

1 (1) Title

2 **Nitrogen resorption efficiency of 13 tree species of a cool temperate deciduous**
3 **forest in Central Japan**

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5 (2) Authors

6 Ryunosuke Tateno^{1*} and Hiroshi Takeda²

7

8 1 *Field Science Research and Education Center, Kyoto University, Kyoto, 6068502,*

9 *Japan*

10 2 *Faculty of Science and Engineering, Doshisha University, Kyotanabe 6100321,*

11 *Japan*

12

13 (3) *Corresponding author

14 Ryunosuke Tateno

15 Field Science Research and Education Center, Kyoto University, Kyoto, 6068502,

16 Japane-mail: rtateno@kais.kyoto-u.ac.jp

17 Fax: +81-75-753-6441

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32 **Abstract**

33 The nitrogen (N) concentration of green and senescent leaves, and the N resorption
34 efficiency of 13 dominant species from three distribution patterns along a slope (Ridge,
35 Valley, and Uniform types) and two types of leaf emergence pattern (determinate and
36 indeterminate flush types) were examined in a cool temperate natural forest in central
37 Japan. Ridge species tended to have lower N concentrations of green and senescent
38 leaves and higher N resorption efficiency than those of Valley species, and those of
39 Uniform species were intermediate to other types with some exceptions. Furthermore,
40 indeterminate flush species tended to have higher N concentrations in green and
41 senescent leaves and lower N resorption efficiency than those of the determinate type.
42 The N concentration of senescent leaves, which is an index of leaf-level N use
43 efficiency (NUE), was significantly correlated with N concentrations in green leaves
44 and with N resorption efficiency. The concentration in green leaves was not correlated
45 with N resorption efficiency, suggesting that interspecific variation in the N
46 concentration was not a major determinate of the N resorption efficiency at this study
47 site. Rather N resorption efficiency was positively correlated with the start date of leaf
48 fall, suggesting that early leaf fall species, which tended to have intermediate leaf flush
49 types (with some exceptions), were not proficient in N resorption. Thus, the patterns in
50 NUE and N resorption efficiencies were affected not only by the distribution pattern
51 along the slope but also by leaf phenology in this cool temperate deciduous forest.

52

53 Key words: Beech forest, Leaf fall phenology, Nitrogen use efficiency, Retranslocation,
54 Topography

55 **Introduction**

56

57 Nitrogen (N) availability limits plant growth in many ecosystems (Vitousek and
58 Howarth 1991), so efficient N utilization is important to plant competition (Chapin
59 1980; Vitousek 1982, 1984; Berendse and Aerts 1987; Eckstein et al. 1999; Aerts and
60 Chapin 2000). Resorption of N from senescing leaves reduces N loss, which may be an
61 important mechanism for N conservation (Chapin 1980; Chabot and Hicks 1982; Aerts
62 1996; Eckstein et al. 1999; Niinemets and Tamm 2005; Millard and Grelet 2010; Sun
63 et al. 2016). N resorption efficiency, which can be calculated from the N
64 concentrations in green and senescent leaves, is higher at sites with low-N availability
65 than those with high-N availability (Small 1972; Flanagan and van Cleve 1983; Enoki
66 and Kawaguchi 1999; but see Lajtha 1987; Aerts 1996), but this is still controversial. N
67 storage in plants and its remobilization to plant organs for growth, maintenance and
68 reproduction could be important determinates of the N resorption efficiency (Lajtha
69 1987; Millard and Grelet 2010). N resorption efficiency increases with stand age due to
70 increasing N requirements (Sun et al. 2016). Furthermore, Estiarte and Peñuelas (2015)
71 recently suggested that warming and drought due to climate change affect the nutrient
72 resorption efficiency. Understanding of plant N utilization, including the N use
73 efficiency (NUE) and N resorption efficiencies, is important for forest management
74 under a changing climate and increasing N deposition (Millard and Grelet 2010;
75 Estiarte and Peñuelas 2015).

