (2009) BLAST+: architecture and applications. *BMC Bioinformatics*  
356 10:421

357 Cowell RK, Coddington JA (1994) Estimating terrestrial biodiversity through

358 extrapolation. *Phil Trans R Soc B* 345:101-118

359 Fukasawa Y, Katsumata S, Mori AS, Osono T, Takeda H (2014) Accumulation and  
360 decay dynamics of coarse woody debris in a Japanese old-growth  
361 subalpine coniferous forest. *Ecol Res* 29:257-269

362 Gates GM, Mohammed C, Wardlaw T, Davidson NJ, Ratkowsky DA (2011)  
363 Diversity and phenology of the macrofungal assemblages supported by  
364 litter in a tall, wet *Eucalyptus obliqua* forest in southern Tasmania,  
365 Australia. *Fungal Ecol* 4:68-75

366 Gómez-Hernández M, Williams-Linera G, Guevara R, Lodge DJ (2012) Patterns of  
367 macromycete community assemblage along an elevation gradient:  
368 options for fungal gradient and metacommunity analyses. *Biodivers*  
369 *Conserv* 21:2247-2268

370 Hansen PA, Tyler G (1992) Statistical evaluation of tree species affinity and soil  
371 preference of the macrofungal flora in south Swedish beech, oak and  
372 hornbeam forest. *Crypt Bot* 2:355-361

373 Hintikka V (1988) On the macromycete flora in oligotrophic pine forests of  
374 different ages in south Finland. *Acta Bot Fennica* 136:89-94

- 375 Hirose D, Osono T (2006) Development and seasonal variations of *Lophodermium*  
376 populations on *Pinus thunbergii* needle litter. *Mycoscience* 47:242-247
- 377 Hongo T (1994) *Fungi*. Yama to Keikoku Sha, Tokyo, Japan (in Japanese)
- 378 Hyodo F, Tayasu I, Wada E (2006) Estimation of the longevity of C in terrestrial  
379 detrital food webs using radiocarbon (<sup>14</sup>C): how old are diets in termites?  
380 *Functional Ecol* 20:385-393
- 381 Ikeda A, Matsuoka S, Masuya H, Mori AS, Hirose D, Osono T (2014) Comparison  
382 of the diversity, composition, and host recurrence of xylariaceous  
383 endophytes in subtropical, cool temperate, and subboreal regions in  
384 Japan. *Popul Ecol* in press
- 385 Imazeki R, Hongo T (1987) *Colored Illustration of Mushrooms of Japan. Vol. I.*  
386 Hoikusha, Tokyo, Japan (in Japanese)
- 387 Imazeki R, Hongo T (1989) *Colored Illustration of Mushrooms of Japan. Vol. II.*  
388 Hoikusha, Tokyo, Japan (in Japanese)
- 389 Imazeki R, Otani Y, Hongo T (1988) *Fungi of Japan*. Yama to Keikoku Sha, Tokyo,  
390 Japan (in Japanese)
- 391 Lange M (1993) Maromycetes under twelve tree species in ten plantations on

392 various soil types in Denmark. *Opera Bot* 120:1-53

393 Levin I, Kromer B (1997) Twenty years of atmospheric  $^{14}\text{CO}_2$  observations at  
394 Schauinsland station, Germany. *Radiocarbon* 39:205-218

395 López-Quintero CA, Straatsma G, Franco-Molano AE, Boekhout T (2012)  
396 Macrofungal diversity in Colombian Amazon forests varies with regions  
397 and regimes of disturbance. *Biodivers Conserv* 21:2221-2243

398 Miyamoto T, Igarashi T, Takahashi K (2000) Lignin-degrading ability of  
399 litter-decomposing basidiomycetes from *Picea* forests of Hokkaido.  
400 *Mycoscience* 41:105-110

401 Mori A, Takeda H (2004) Functional relationships between crown morphology and  
402 within-crown characteristics of understory saplings of three codominant  
403 conifers in a subalpine forest in central Japan. *Tree Physiol* 24:661-670

404 Mueller GM, Schmit JP, Leacock PR, Buyck B, Cifuentes J, Desjardin DE, Halling  
405 RE, Hjortstam K, Iturriaga T, Larsson KH, Lodge DJ, May TW, Minter D,  
406 Rajchenberg M, Redhead SA, Ryvarde L, Trappe JM, Watling R, Wu Q  
407 (2007) Global diversity and distribution of macrofungi. *Biodivers Conserv*  
408 16:37-48



