# 1 Impact of Feed Water pH and Membrane Material on Nanofiltration of Perfluorohexanoic Acid

- 2 in Aqueous Solution
- 3
- 4 Chenghui Zeng<sup>1</sup>, Shuhei Tanaka<sup>2</sup>, Yuji Suzuki<sup>2</sup>, Shigeo Fujii<sup>2</sup>
- 5 1. Graduate School of Engineering, Kyoto University, Yoshida, Sakyo-ku, Kyoto, Japan
- 6 2. Graduate School of Global Environmental Studies, Kyoto University, Yoshida, Sakyo-ku, Kyoto,
- 7 Japan
- 8

## 9 **Graphical abstract**



# 16 Highlights

17	•	PFHxA in ac	queous solution	was concentrated	by two	kinds of NI	F membranes.
					~		

- 18 Polyamide and sulfonated polyethersulfone NF membranes were used.
- 19 Feed water pH showed different impact on these two kinds of NF membranes.
- The membrane flux or PFHxA rejection rate deteriorated at acidic pH.
- Increasing the pH can enhance flux and rejection performance of NF membranes.

#### 23 Abstract

24Nanofiltration was thought to be a good option for the recovery of perfluohexanoic acid (PFHxA) from industrial wastewater. In this study, two commercially available nanofiltration (NF) 2526membranes (NF 270 and NTR-7450) were tested to concentrate the PFHxA in aqueous solution. Filtration test was conducted in crossflow filtration mode. Membrane flux and PFHxA rejection rate 2728were monitored throughout the filtration test. The impact of initial feed water pH on membrane 29performance was investigated. Results demonstrated that the two NF membranes showed different 30 response to the change of initial feed water pH, which was caused by the intrinsic properties of 31membrane material. The flux performance of NF 270 was stable, while its rejection rate of PFHxA 32was very sensitive to the change of initial feed water pH. Opposite result was obtained with NTR-7450. It has a very good stability on rejection rate, while its flux is very sensitive to the change 33 34of initial feed water pH. The mechanisms behind these phenomena were also discussed. The results 35obtained in this study should be very useful for the process design in practical engineering. 36

Keywords: perfluorohexanoic acid (PFHxA), recovery, nanofiltration (NF), polyamide membrane,
 sulfonated polyethersulfone membrane, pH

39 Corresponding author: Chenghui Zeng, Email: zeng.chenghui.35z@st.kyoto-u.ac.jp

40

#### 42 **1. Introduction**

43Per- and polyfluoroalkyl substances (PFASs) are a group of anthropogenic compounds, which have 44 been widely used in industries due to their unique characteristics, such as surface activity, thermal and chemical stability, etc (Herzke et al., 2012; Wang et al., 2013). These compounds have been 4546detected in the environment ubiquitously, including wastewater treatment plants, water purification plant, lakes, sea water and even arctic area, etc (Kunacheva et al., 2011; Yu et al., 2013; Rahman et 4748al., 2014; Lu et al., 2015; Guo et al., 2016). In recent years, they have received continuously 49increasing concern as they found to be toxic, persistent and bioaccummulative (Kannan et al., 2004; 50Yeung et al., 2006; Fujii et al., 2007; Svihlikova et al., 2015). Industrial discharge was considered as one of the main sources of PFASs pollution in water environment (Prevedouros et al., 2006). In 51order to provide protection to human health and environment, governments and international 5253organizations have been involved in phasing out the production and use of PFASs globally (ECHA, 542013; US EPA, 2006). PFASs-related industries are also seeking cost-effective techniques on the recovery of PFASs from industrial wastewater so that the discharge of PFASs to the environment 55could be reduced. Membrane technology has been thought to be a good option for the recovery of 56PFASs from industrial wastewater (Tang et al., 2006; Tang et al., 2007). 57

58

Among all membranes ranging from reverse osmosis (RO) to microfiltration (MF), nanofiltration
(NF) membrane has a molecular weight cut-off ranging from 200 to 1,000 Da. It offers higher flux