76 NUE can be defined as dry matter productivity divided by plant N uptake
77 (Hirose 1971, 2011; Vitousek 1982). Ideally, dry matter productivity should include
78 plant dry mass (including litterfall) and respiratory consumption in non-photosynthetic

79 tissues, including stems, roots, and reproductive organs (e.g., Hirose 2011). Many
80 studies have assumed a steady state and have simply used dry matter lost (i.e., leaf fall
81 or litterfall weight) as an index of dry matter productivity (e.g., Vitousek 1982; Hirose
82 2011), given the complexity of estimating NUE and its components in natural
83 environments, especial for canopy tree species. Likewise, plant N uptake from soil is
84 equivalent to N loss, such as N content lost to leaf fall or litterfall (leaf fall or litterfall
85 weight \times N concentration of senescent leaves or organs). Thus, the inverse N
86 concentration of leaf fall or litterfall is an index of NUE (e.g., Vitousek 1982).

87 Furthermore, NUE can be divided into mean residence time (MRT) and
88 nitrogen productivity (NP) (Berendse and Aerts 1987; Eckstein et al. 1999; Hirose
89 2011). MRT can be defined as the mean standing plant N during the growth period
90 (N_{mean}) divided by the plant N uptake, and NP can be defined as the dry matter
91 productivity divided by N_{mean} (Berendse and Aerts 1987; Eckstein et al. 1999; Hirose
92 2011). The N_{mean} , which is the numerator and denominator of MRT and NP,
93 respectively, can be the product of biomass, growth period, and the N concentration of
94 plant organs. If we assume that biomass and organ concentration do not change over
95 the growth period and that the N concentration of organs equals the leaf N
96 concentration, MRT can be explained simply as a function of the growth period and N
97 resorption efficiency, and NP can be simply explained as a function of the N
98 concentration of the green leaves and the growth period. Thus, an increased growth
99 period and N resorption efficiency increase MRT; similarly, an increased growth period
100 and N concentration in green leaves decrease NP.

101 Leaf longevity, which varies widely from several months to several years
102 (Kikuzawa 1983, 1984; Reich et al. 1992; Osada et al. 2001), is also related to the

103 NUE (Eckstein et al. 1999) and soil N availability (Pornon et al. 2011). Major
104 acclimations to soils with low-N availability include increased leaf longevity and lower
105 leaf N concentration (Aerts and Chapin 2000). Leafing duration, which refers to the
106 period of leaf holding on the crown and differs from leaf longevity due to turnover of
107 leaves and abrupt leaf fall during the growing season, is also an important parameter
108 for NUE and its components in deciduous forest ecosystems. In temperate deciduous
109 tree species, which flush their leaves in the spring and shed them in the fall, leaf
110 longevity is less than 1 year (Kikuzawa 1983, 1984; Aikawa et al. 2002). The reported
111 leafing duration of cool temperate deciduous species in central Japan ranged from 150
112 to 200 days and was determined mostly by the leaf fall dates rather than by the leaf
113 flushing dates (Aikawa et al. 2002). In cool temperate forests, abrupt early leaf fall
114 resulted in incomplete N resorption and low NUE (Yasumura et al. 2002). Thus, leaf
115 phenology is also important as an N conservation mechanism. More recently, plant
116 phenology has been considered a crucial factor affecting vegetation feedback by
117 influencing the seasonality of albedo, surface roughness length, canopy conductance
118 and canopy water and energy fluxes in future climate changes (Richardson et al. 2013).

119 Trees of the temperate region exhibit two types of leaf emergence:
120 determinate growth (producing the bulk of their foliage in a short spring flush) and
121 indeterminate growth (producing new leaves in spring but also late into the growing
122 season; Kikuzawa 1983). Species of the indeterminate growth type tend to drop their
123 leaves earlier than do those of the determinate growth type (Aikawa et al. 2002; Tateno
124 et al. 2005). Leaf emergence patterns of tree species are correlated with shade tolerance,
125 i.e., pioneer species have indeterminate growth and climax species have determinate
126 growth (Kikuzawa 1983, 1984; Hirayama and Sakimoto 1999). In terms of differences

127 in growth strategy, plants may change their N utilization patterns; however, few studies
128 have investigated the effects of leaf phenology on NUE and N resorption efficiency
129 (but see del Arco et al. 1991; Yasumura et al. 2002; Niinemets and Tamm 2005;
130 Inagaki et al. 2008).