- 409 Murakami Y (1989) Spatial changes of species composition and seasonal fruiting  
410 of the Agaricales in *Castanopsis cuspidata* forest. Trans Mycol Soc Japan  
411 30:89-103
- 412 O'Donnell K (1993) *Fusarium* and its near relatives. In: Reynolds DR, Taylor JW  
413 (eds) The fungal holomorph: mitotic, meiotic and pleomorphic speciation  
414 in fungal systematics. CAB International, Wallingford, UK, pp 225-233
- 415 O'Hanlon R, Harrington TJ (2012) Macrofungal diversity and ecology in four Irish  
416 forest type. Fungal Ecol 5:499-508
- 417 Okabe H (1983) Mycosociological research of Agaricales in natural forests (II)  
418 Seasonal changes on each stand and life form. Bull Kyoto Univ Forest  
419 53:20-32 (in Japanese with English abstract)
- 420 Okabe H (1986) Ecological study of distribution of fungi within forests. PhD thesis,  
421 Kyoto University, Kyoto (in Japanese)
- 422 Oksanen J (2013) Multivariate analysis of ecological communities in R: vegan  
423 tutorial. <http://cc.oulu.fi/~jarioksa/opetus/metodi/vegantutor.pdf>  
424 (accessed 14.5.14)
- 425 Osono T (2007) Ecology of ligninolytic fungi associated with leaf litter

426 decomposition. Ecol Res 22:955-974

427 Osono T (2011) Diversity and functioning of fungi associated with leaf litter  
428 decomposition in Asian forests of different climatic regions. Fungal Ecol  
429 4:375-385

430 Osono T. Mycelial biomass in the forest floor and soil of subtropical, temperate,  
431 and subalpine forests. J For Res:submitted

432 Osono T, Hobara S, Fujiwara S, Koba K, Kameda K (2002) Abundance, diversity,  
433 and species composition of fungal communities in a temperate forest  
434 affected by excreta of the Great Cormorant *Phalacrocorax carbo*. Soil Biol  
435 Biochem 34:1537-1547

436 Osono T, Ono Y, Takeda H (2003) Fungal ingrowth on forest floor and  
437 decomposing needle litter of *Chamaecyparis obtusa* in relation to  
438 resource availability and moisture condition. Soil Biol Biochem  
439 35:1423-1431

440 Osono T, Hobara S, Hishinuma T, Azuma JI (2011a) Selective lignin  
441 decomposition and nitrogen mineralization in forest litter colonized by  
442 *Clitocybe* sp. Eur J Soil Biol 47:114-121

443 Osono T, To-Anun C, Hagiwara Y, Hirose D (2011b) Decomposition of wood, petiole  
444 and leaf litter by *Xylaria* species from northern Thailand. *Fun Ecol*  
445 4:210-218

446 Outerbridge RAM (2002) Macrofungus ecology and diversity under different  
447 conifer monocultures on southern Vancouver Island. PhD thesis,  
448 University of Victoria

449 Peterson SW (2000) Phylogenetic analysis of *Penicillium* species based on ITS and  
450 lsu-rDNA nucleotide sequences. In: Samson RA, Pitt JI (eds) Integration  
451 of modern taxonomic methods for *Penicillium* and *Aspergillus*  
452 classification. Harwood, Amsterdam, the Netherland, pp 163-178

453 Rastin N, Schlechte G, Hüttermann A (1990) Soil macrofungi and some soil  
454 biological, biochemical and chemical investigations on the upper and  
455 lower slope of a spruce forest. *Soil Biol Biochem* 22:1039-1047

456 R Development Core Team (2009) R: a language and environment for statistical  
457 computing. <http://www.r-project.org/> (accessed 14.5.14)

458 Richard F, Moreau PA, Selosse MA, Gardes M (2004) Diversity and fruiting  
459 patterns of ectomycorrhizal and saprobic fungi in an old-growth

- 460 Mediterranean forest dominated by *Quercus ilex* L. Can J Bot  
461 82:1711-1729
- 462 Såstad SM (1995) Fungi - vegetation relationships in a *Pinus sylvestris* forest in  
463 central Norway. Can J Bot 73:807-816
- 464 Schmit JP, Murphy JF, Mueller GM (1999) Macrofungal diversity of a temperate  
465 oak forest: a test of species richness estimators. Can J Bot 77:1014-1027
- 466 Steffen KT, Cajthaml T, Šnajdr J, Baldrian P (2007) Differential degradation of  
467 oak (*Quercus petraea*) leaf litter by litter-decomposing basidiomycetes.  
468 Res Microbiol 158:447-455
- 469 Straatsma G, Ayer F, Egli S (2001) Species richness, abundance, and phenology of  
470 fungal fruit bodies over 21 years in a Swiss forest plot. Mycol Res  
471 105:515-523
- 472 Takeda H, Kaneko N (1988) Patterns of soil humus accumulation in forests. I.  
473 Mull and moder types humus in a broad-leaved forest. Bull Kyoto Univ  
474 Forest 60:33-45 (in Japanese with English abstract)
- 475 Tian X, Takeda H, Ando T (1997) Application of a rapid thin section method for  
476 observations on decomposing litter in mor humus form in a subalpine

477                   coniferous forest. Ecol Res 12:289-300

478 Tsukamoto J (1996) Soil macro-invertebrates and litter disappearance in a  
479                   Japanese mixed deciduous forest and comparison with European  
480                   deciduous forests and tropical rainforests. Ecol Res 11:35-50

481 Tyler G (1985) Macrofungal flora of Swedish beech forest related to soil organic  
482                   matter and acidity characteristics. For Ecol Manag 10:13-29