61	than reverse osmosis (RO) and better retention performance than ultrafiltration (UF) for treating
62	wastewater containing small molecule. These advantages make NF membrane became very
63	promising in the recovery of PFASs from industrial wastewater. Before, most of the researches had
64	focused on rejecting PFOA and PFOS (C8) by NF or RO membrane as they were the main kinds of
65	PFASs used in industries (Tang et al., 2006; Tang et al., 2007; Zhao et al., 2013; Hang et al., 2015;
66	Zhao et al., 2016). As the regulations and legislation on the use and manufacture of PFOA and
67	PFOS become more stringent, some industries have shifted to use the shorter chain PFASs as the
68	alternative (Wang et al., 2013; Wang et al., 2017). For example, a type of short chain PFASs-
69	perfluorohexanoic acid (CF <sub>3</sub> (CF <sub>2</sub> ) <sub>4</sub> COOH, PFHxA)- has been detected in water environment with its
70	concentration as high as 16,000 ng/L (Takemine et al., 2014), which suggest that some industries have
71	started to use PFHxA as the alternative of PFOS and PFOA. PFASs with different chain lengths have
72	different physicochemical properties (Rahman et al., 2014). Thus, membrane's rejection properties to
73	them might be also different. However, limited information on rejection of PFHxA by NF membranes
74	is available.

A typical structure of NF membrane consists of three parts: top selective layer, microporous substrate and nonwoven fabric. The separation properties of NF membrane mostly depend on the selective layer. Generally, the selective layer of commercialized NF membranes was made from two types of materials. One is polyamide (PA) and another one is sulfonated polyethersulfone (SPES).

80	These two kinds of membranes showed different response to the change of feed water pH (Mänttäri
81	et al., 2006). Thus, they might also have different rejection properties to PFHxA in water.
82	Furthermore, most of the previous researches were conducted with a PA membrane - NF 270
83	(Filmtech, USA), which showed a relatively high flux and salt rejection rate among commercialized
84	NF membranes (Tang et al., 2009). The rejection properties of SPES NF membrane to PFHxA in
85	wastewater were still not well investigated.
86	
87	This paper describes an investigation on the recovery of PFHxA in pure waterby both PA and SPES
88	NF membrane. The impact of pH on PFHxA rejection rate and membrane flux was investigated.
89	PFHxA aqueous solution was used and pH of the feed water was adjusted by NaOH. Filtration test
90	was conducted in crossflow filtration mode. PFHxA concentration in water samples was measured
91	by High-performance Liquid Chromatography coupled with tandem mass spectrometer (1200 series,
92	HPLC/MS/MS, Agilent Technologies, USA).
93	
94	2. Materials and Methods
95	2.1 Chemicals and Materials
96	2.1.1 Chemicals
97	All chemicals used for PFHxA analysis were of HPLC/MS/MS grade unless otherwise specified.
98	Perfluorohexanoic acid, methanol, acetonitrile, ammonium acetate solution (1 mole/L) were purchased

99	from Wako Pure Chemical (Japan). Milli-Q water purified by PresepC-Agri (C18) cartridge (Wako,
100	Japan) and Oasis <sup>®</sup> HLB cartridge (Waters, Japan) was used as the solvent for PFHxA analysis in order
101	to avoid PFASs contamination.
102	
103	2.1.2 NF membranes
104	Two membranes were used in this study: NTR-7450 was purchased from Nitto Denko (Japan) through
105	Membrane Soltech (Japan), NF 270 was supplied by Dow/Filmtech (USA) through Mikuniki (Japan).
106	NTR-7450 was obtained and stored as dry coupons, while NF 270 was supplied and stored in water.
107	All the coupons were stored in refrigerator (5 °C) once arrived to our lab. The information of these
108	two membranes were summarized in Table 1.

## 110 **Table 1 Characteristics of membranes used in this research**

111

Membrane	NF 270	NTR-7450
MWCO (Da) <sup>(a)</sup>	200	600 - 800
zeta-potential at pH 7.0 (mV) <sup>(a)</sup>	- 21.6	- 16.6
Pure water permeability $(m^3/(m^2 \cdot MPa^1 \cdot d^1))$	3.92	2.62
NaCl rejection rate (%)	58	55
Material of selective layer	polyamide	sulfonated polyethersulfone
Contact angle (°)	$32.6\pm1.3^{(b)}$	$69.6\pm5.9^{(c)}$
Manufacturer	FilmTech, USA	Nitto Denko, Japan

112 (a) The molecular weight cut-off (MWCO) of membrane, zeta-potential were obtained from literature

113 (Braeken et al., 2006); (b) Contact angle of NF 270 was adopted from (Tang et al., 2009); (c) Contact

angle of NTR-7450 was adopted from (Boussu et al., 2005); (d) The pure water permeability and NaCl

rejection rate were measured in this study. The test conditions for NaCl rejection rate measurement:

116 NaCl 2000 mg/L, 1 L/min and 0.7 MPa. The experimental set-up was the same with filtration

117 experiment.