131 This study investigated the NUE and N resorption efficiencies of 13 dominant
132 species (seven with determinate growth and six with indeterminate growth) in a cool
133 temperate natural forest in central Japan growing on a wide range of soils. The focal
134 species could be categorized into three types (Valley, Ridge, and Uniform species)
135 according to their distribution along the topographic gradient (Tateno and Takeda
136 2003). This study investigated how tree distribution patterns, leaf emergence type, and
137 leaf fall phenology affect plant NUE and N proficiency in a cool temperate forest.

138

139 **Materials and Methods**

140

141 **Study site**

142

143 This study was conducted in a cool temperate broadleaf deciduous forest at Ashiu
144 Forest Research Station, Kyoto University (35°18'N, 135°43'E), Kyoto Prefecture,
145 Japan. The mean annual temperature and precipitation are 10–11°C and 2895 mm,
146 respectively, at a weather station (640 m elevation) about 1 km from the study site.

147 We established a 30 × 200 m (0.6-ha) transect plot extending from a valley
148 bottom to the top of a northwest-facing slope. Diameter at breast height (DBH) was
149 measured for all trees larger than 5 cm DBH. The details of the distributions of the
150 species studied are reported elsewhere (Tateno and Takeda 2003). From this

151 preliminary vegetation census, we selected 13 dominant tree species (Table 1), which
152 accounted for more than 81% of the total basal area and 66% of the total number of
153 trees in the study plot. Details of the leaf phenology of the species studied are reported
154 elsewhere (Aikawa et al.2002; Tateno et al. 2005).

155 The study species were categorized into two groups based on their type of leaf
156 emergence: those with determinate (seven species) and indeterminate (six species)
157 growth (Kikuzawa 1983, 1984; Hirayama and Sakimoto 1999; S. Morozumi personal
158 observation for *Pterostyrax hispida* and *Lyonia ovalifolia*). Topography is one of the
159 most important environmental gradients in terms of soil moisture and nutrients in cool
160 temperate ecosystems and the distribution patterns of tree species often show a
161 topographic tendency (Tateno and Takeda 2003). The distribution pattern of tree
162 species reflects the soil habitat preference of a species (e.g., Chapin 1980). The sample
163 species were categorized into the three types determined by Tateno and Takeda (2003)
164 and are summarized in Table 1. Only one study species, *Aesculus turbinata*, was not
165 categorized by Tateno and Takeda (2003) because of its low sample size; however,
166 *Aesculus turbinata* is a riparian species in cool temperate forests (Kaneko et al. 1999)
167 and thus was categorized as a Valley species.

168

169 Leaf sampling

170

171 To avoid the effects of light conditions on plant N utilization traits, we selected only
172 canopy trees for this study, because the N resorption efficiency of canopy trees is
173 different from that of subcanopy trees (Tateno and Kawaguchi 2002). We conducted
174 leaf sampling in early September 1998, the end of the growing season, for measuring N

175 concentration of green leaves. We sampled five trees for each species and collected one
176 sunlit branch from the crown surface of each tree to minimize the variation in the N
177 concentration in the crown within and outside the study plot. Uniform species were
178 collected from a wide range of slope positions to avoid bias in soil conditions. All
179 sampled branches included more than 10 current-year shoots, except for species with
180 compound leaves (*Acanthopanax sciadophylloides* and *Aesculus turbinata*). For these
181 species, we collected more than 10 compound leaves per tree from three or four shoots.
182 A total of 10 shoots or 10 compound leaves were randomly selected from the sampled
183 branch for each tree. Details of leaf sampling are reported elsewhere (Tateno et al.
184 2003). To collect senescent leaves, we gently shook the branches of sampled species
185 beneath the crown to obtain newly shed leaves; we repeated this several times
186 throughout the leaf-fall period of each species (from September to November).

187 All leaf samples (both green and senescent) were oven dried and ground into a
188 powder, and then the total carbon (C) and N content of each species was determined
189 using a C/N analyzer (NC-900, Shimadzu, Kyoto, Japan). N resorption efficiency was
190 calculated using the following equation:

$$191 \quad \text{N resorption efficiency (\%)} = (N_g - N_s) / N_g \times 100,$$

192 where N_g is the N concentration of green leaves and N_s is the N concentration of
193 senescent leaves.