483 Valášková V, Šnajdr J, Bittner B, Cajthaml T, Merhautová V, Hofrichter M,  
484                   Baldrian P (2007) Production of lignocellulose-degrading enzymes and  
485                   degradation of leaf litter by saprotrophic basidiomycetes isolated from a  
486                   *Quercus petraea* forest. Soil Biol Biochem 39:2651-2660

487 Van der Wal A, Geydan TD, Kuyper TW, de Boer W (2013) A thready affair:  
488                   linking fungal diversity and community dynamics to terrestrial  
489                   decomposition processes. FEMS Microbiol Rev 37:477-494

490 Villeneuve N, Grandtner MM, Fortin JA (1989) Frequency and diversity of  
491                   ectomycorrhizal and saprophytic macrofungi in the Laurentide  
492                   Mountains of Quebec. Can J Bot 67:2616-2629

493 White TJ, Bruns T, Lee S, Taylor JW (1990) Amplification and direct sequencing

494 of fungal ribosomal RNA genes for phylogenetics. In: Innis MA, Gelfand  
495 DH, Sninsky JJ, White TJ (eds) PCR Protocols: a Guide to Methods and  
496 Applications. Academic Press, New York, USA, pp 315-322

497 Xu X, Tokashiki Y, Enoki T, Hirata E (1998a) Characteristics of nutrient  
498 accumulation in forest floor under evergreen broadleaved forests in  
499 Okinawa Island. Sci Bull Fac Agr Univ Ryukyus 45:185-193

500 Xu X, Enoki T, Tokashiki Y, Hirata E (1998b) Litterfall and the nutrient returns in  
501 evergreen broadleaved forests in Northern Okinawa Island. Sci Bull Fac  
502 Agr Univ Ryukyus 45:195-208

503 Yamashita S, Hijii N (2004) Relationships between seasonal appearance and  
504 longevity of fruitbodies of Agaricales and meteorological factors in a  
505 Japanese red pine forest. J For Res 9:165-171

Osono Table 1

Table 1. Forest floor and field survey of fruiting bodies in the study sites.

Site	ST	CT	SA
Humus type	Mull	Mull (lower), Moder (upper) <sup>a</sup>	Moder <sup>b</sup>
Depth of L layer (cm) <sup>c</sup>	1.1 ± 0.1	1.2 ± 0.1	2.7 ± 0.3
Depth of FH layer (cm) <sup>c</sup>	1.0 ± 0.2	4.0 ± 0.4	13.4 ± 1.2
Forest floor mass (Mg/ha)	12.0 <sup>d</sup>	7.7 (lower) 33.3 (upper) <sup>e</sup>	104.6 <sup>f</sup>
Leaf fall mass (Mg/ha/yr)	7.95 <sup>d</sup>	4.10 (lower) 3.20 (upper) <sup>e</sup>	3.59 <sup>f</sup>
Turnover time (yr) <sup>g</sup>	1.5	1.9 (lower) 10.4 (upper)	29.1

<sup>a</sup>Takeda and Kaneko (1988). <sup>b</sup>Tian et al. (1997). <sup>c</sup>Values are means ± standard errors (n=20). Measurement was carried out in the three study sites in October 2012. Values of CT were from the lower slope. <sup>d</sup>Xu et al. (1998a, 1998b). <sup>e</sup>Tsukamoto (1996). <sup>f</sup>Fukasawa et al. (2014). <sup>g</sup>Turnover time = forest floor mass / annual leaf fall mass.

Osono Table 2

Table 2. Assemblage structure and family composition of macrofungi and the soil layer from which fruiting bodies occurred. Numbers of macrofungal species are indicated for fungal families and the litter layers. Numbers in parentheses indicate the proportion relative to the observed number of species. Values of  $S_{est}$  indicate means  $\pm$  standard deviations. See text for  $S_{est}$ .

	ST	CT (lower)	CT (upper)	SA
<b>Diversity</b>				
Observed number of species ( $S_{obs}$ )	35	25	11	18
Singleton species	17 (49)	11 (44)	4 (36)	3 (17)
Simpson's D	7.68	4.47	3.40	4.86
Equitability	0.22	0.18	0.31	0.27
Estimated number of species ( $S_{est}$ )	14.2 $\pm$ 2.1	17.2 $\pm$ 2.0	11.0 $\pm$ 0.0	11.9 $\pm$ 1.7
<b>Family composition</b>				
Mycenaceae	13 (37)	10 (40)	3 (27)	9 (50)
Marasmiaceae	12 (34)	3 (12)	3 (27)	2 (11)
Agaricaceae	3 (9)	6 (24)	2 (18)	1 (6)
Tricholomataceae	5 (14)	2 (8)	0 (0)	3 (17)
Strophariaceae	0 (0)	1 (4)	2 (18)	0 (0)
Psathyrellaceae	0 (0)	2 (8)	0 (0)	0 (0)
Pluteaceae	0 (0)	1 (4)	0 (0)	0 (0)
Hygrophoraceae	0 (0)	0 (0)	1 (9)	0 (0)
Hymenogasteraceae	0 (0)	0 (0)	0 (0)	1 (6)
Xylariaceae	1 (3)	0 (0)	0 (0)	0 (0)
Unidentified	1 (3)	0 (0)	0 (0)	2 (11)
<b>Soil layer</b>				
L layer	26 (74)	0 (0)	0 (0)	0 (0)
L-F border	9 (26)	23 (92)	8 (73)	3 (17)
F layer	0 (0)	0 (0)	1 (9)	14 (78)
A layer	0 (0)	2 (8)	2 (18)	1 (6)



Osono Table 3

Table 3. Radiocarbon content in fruiting bodies of macrofungi. Samples from CT were from the lower slope.