### 118 **2.2 Experimental set-up and filtration test**

119 2.2.1 Experimental set-up

120 The experimental set-up consisted of two customized membrane test set-ups (FTU-1 and C-10T)

121 purchased from Membrane Soltech, Japan. More details of these two set-ups can be found on the

122 website of manufacturer (Membrane SolTech. 2017). The whole system was schematically described

123 in **Fig. 1**. Basically, it has four parts: feed tank, permeate tank, diaphragm pump and membrane cell.



All of them were made from non-PFASs adsorptive material. The feed water was pumped into the membrane cell by a diaphragm pump. The transmembrane pressure (TMP) and crossflow rate were kept at 1 L/min and 0.7 MPa, respectively, by adjusting the bypass valve manually throughout the whole filtration test. Permeate was discharged to the permeate tank directly and the retentate was recycled back to the feed tank. The solutes in feed tank were concentrated throughout the filtration test.

133The whole membrane set-up was put in an air-conditioned room so the feed water temperature was134almost kept constant  $(29 - 30 \ ^{\circ}C)$  during the filtration test.

135

136 2.2.2 Filtration test

In this study, synthetic wastewater prepared by adding PFHxA into the pure water was used for the 137138filtration test. Pure water was supplied by an ion exchanger system (Kurita MD-05, Japan), in which 139most of the matrix in tap water was removed. The initial concentration of PFHxA in synthetic 140wastewater was about 100 mg/L, which is a typical concentration for PFHxA in industrial wastewater (Karnwadee W., 2015). For each filtration test, 20 L synthetic wastewater was added into the feed tank. 141142Also, a new membrane coupon (60 cm<sup>2</sup>) was used, which was pretreated by filtration with pure water 143at 0.7 MPa for at least 1 hour. In order to know the impact of initial feed water pH on membranes performance, the filtration tests were conducted with feed water at three initial pH values: pH 3.3 144(only PFHxA was added), pH 7.0 (addition with PFHxA and NaOH) and pH 10.0 (addition with 145146 PFHxA and NaOH).

147

Throughout the filtration, membrane flux, pH, temperature were monitored. Membrane flux was measured gravimetrically by weighing the mass of permeate collected at predetermined intervals. pH of feed and permeate was monitored by B-172 pH meter (Horiba Ltd., Japan). Temperature was measured by CT-300WP digital thermal meter (Custom Corp., Japan). To obtain the rejection rates of

152	PFHxA, the feed and permeate samples were collected continuously throughout the whole filtration	on
153	test. The overall rejection rate of PFHxA by membrane was calculated by following equation:	
154		
155	Rejection rate (%) = $(1 - C_p/C_f) \times 100$ (1	)
156	Where C <sub>p</sub> is the concentration of PFHxA in permeate (mg/L);	
157	$C_{\rm f}$ is the concentration of PFHxA in feed (mg/L).	
158		
159	In this study, normalized flux was used to evaluate the change of membrane flux throughout the	
160	filtration test. It was defined as:	
161		
162	Normalized flux (%) = $Q_t \times 100/Q_0$ (2)	.)
163	Where $Q_t$ is the membrane flux at filtration time t (m/d);	
164	$Q_0$ is the membrane flux at the beginning of filtration (m/d).	
165		
166	2.3 Analytical method for PFHxA	
167	HPLC/MS/MS was used for PFHxA analysis. In this study, both feed and permeate samples we	ere
168	diluted by 40% acetonitrile solution to a desired concentration before analyzing by HPLC/MS/N	IS.
169	Chromatographic separation was conducted on a Zorbax Eclipse Plus C18 (Agilent Technologi	es,
170	USA) and the column temperature was set at 40 °C. The mobile phase used in this study was	s a