194

195 Statistical analysis

196

197 We used one-way ANOVA to test the effects of species on the N concentrations of
198 green and senescent leaves and on N resorption efficiency. For multiple comparisons,

199 we used Tukey's HSD test to evaluate the N concentrations of green and senescent
200 leaves and the Games–Howell test to evaluate N resorption efficiency depending on
201 the results of Levene's test. We also used one-way ANOVA with Games-Howell test to
202 test the effects of distribution pattern and *t*-test to test the effects of leaf emergence
203 pattern on the N concentrations of green and senescent leaves and on N resorption
204 efficiency. We used linear regression to analyze the relationship between leaf fall
205 timing and plant N parameters, such as the N concentrations of green and senescent
206 leaves and N resorption efficiency, and to analyze the relationships among plant N
207 parameters. Percentage data were arcsine transformed before the statistical analysis.

208 We used the start date of leaf fall (the day on which 10% of leaves were shed)
209 as an index of leaf fall phenology, following Tateno et al. (2005), as summarized in
210 Table 1. For some species that occurred at both upper and lower slope positions, we
211 calculated the average start date across trees at both sites. We should note that leaf
212 sampling was conducted in 1998, but leaf phenology was estimated in 2001. Leaf
213 phenology tends to be similar from year to year, regardless of the species (Vitasse et al.
214 2009), so we assumed that sampling year effects were negligible. All statistical
215 analyses were performed using IBM SPSS Statistics (IBM SPSS 22.0, IBM, Armonk,
216 NY, USA).

217

218 **Results**

219

220 The N concentration of green leaves ranged from $17.1 \pm 1.5 \text{ mg g}^{-1}$ (mean \pm SD) in
221 *Clethra barbinervis* to $30.7 \pm 1.3 \text{ mg g}^{-1}$ in *Pterostyrax hispida* (Table 1). Especially in
222 Valley species of the indeterminate leaf flush type, the N concentration in green leaves

223 was higher than that in species of the other types. Overall mean of Ridge, Uniform and
224 Valley species were 17.1 ± 1.3 , 21.3 ± 3.0 and 26.3 ± 4.3 mg g⁻¹, respectively, which
225 were significantly different each other (one-way ANOVA with Games-Howell test, $p <$
226 0.05). Overall mean of indeterminate and determinate leaf flush type were 25.2 ± 5.2
227 and 20.6 ± 2.7 mg g⁻¹, respectively, and there was a significant difference between leaf
228 flush type (t-test, $p < 0.05$).

229 The N concentration of senescent leaves ranged from 10.1 ± 2.0 mg g⁻¹ in
230 *Sorbus alnifolia* to 28.9 ± 0.8 mg g⁻¹ in *Pterostyrax hispida* (Table 1). Especially in
231 Valley species of the indeterminate leaf flush type, the N concentration in senescent
232 leaves was higher than that in species of the other types, similar to the case with green
233 leaves. Overall mean of Ridge, Uniform and Valley species were 12.3 ± 2.5 , 14.1 ± 3.3
234 and 22.0 ± 6.3 mg g⁻¹, respectively, and Valley species had lower N concentrations in
235 senescent leaves than Valley species and Uniform species (one-way ANOVA with
236 Games-Howell test, $p < 0.05$). However, these pattern included some exceptions, for
237 example, the N concentration in senescent leaves of *Aesculus turbinata*, which is
238 categorized as a Valley species, was low. Overall mean of indeterminate and
239 determinate leaf flush type were 20.8 ± 6.3 and 13.5 ± 3.6 mg g⁻¹, respectively, and
240 there was a significant difference between leaf flush type (t-test, $p < 0.05$).

241 The N resorption efficiency ranged from 6.0 ± 2.7 % in *Pterostyrax hispida* to
242 49.1 ± 12.0 % in *Sorbus alnifolia* (Table 1). Overall mean of Ridge, Uniform and Valley
243 species were 30.8 ± 10.7 , 35.3 ± 9.6 and 21.5 ± 13.5 %, respectively. Significant
244 differences were only found between Valley and Uniform species (one-way ANOVA
245 with Games-Howell test, $p < 0.05$). Thus the N resorption efficiency did not show
246 consistent patterns with habitat preference. Overall mean of indeterminate and

247 determinate leaf flush type were 23.0 ± 12.1 and $34.7 \pm 11.2\%$, respectively, and there
248 was a significant difference between leaf flush type (t-test, $p < 0.05$). However, these
249 pattern included some exceptions, for example, the N resorption efficiency of *Acer*
250 *mono* and *Fagus crenata* were low among determinate species and those of
251 *Acanthopanax Sciadophylloides* were high among indeterminate species.