	Species	Soil layer	Collection date	Laboratory code	$\delta^{13}\text{C}$	$\Delta^{14}\text{C}$	Carbon age (yr)
ST	<i>Mycena</i> sp. ST1	L layer	Apr 11	IAAA-111556	$-27.3 \pm 0.5$	$48.2 \pm 2.8$	2.8
	<i>Gymnopus</i> sp. ST1	L layer	Oct 11	IAAA-111557	$-30.4 \pm 0.5$	$52.0 \pm 3.0$	4.5
	<i>Marasmius</i> sp. ST4	L-F border	Oct 11	IAAA-111558	$-22.9 \pm 0.4$	$68.6 \pm 2.7$	9.0
CT	<i>Mycena polygramma</i>	L-F border	Oct 01	IAAA-81685	$-25.5 \pm 0.3$	$132.6 \pm 3.7$	9.5
	<i>Mycena amygdalina</i>	L-F border	Nov 11	IAAA-111554	$-20.2 \pm 0.4$	$79.6 \pm 2.8$	11.4
	<i>Gymnopus peronatus</i>	L-F border	Nov 01	IAAA-111555	$-24.3 \pm 0.4$	$91.0 \pm 2.7$	3.6
SA	<i>Mycena aurantiidisca</i>	L-F border	Oct 08	IAAA-81687	$-24.9 \pm 0.4$	$96.0 \pm 3.8$	11.3
	<i>Mycena epipterygia</i>	L-F border	Oct 11	IAAA-111552	$-23.0 \pm 0.4$	$139.7 \pm 2.8$	20.3
	<i>Galerina atkinsoniana</i>	L-F border	Oct 11	IAAA-111553	$-28.9 \pm 0.4$	$56.7 \pm 2.7$	5.9

Osono Table 4

Table 4. Size of fruiting bodies of macrofungi. DP, diameter of pileus; LS, length of stipe. Values are means  $\pm$  standard errors in cm. Numbers in parentheses indicate the number of samples. The same letters indicate that the values are not significantly different at 5% level by Tukey's HSD test. Data from CT are from the lower slope.

	Total		Major species 1		Major species 2		Major species 3	
	DP	LS	DP	LS	DP	LS	DP	LS
ST	3.7 $\pm$ 1.3 b	13.9 $\pm$ 2.3 c	1.6 $\pm$ 0.3	11.3 $\pm$ 1.1	4.6 $\pm$ 0.7	12.0 $\pm$ 0.8	22.5	46.0
	(16)		<i>Mycena</i> sp. ST1 and ST2 (10)		<i>Gymnopus</i> sp. ST1 (4)		<i>Marasmius</i> sp. ST4 (1)	
CT	12.0 $\pm$ 1.2 a	63.8 $\pm$ 5.7 a	3.8	30.0	14.0	116.5	15.8 $\pm$ 2.3	73.5 $\pm$ 7.0
	(23)		<i>Mycena filopes</i> (2)		<i>Mycena polygramma</i> (2)		<i>Mycena crocata</i> (6)	
SA	5.8 $\pm$ 0.5 b	35.1 $\pm$ 2.1 b	4.1 $\pm$ 0.3	30.9 $\pm$ 2.3	6.9 $\pm$ 0.5	46.9 $\pm$ 5.6	4.6 $\pm$ 0.7	25.5 $\pm$ 3.3
	(38)		<i>Mycena aurantiidisca</i> (15)		<i>Mycena epipterygia</i> (6)		<i>Galerina atkinsoniana</i> (8)	

1 Figure legends

2

3 Fig. 1. Rank-relative frequency distribution of macrofungal assemblages in  
4 subtropical (ST), cool temperate (CT), and subalpine forests (SA). Open,  
5 Mycenaceae; filled, Marasmiaceae; coarse oblique mesh, Agaricaceae; fine oblique  
6 mesh, Tricholomataceae; horizontal mesh, others (Hygrophoraceae,  
7 Hymenogasteraceae, Pluteaceae, Psathyrellaceae, Strophariaceae, Xylariaceae,  
8 and unidentified). The survey in CT was performed at lower and upper parts of a  
9 forest slope.