171	mixture of solvent A (acetonitrile) and B (5 mmol/L ammonium acetate) and the flow rate was set at
172	0.25 mL/min. The eluent started from 30% solvent A, increased to 45% at 1.0 min, then 60% at 4
173	min, held at 60% for 0.5 min, went up to 90% at 4.5 min, kept at 90% for 3.5 min, and then ramped
174	down to 30%. The post time was 5 min and the total running time for one sample was 13 min. Triple
175	Quadrupole (Agilent Technologies, USA) mass spectrometer (MS/MS) was used for quantitative
176	analysis. Mass spectrometer was operated with the electrospray ionization (ESI) in negative mode.
177	Analyte ion was monitored by using multiple reaction monitoring (MRM) mode. The $m/z$ for parent
178	ion and daughter ion were 313 and 269, respectively. Collision energy was set at 2 V.
179	
180	3. Results and discussions
181	
	3.1 Rejection rates of PFHxA by two membranes
182	<b>3.1 Rejection rates of PFHxA by two membranes</b> In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF
182 183	3.1 Rejection rates of PFHxA by two membranes In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF membranes, a series of experiments was carried out with feed water at different pH. In Fig. 2, the
182 183 184	3.1 Rejection rates of PFHxA by two membranes In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF membranes, a series of experiments was carried out with feed water at different pH. In Fig. 2, the rejection rates of PFHxA by NF 270 and NTR-7450 under different test conditions were shown.
182 183 184 185	3.1 Rejection rates of PFHxA by two membranes In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF membranes, a series of experiments was carried out with feed water at different pH. In Fig. 2, the rejection rates of PFHxA by NF 270 and NTR-7450 under different test conditions were shown. When only PFHxA was added into the feed tank (initial pH: 3.3), the rejection rate of PFHxA by NF
182 183 184 185 186	<ul> <li>3.1 Rejection rates of PFHxA by two membranes</li> <li>In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF membranes, a series of experiments was carried out with feed water at different pH. In Fig. 2, the rejection rates of PFHxA by NF 270 and NTR-7450 under different test conditions were shown.</li> <li>When only PFHxA was added into the feed tank (initial pH: 3.3), the rejection rate of PFHxA by NF 270 was quite low. Initially, the rejection rate of PFHxA was around 87%. Then, it started to</li> </ul>
182 183 184 185 186 187	3.1 Rejection rates of PFHxA by two membranes In order to investigate the impact of initial feed water pH on the rejection rates of PFHxA by NF membranes, a series of experiments was carried out with feed water at different pH. In Fig. 2, the rejection rates of PFHxA by NF 270 and NTR-7450 under different test conditions were shown. When only PFHxA was added into the feed tank (initial pH: 3.3), the rejection rate of PFHxA by NF 270 was quite low. Initially, the rejection rate of PFHxA was around 87%. Then, it started to decrease along with the concentration increase in feed tank. Finally, it kept constant at about

189 pH of feed water was adjusted to 7.0 by adding NaOH solution, the initial rejection rate of PFHxA

190	by NF 270 increased to 96%. When the feed concentration of PFHxA became higher than 400 mg/L,
191	rejection rate was enhanced further (>98%). After increasing initial pH of feed water to 10.0, the
192	rejection rate of PFHxA by NF 270 also increased further. The initial rejection rate was 97%
193	initially and maintained at 99% at last. Based on above discussions, it can be concluded that NF 270
194	could reject the PFHxA in pure water more effectively at neutral or alkaline conditions than that at
195	acidic conditions. However, for NTR-7450, its rejection rate to PFHxA seems much less sensitive to
196	the change of initial pH of feed water. Under all three pH conditions tested, the rejection rates were
197	maintained above 96%.



Fig. 2 The rejection rates of PFHxA by NF 270 and NTR-7450 throughout the filtration tests at
 different initial feed water pH conditions

204	According to the information shown in Table 1, NF 270 is tighter than NTR-7450. However, its
205	rejection rate to PFHxA was lower than that of NTR-7450 at acidic condition (pH: 3.3). This
206	suggested that the rejection mechanism of PFHxA by NF membranes was not just dependent on the
207	size exclusion. The interaction between PFHxA and membrane surface might also play an important
208	role. Additionally, NF 270 also showed a higher sensitivity to the change of initial feed water pH
209	than NTR-7450. This could be explained by the change of interaction between PFHxA and
210	membrane surface at different initial feed water pH. On one hand, PFHxA is a kind of strong acid
211	and has pKa value of -0.16 (Steinle-Darling and Reinhard, 2008). When only PFHxA was added into
212	the feed water (initial pH: 3.3), it can present in water by both ionic ( $CF_3(CF_2)_4COO^-$ ) and non-ionic
213	(CF <sub>3</sub> (CF <sub>2</sub> ) <sub>4</sub> COOH) form. When initial feed water pH was adjusted to 7.0 or 10.0, PFHxA could
214	dissociate almost completely. On the other hand, the surface charge of NF membrane could also be
215	changed at different pH. Thus, the interaction between PFHxA and membrane surface charge may
216	change at different pH.