10

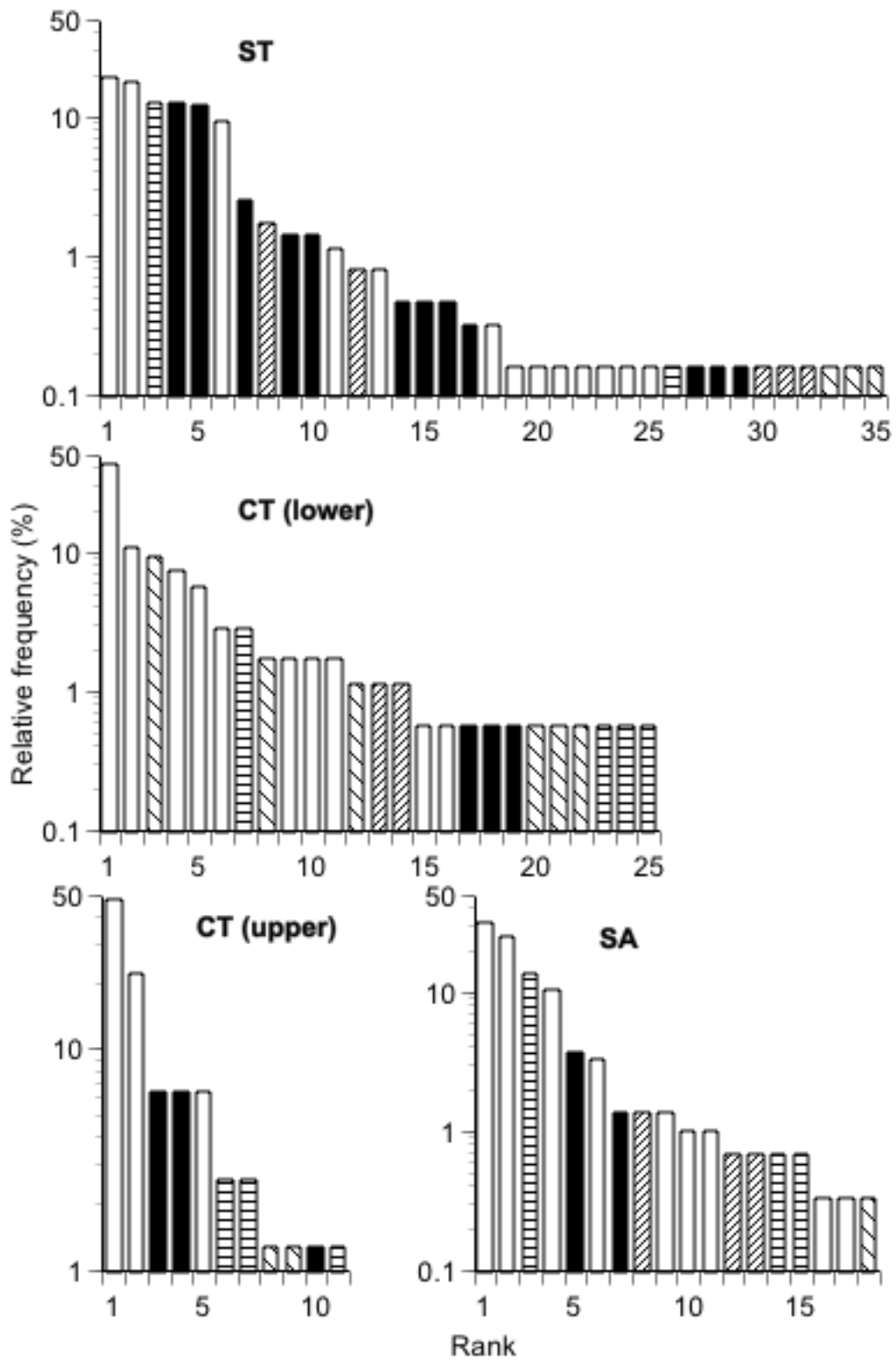
11 Fig. 2. Seasonal changes in the frequency of occurrence (upper) and the number of  
12 species (lower) of fruiting bodies of macrofungi. □, subtropical forest (ST); ○,  
13 lower part of a slope in cool temperate forest, [CT (lower)]; ●, upper part of a slope  
14 in cool temperate forest [CT (upper)]; ▲, subalpine forest (SA).

15

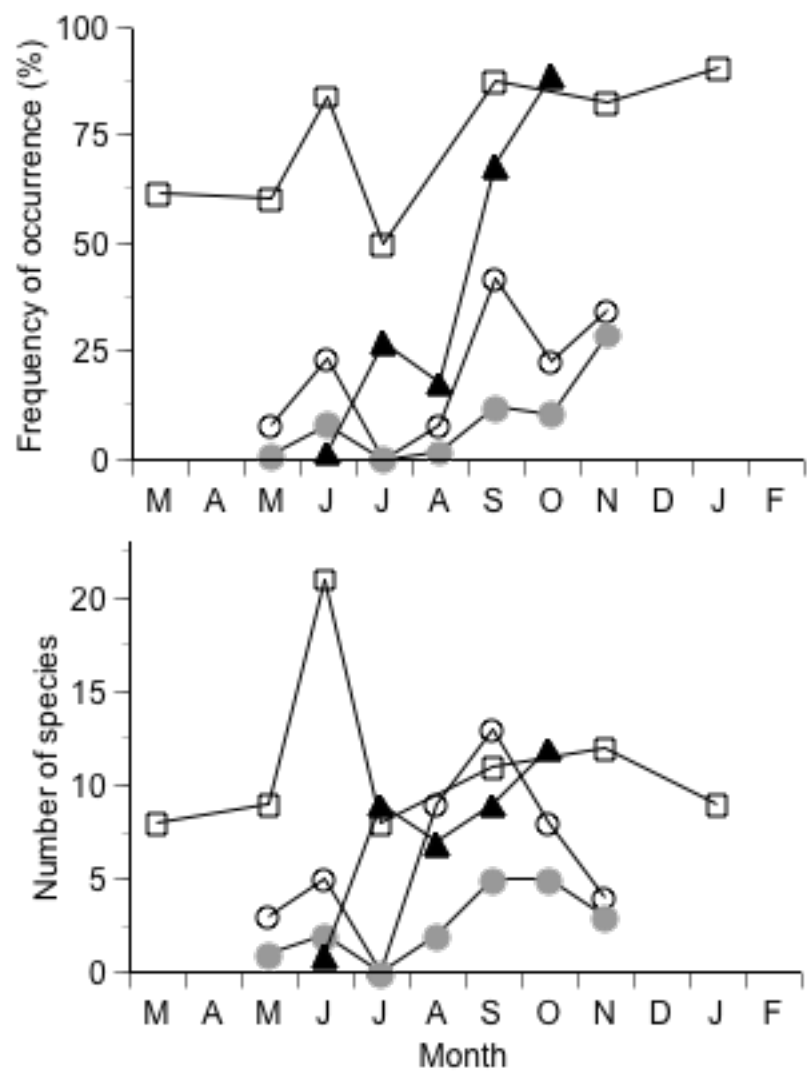
16 Fig. 3. Seasonal changes in the frequency of occurrence of major macrofungal  
17 species in subtropical forest (ST), cool temperate forest (CT), and subalpine forest  
18 (SA). For CT, blank and shaded bars indicate lower and upper slopes, respectively.