For NF 270, its selective layer was made from polyamide, which possesses dissociable carboxylic and amine groups. According to the data reported by Mänttäri et al., 2006, this membrane has an isoelectric point (IEP) between pH 5.0 and pH 6.0. At pH below 5.0, the amine functional groups on membrane surface would protonate ( $\equiv NH_2 \rightarrow \equiv NH_3^+$ ) and resulted in positive charge. At pH above

6.0, the carboxyl groups on membrane surface would deprotonate ( $\equiv COOH \rightarrow \equiv COO^{-}$ ) and resulted in negative charge.

224

225	When the initial pH of feed water was at 3.3, membrane surface of NF 270 was positively charged
226	as a result of protonation of amine groups. The electrical repulsion did not happen between PFHxA
227	ion and membrane surface charge. On the contrary, PFHxA ion might even be adsorbed onto the
228	membrane surface due to the electrostatic attraction between the hydrophilic charged head group of
229	PFHxA (R-COO <sup>-</sup> ) and membrane surface charge. Once the charged head group of PFHxA ion was
230	adsorbed onto the membrane surface of NF 270, the hydrophobic tail of PFHxA ion would be
231	oriented towards the aqueous phase, which made the membrane surface became more neutral. Thus,
232	PFHxA ion can pass through membrane easily and resulted in a relatively low rejection rate of
233	PFHxA. However, when the initial pH of feed water was increased to 7.0 and 10.0, the membrane
234	surface of NF 270 became negatively charged as a result of deprotonation of carboxyl groups.
235	Electrical repulsion between PFHxA ion and membrane surface took place and lead to high rejection
236	rate to PFHxA.

237

For NTR-7450, its selective layer was made from sulfonated polyethersulfone and the functional group on selective layer is sulfonate group ( $-SO_3H$ ). This membrane can only be negatively charged, which was proved by its zeta-potential measured by Mänttäri et al., 2006. Thus, electrical repulsion

241	between PFHxA ion and membrane surface charge took place at all pH conditions tested. When the
242	initial feed water pH was 3.3, PFHxA in feed water could be in ionic ( $CF_3(CF_2)_4COO^-$ ) and
243	non-ionic (CF <sub>3</sub> (CF <sub>2</sub> ) <sub>4</sub> COOH) form. Electrical repulsion alone might not be able to reject the PFHxA
244	effectively as non-ionized PFHxA cannot be repelled by membrane surface charge. But the rejection
245	rate of PFHxA was still higher than 96%. This could be explained by the shrinkage of membrane. At
246	acidic conditions, the dissociation of sulfonate group was depressed and resulted in less electrical
247	repulsion inside membrane "pores". Thus, the shrinkage of membrane happened. This made the
248	"pore size" of NTR-7450 became smaller and the rejection rate of PFHxA was enhanced. When the
249	initial feed water pH was 7.0 and 10.0, the sulfonate groups on selective layer were dissociated and
250	membrane surface became negatively charged. Meanwhile, PFHxA was also ionized at neutral and
251	alkaline conditions. Thus, PFHxA could be rejected via electrical repulsion.
252	
253	Another interesting finding shows that the change of feed water pH throughout the filtration tests
254	was quite different at different initial feed water pH. When only PFHxA was added (initial pH: 3.3)
255	into the feed water, the pH of feed water kept decreasing throughout the filtration test due to the
256	concentration of PFHxA. However, when the initial pH of feed water was adjusted to pH 7.0, the pH
257	of feed water throughout the filtration was changed slightly. After adjusting the initial pH of feed
258	water to 10.0, the pH of feed water decreased significantly first and then kept between 7.0 and 8.5.
259	Such kind of change may be caused by two factors: co-ion competition effect and the dissociation of

260	PFHxA. When NaOH was added into the feed water, there are mainly three kinds of ions presented
261	in the feed water, including Na <sup>+</sup> , OH <sup>-</sup> , CF <sub>3</sub> (CF <sub>2</sub> ) <sub>4</sub> COO <sup>-</sup> . OH <sup>-</sup> is smaller than CF <sub>3</sub> (CF <sub>2</sub> ) <sub>4</sub> COO <sup>-</sup> and
262	has higher mobility. Thus, it can penetrate through the membrane much more easily than
263	$CF_3(CF_2)_4COO^-$ and lead to a decrease of feed water pH. Additionally, the percentage of PFHxA
264	dissociated also may change at different concentrations. Therefore, the pH of feed water changed
265	significantly throughout the filtration.