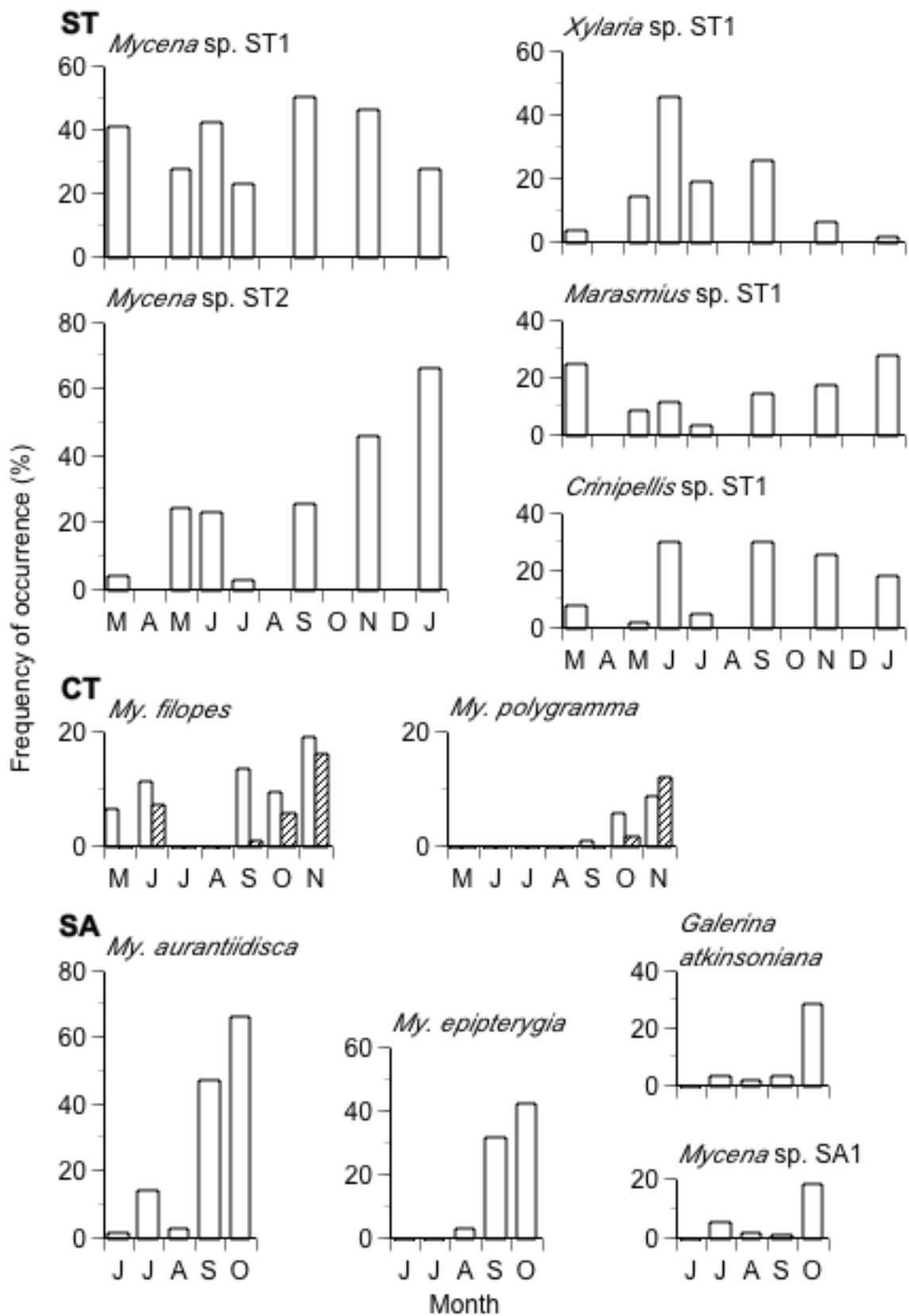
19

1 Osono Fig. 1



1 Osono Fig. 2  
2





# Electronic Supplementary Material

## Patterns in diversity, resource utilization, and phenology of fruiting bodies of litter-decomposing macrofungi in subtropical, temperate, and subalpine forests

Takashi Osono

**Table S1: Frequency of occurrence of fruiting bodies of macrofungi and the soil layer from which the fruiting bodies occurred. Ag, Agaricaceae; Hg, Hygrophoraceae; Hm, Hymenogasteraceae; Mr, Marasmiaceae; My, Mycenaceae; Tr, Tricholomataceae; Pl, Pluteaceae; Ps, Psathyrellaceae; St, Strophariaceae; Xy, Xylariaceae; and Un, unidentified. L/F, L-F border.**

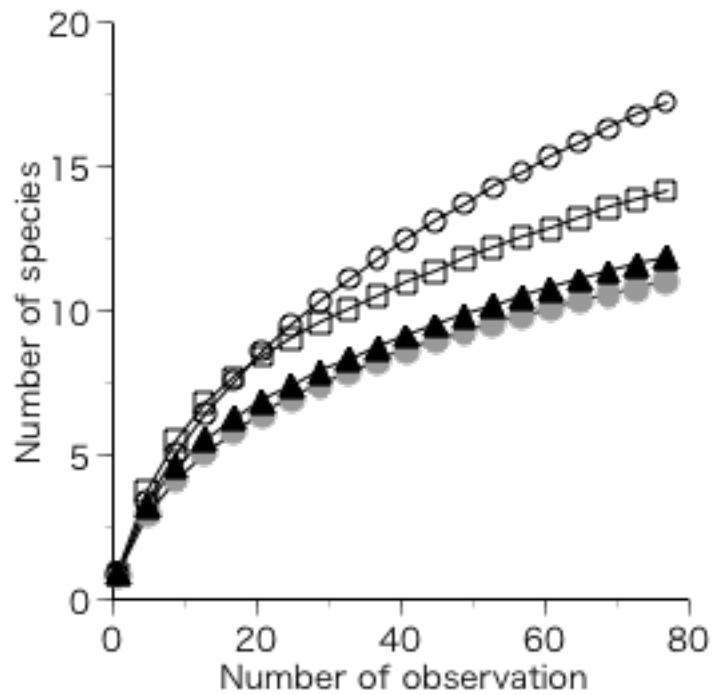
Taxa	Family	Soil layer	Frequency (%)
<b>Subtropical forest</b>			
<i>Mycena</i> section Basipedes 'sp. ST1'	My	L	96.8
<i>Mycena</i> section Roridae 'sp. ST2'	My	L	90.4
<i>Xylaria</i> spp. ST1	Xy	L	64.8
<i>Marasmius</i> spp. ST1	Mr	L	63.2
<i>Crinipellis</i> sp. ST1	Mr	L	60.8
<i>Mycena</i> sp. ST3	My	L	46.4
<i>Gymnopus</i> sp. ST1	Mr	L	12.8
<i>Tricholomataceae</i> sp. ST1	Tr	A	8.8
<i>Gymnopus</i> sp. ST2	Mr	L	7.2
<i>Marasmiellus</i> sp. ST1	Mr	L	7.2
<i>Mycena</i> sp. ST4	My	L	5.6
<i>Tricholomataceae</i> sp. ST2	Tr	A	4.0
<i>Mycena</i> sp. ST5	My	L	4.0
<i>Crinipellis</i> sp. ST2	Mr	L	2.4
<i>Gymnopus</i> sp. ST3	Mr	L	2.4
cf. <i>Calyprella</i> sp. ST1	Mr	L	2.4