### 267 **3.2 Membrane flux of two NF membranes**

The membrane flux of both NF 270 and NTR-7450 throughout the filtration test under different 268269filtration conditions were shown in Fig.3. It can be found that, under three kinds of testing 270conditions, flux of NF 270 was more stable than that of NTR-7450. When the initial pH is 3.3, the 271flux of NF 270 has a slightly decrease at the end of filtration, which might be caused by the 272concentration polarization or membrane fouling. After adjusting the initial pH of feed water to 7.0 273and 10.0, the flux did not change so much throughout the filtration. The membrane flux was even 274increased after filtrating for some time when the initial pH of feed water was 10.0. However, for 275NTR-7450, the membrane flux keeps decreasing throughout the filtration test when the initial pH of 276feed water was 3.3. After increasing the initial pH of feed water to 7.0, the flux became stable 277during the whole filtration. When the initial pH of feed water became 10.0, the flux even increased

along with the filtration. Based on above information, it can be concluded that increasing the pH to





the conformation of polymer chains at different pH was quite different. Thus, membrane flux of NF
membrane at different pH was also quite different.

293

At all pH condition tested (Initial pH 3.3, 7.0 and 10.0), NF 270 was charged and the conformation of polymer chains was expanded as a result of electrostatic repulsion between polymer chains. This is the reason why the flux of NF 270 was not changed by the change of feed water initial pH.

297

298However, for NTR-7450, its selective layer was made from sulfonated polyethersulfone, which possesses dissociable sulfonate groups. According to the data reported by Mänttäri et al., 2006, it 299300 showed negative surface charge and seems did not have IEP. Increasing the pH of feed water might promote the dissociation of sulfonate groups on membrane surface and the membrane "pores" can 301 302be more open, leading to flux increase finally. In contrast, decreasing the pH of feed water may depress the dissociation of sulfonate group and leading to the shrinkage of polymer chains, which 303 could lead to the flux decline finally. Additionally, Mänttäri et al., 2006 also investigate the 304 305hodrophilicity of NTR-7450 at different pH. They found that this membrane became very hydrophobic at acidic conditions. Thus, the low membrane flux of NTR-7450 at initial pH 3.3 might 306 be caused by the shrinkage of membrane selective layer and the intrinsic hydrophobicity of 307 308 membrane surface. Once the initial pH of feed water was increased to 7.0 and 10.0, the sulfonate 309 groups became ionized and the membrane surface became more hydrophilic, which could reduce the

resistance for water to pass through the membrane. Thus, the membrane flux of NTR-7450 increasedand kept constant throughout the filtration.

312

### 313 **4.** Conclusions

In conclusion, PFHxA in aqueous solution was concentrated successfully by two commercially 314315available NF membranes. The impact of initial feed water pH on the rejection performances of NF 316 membranes was investigated. The flux of polyamide NF membrane (NF 270) kept stable when 317changing the initial feed water pH. But its rejection rate to PFHxA was greatly enhanced once the initial feed water pH was increased from 3.3 to 7.0 or 10.0. However, opposite results were obtained 318319 with sulfonated polethersulfone NF membrane (NTR-7450). It showed very high stability on rejection rate of PFHxA (>94.0%) while changing the initial feed water pH. But its membrane flux 320321was sensitive to the change of initial feed water pH. When the initial feed water pH was increased 322from 3.3 to 7.0 and 10.0, the flux was increased significantly. The mechanism behind these 323phenomena was also discussed. The findings reported in this study could provide some useful guides 324for process design in treating industrial wastewater containing PFHxA.

325

326 **References** 

Boussu, K., Van der Bruggen, B., Volodin, A., Snauwaert, J., Van Haesendonck, C., Vandecasteele, C.,
 2005. Roughness and hydrophobicity studies of nanofiltration membranes using different modes