<i>Marasmius</i> sp. ST2	Mr	L	1.6	
<i>Mycena</i> sp. ST6	My	L	1.6	
Agaricaceae sp. ST1	Ag	A	0.8	
Agaricaceae sp. ST2	Ag	A	0.8	
<i>Gymnopus</i> sp. ST4	Mr	L	0.8	
<i>Mycena</i> sp. ST7	My	L	0.8	
<i>Leucocorpinus</i> sp. ST1	Ag	A	0.8	
<i>Marasmiellus</i> sp. ST2	Mr	A	0.8	
<i>Marasmiellus</i> sp. ST3	Mr	L	0.8	
<i>Mycena</i> sp. ST8	My	L	0.8	
<i>Mycena</i> sp. ST9	My	L	0.8	
<i>Mycena</i> sp. ST10	My	L	0.8	
Unidentified ST1	Un	L	0.8	
<i>Mycena</i> sp. ST11	My	A	0.8	
<i>Mycena</i> sp. ST12	My	L	0.8	
Tricholomataceae sp. ST3	Tr	A	0.8	
Tricholomataceae sp. ST4	Tr	A	0.8	
Tricholomataceae sp. ST5	Tr	L	0.8	
<i>Xeromphalina</i> sp. ST1	My	L	0.8	
<b>Cool temperate forest</b>			<b>Upper</b>	<b>Lower</b>
<i>Mycena amygdalina</i>	My	L/F	29.6	60
<i>Mycena polygramma</i>	My	L/F	13.6	15.2
<i>Gymnopus peronatus</i>	Mr	L/F	4.0	0.8
<i>Gymnopus</i> sp. CT1	Mr	L/F	4.0	0.0
<i>Mycena</i> sp. CT2	My	L/F	4.0	0.0
<i>Stropharia aeruginosa</i>	St	L/F	1.6	0.8
<i>Hygrocybe cantharellus</i>	Hy	L/F	1.6	0.0
Agaricaceae sp. CT1	Ag	L/F	0.8	0.0
<i>Lepiota fusciceps</i>	Ag	F	0.8	0.0
<i>Marasmius</i> sp. CT2	Mr	A	0.8	0.0
<i>Naematoloma sublateritium</i>	St	A	0.8	0.0
<i>Lycoperdon perlatum</i>	Ag	A	0.0	12.8



<i>Mycena pura</i>	My	L/F	0.0	10.4
<i>Mycena</i> sp. CT1	My	L/F	0.0	8.0
<i>Mycena amicta</i>	My	L/F	0.0	4.0
<i>Psathyrella candolleana</i>	Ps	L/F	0.0	4.0
Agariaceae sp. CT2	Ag	L/F	0.0	2.4
<i>Mycena crocata</i>	My	L/F	0.0	2.4
<i>Mycena luteopallens</i>	My	L/F	0.0	2.4
<i>Mycena</i> sp. CT3	My	L/F	0.0	2.4
<i>Agaricus praeclaresquamosus</i>	Ag	L/F	0.0	1.6
<i>Clitocybe</i> sp. CT1	Mr	L/F	0.0	1.6
<i>Pseudoclitocybe cyathiformis</i>	Tr	L/F	0.0	1.6
<i>Lepiota</i> cf. <i>pseudogranulosa</i>	Ag	L/F	0.0	0.8
<i>Lepiota cygnea</i>	Ag	L/F	0.0	0.8
<i>Lepiota</i> sp. CT1	Ag	A	0.0	0.8
<i>Marasmius pulcheriipes</i>	Mr	L/F	0.0	0.8
<i>Marasmius</i> sp. CT1	Mr	L/F	0.0	0.8
<i>Mycena</i> cf. <i>osmundicola</i>	My	L/F	0.0	0.8
<i>Mycena</i> sp. CT4	My	L/F	0.0	0.8
<i>Psathyrella piluliformis</i>	Ps	L/F	0.0	0.8
<i>Volvariella speciosa</i> var. <i>gloiocephala</i>	Pl	L/F	0.0	0.8
<b>Subalpine forest</b>				
<i>Mycena aurantiidisca</i>	My	F	75.2	
<i>Mycena epipterygia</i>	My	F	60.8	
<i>Galerina atkinsoniana</i>	Hm	F	32.8	
<i>Mycena</i> cf. <i>filopes</i>	My	L/F	24.8	
<i>Clitocybe</i> sp. SA1	Mr	F	8.8	
<i>Mycena</i> sp. SA2	My	L/F	8.0	
<i>Marasmius androsaceus</i>	Mr	L/F	3.2	
Tricholomataceae sp. SA1	Tr	F	3.2	
<i>Mycena</i> cf. <i>stipata</i>	My	F	3.2	
<i>Mycena</i> sp. SA3	My	F	2.4	
<i>Mycena</i> cf. <i>pura</i>	My	F	2.4	

<i>Collybia cookei</i>	Tr	F	1.6
Tricholomataceae sp. SA2	Tr	F	1.6
Unidentified SA1	Un	F	1.6
Unidentified SA2	Un	F	1.6
<i>Mycena</i> sp. SA5	My	F	0.8
<i>Mycena</i> sp. SA4	My	F	0.8
<i>Lycoperdon perlatum</i>	Ag	A	0.8



**Fig. S2. Rarefaction curves for litter-decomposing macrofungal (LDM) assemblages. □, subtropical forest (ST); ○, lower part of a slope in cool temperate forest, [CT (lower)]; ●, upper part of a slope in cool temperate forest [CT (upper)]; ▲, subalpine forest (SA).**