329	of AFM. J. Colloid Interface Sci. 286, 632-638. doi:10.1016/j.jcis.2005.01.095
330	Braeken, L., Bettens, B., Boussu, K., Van Der Meeren, P., Cocquyt, J., Vermant, J., Van der Bruggen,
331	B., 2006. Transport mechanisms of dissolved organic compounds in aqueous solution during
332	nanofiltration. J. Membr. Sci. 279, 311-319. doi.org/10.1016/j.memsci.2005.12.024
333	ECHA, European Chemical Agency, Candidate list of substances of very high concern for
334	authorisation. https://echa.europa.eu/web/guest/candidate-list-table [Accessed on February 7,
335	2017]
336	Fujii, S., Polprasert, C., Tanaka, S., Lien, N.P.H., Qiu, Y., 2007. New POPs in the water environment:
337	distribution, bioaccumulation and treatment of perfluorinated compounds-a review paper. J.
338	Water Supply Res. T. 56, 313-326. doi: 10.2166/aqua.2007.005Giesy, J.P., Kannan, K., 2001.
339	Global distribution of perfluorooctane sulfonate in wildlife. Environ. Sci. Technol. 35, 1339-1342.
340	doi: 10.1021/es001834k
341	Guo, R., Megson, D., Myers, A.L., Helm, P.A., Marvin, C., Crozier, P., Mabury, S., Bhavsar, S.P.,
342	Tomy, G., Simcik, M., 2016. Application of a comprehensive extraction technique for the
343	determination of poly-and perfluoroalkyl substances (PFASs) in Great Lakes Region sediments.
344	Chemosphere 164, 535-546. doi.org/10.1016/j.chemosphere.2016.08.126
345	Hang, X., Chen, X., Luo, J., Cao, W., Wan, Y., 2015. Removal and recovery of perfluorooctanoate

- 346 from wastewater by nanofiltration. Sep. Purif. Technol. 145, 120-129.
- 347 doi.org/10.1016/j.seppur.2015.03.013

348	Herzke, D., Olsson, E., Posner, S., 2012. Perfluoroalkyl and polyfluoroalkyl substances (PFASs) in
349	consumer products in Norway–A pilot study. Chemosphere 88, 980-987.
350	doi.org/10.1016/j.chemosphere.2012.03.035
351	Kannan, K., Corsolini, S., Falandysz, J., Fillmann, G., Kumar, K.S., Loganathan, B.G., Mohd, M.A.,
352	Olivero, J., Wouwe, N.V., Yang, J.H., 2004. Perfluorooctanesulfonate and related fluorochemicals
353	in human blood from several countries. Environ. Sci. Technol. 38, 4489-4495.
354	doi: 10.1021/es0493446
355	Karnwadee, W., 2015. Development Of Effective Removal Procedures Of Perfluorohexanoic Acid
356	(PFHxA) From Industrial Wastewater By Adsorption And Regeneration (PhD thesis). Graduate
357	School of Global Environmental Studies, Kyoto University.
358	Kunacheva, C., Tanaka, S., Fujii, S., Boontanon, S.K., Musirat, C., Wongwattana, T., Shivakoti, B.R.,
359	2011. Mass flows of perfluorinated compounds (PFCs) in central wastewater treatment plants of
360	industrial zones in Thailand. Chemosphere 83, 737-744.
361	doi.org/10.1016/j.chemosphere.2011.02.059
362	Lu, Z., Song, L., Zhao, Z., Ma, Y., Wang, J., Yang, H., Ma, H., Cai, M., Codling, G., Ebinghaus, R.,
363	2015. Occurrence and trends in concentrations of perfluoroalkyl substances (PFASs) in surface
364	waters of eastern China. Chemosphere 119, 820-827. doi.org/10.1016/j.chemosphere.2014.08.045
365	Mänttäri, M., Pihlajamäki, A., Nyström, M., 2006. Effect of pH on hydrophilicity and charge and their
366	effect on the filtration efficiency of NF membranes at different pH. J. Membr. Sci. 280, 311-320.

### 367 doi.org/10.1016/j.memsci.2006.01.034

- 368 Prevedouros, K., Cousins, I.T., Buck, R.C., Korzeniowski, S.H., 2006. Sources, fate and transport of
- 369 perfluorocarboxylates. Environ. Sci. Technol. 40, 32-44. doi: 10.1021/es0512475
- 370 Rahman, M.F., Peldszus, S., Anderson, W.B., 2014. Behaviour and fate of perfluoroalkyl and
- 371 polyfluoroalkyl substances (PFASs) in drinking water treatment: a review. Water Res. 50,
  372 318-340. doi.org/10.1016/j.watres.2013.10.045
- 373 Steinle-Darling, E., Reinhard, M., 2008. Nanofiltration for trace organic contaminant removal:
- 374 structure, solution, and membrane fouling effects on the rejection of perfluorochemicals. Environ.
- 375 Sci. Technol. 42, 5292-5297. doi: 10.1021/es703207s
- 376 Svihlikova, V., Lankova, D., Poustka, J., Tomaniova, M., Hajslova, J., Pulkrabova, J., 2015.
- Perfluoroalkyl substances (PFASs) and other halogenated compounds in fish from the upper Labe
- 378 River basin. Chemosphere 129, 170-178. doi.org/10.1016/j.chemosphere.2014.09.096
- Takemine, S., Matsumura, C., Yamamoto, K., Suzuki, M., Tsurukawa, M., Imaishi, H., Nakano, T.,
- 380 Kondo, A., 2014. Discharge of perfluorinated compounds from rivers and their influence on the
- 381 coastal seas of Hyogo prefecture, Japan. Environ. Pollut. 184, 397-404.
- 382 doi.org/10.1016/j.envpol.2013.09.016
- 383 Tang, C.Y., Fu, Q.S., Criddle, C.S., Leckie, J.O., 2007. Effect of flux (transmembrane pressure) and
- membrane properties on fouling and rejection of reverse osmosis and nanofiltration membranes
- treating perfluorooctane sulfonate containing wastewater. Environ. Sci. Technol. 41, 2008-2014.

387	Tang, C.Y., Fu, Q.S., Robertson, A., Criddle, C.S., Leckie, J.O., 2006. Use of reverse osmosis
388	membranes to remove perfluorooctane sulfonate (PFOS) from semiconductor wastewater.
389	Environ. Sci. Technol. 40, 7343-7349. doi: 10.1021/es060831q
390	Tang, C.Y., Kwon, YN., Leckie, J.O., 2009. Effect of membrane chemistry and coating layer on
391	physiochemical properties of thin film composite polyamide RO and NF membranes: II.
392	Membrane physiochemical properties and their dependence on polyamide and coating layers.
393	Desalination 242, 168-182. doi.org/10.1016/j.desal.2008.04.004
394	Membrane SolTech, http://www.membrane-soltech.com/pg178.html [Accessed in February 7th, 2017]
395	US EPA, United States Environmental Protection Agency, 2006. PFOA Stewardship Program.
396	hhttp://www.epa.gov/assessing-and-managing-chemicals-under-tsca/and-polyfluoroalkyl-substan
397	ces-pfapf-under-tsca#tab-3. [Accessed in February 7th, 2017]
398	Wang, Z., Cousins, I.T., Scheringer, M., Hungerbühler, K., 2013. Fluorinated alternatives to long-chain
399	perfluoroalkyl carboxylic acids (PFCAs), perfluoroalkane sulfonic acids (PFSAs) and their
400	potential precursors. Environ. Int. 60, 242-248. doi.org/10.1016/j.envint.2013.08.021
401	Wang, Z., DeWitt, J.C., Higgins, C.P., Cousins, I.T., 2017. A Never-Ending Story of Per-and
402	Polyfluoroalkyl Substances (PFASs)? . Environ. Sci. Technol. 51, 2508-2518. doi:
403	10.1021/acs.est.6b04806

404 Yeung, L.W., So, M., Jiang, G., Taniyasu, S., Yamashita, N., Song, M., Wu, Y., Li, J., Giesy, J., Guruge,

406

K., 2006. Perfluorooctanesulfonate and related fluorochemicals in human blood samples from China. Environ. Sci. Technol. 40, 715-720. doi: 10.1021/es052067y

- 407 Yu, N., Shi, W., Zhang, B., Su, G., Feng, J., Zhang, X., Wei, S., Yu, H., 2013. Occurrence of
- 408 perfluoroalkyl acids including perfluorooctane sulfonate isomers in Huai River Basin and Taihu
- 409 Lake in Jiangsu Province, China. Environ. Sci. Technol. 47, 710-717. doi: 10.1021/es3037803
- 410 Zhao, C., Tang, C.Y., Li, P., Adrian, P., Hu, G., 2016. Perfluorooctane sulfonate removal by
- 411 nanofiltration membrane—the effect and interaction of magnesium ion/humic acid. J. Membr. Sci.
- 412 503, 31-41. doi.org/10.1016/j.memsci.2015.12.049
- 413 Zhao, C., Zhang, J., He, G., Wang, T., Hou, D., Luan, Z., 2013. Perfluorooctane sulfonate removal by
- 414 nanofiltration membrane the role of calcium ions. Chem. Eng. J. 233, 224-232.
- 415 doi.org/10.1016/j.cej.2013.08.027