252 The N concentration and N resorption efficiency were negatively correlated in
253 senescent leaves (Fig. 1b, $r^2 = 0.762$, $p < 0.001$) but showed no linear correlation in
254 green leaves (Fig. 1a, $p > 0.05$). In addition, there was a significant linear correlation
255 between N concentrations in senescent and green leaves (Fig. 1c, $r^2 = 0.752$, $p < 0.001$),
256 and between N resorption efficiency and the start date of leaf fall (Fig. 2a, $r^2 = 0.506$,
257 $p < 0.001$), but not between N concentrations of green and senescent leaves and the start
258 date of leaf fall (Fig. 2b, c, $p > 0.05$).

259

260 **Discussion**

261

262 Leaf emergence patterns are often correlated with shade tolerance; the leaf N
263 concentration tended to be higher in indeterminate species than in determinate species
264 in this study. Pioneer species tend to have higher leaf N concentrations than do
265 shade-tolerant climax species (Ellsworth and Reich 1996). Differences in leaf N
266 concentration between leaf emergence type could reflect their shade tolerance, because
267 shade-intolerant pioneer species show indeterminate growth and shade-tolerant climax
268 species show determinate growth (Kikuzawa 1983, 1984; Hirayama and Sakimoto
269 1999). However, the green leaf N concentration was not related to the N resorption
270 efficiency. Rather, leaf fall occurred earlier in indeterminate species than in

271 determinate species, which could explain the low N resorption efficiencies found in
272 indeterminate species. Niinemets and Tamm (2005) suggested that the inherent
273 potential of N resorption in tree species depends not on the initial N content, but on the
274 greater N losses associated with earlier leaf fall. Thus, indeterminate species did not
275 re-translocate N efficiently before leaf shedding.

276 Habitat preference with respect to topography is also a potential factor
277 affecting the NUE and N resorption efficiencies of tree species, as these traits are
278 important for acclimation to infertile soil environments (e.g., Aerts and Chapin 2000).
279 In this study, N concentrations in green leaves showed clear patterns with topographic
280 preference. However, the N resorption efficiency did not show consistent patterns with
281 habitat preference. At this study site, leaf fall occurred slightly earlier on the upper slope
282 than on the lower slope (Tateno et al. 2005), suggesting that incomplete N resorption
283 tends to occur more on upper slopes, where soil N availability is lower (Tateno and
284 Takeda 2003), than on lower slopes. However, at least at the stand level, N resorption
285 efficiencies were negatively correlated with N availability in the soil (Tateno and
286 Takeda 2010), although previous studies have demonstrated that N resorption efficiency
287 is not always correlated with soil N conditions (e.g., Lahja 1987). This inconsistency of
288 correlation may be due partly to the range in timing of the start of leaf fall, which
289 overlaps considerably between species on upper and lower slopes (Tateno et al. 2005).
290 Late leaf fall species may also be more dominant on upper slopes than on lower slopes.
291 At the species level, the pattern of N resorption was affected not only by soil habitat
292 preference, but also by leaf fall phenology. Thus, the timing of leaf fall appears to affect
293 the N utilization strategy of each species.

294 The N resorption efficiency among the studied species showed wide variation,

295 from 6.0% to 49.1%, which is slightly lower than the worldwide range reported by Aerts
296 (1996). In contrast to the variation in the N resorption efficiency among the species
297 evaluated, the variation in leaf longevity differed among tree species thought to be
298 relatively less in deciduous trees. For example, Kikuzawa (1983) reported that leaf
299 longevity of cool temperate tree species ranged from 90 to 170 days. The reported
300 leafing duration of cool temperate deciduous species, which is not accurately consistent
301 with leaf longevity, was 150–200 days in this site (Aikawa et al. 2002). In the present
302 study, indeterminate species tended to have shorter leafing durations than those of
303 determinate species (Aikawa et al. 2002), and the difference in leaf longevity between
304 the two leaf emergence types may be larger because indeterminate species produce and
305 shed leaves throughout the growing season (Kikuzawa 1983). Overall, the variation in
306 leaf longevity and leafing duration among species was less than twofold, which was
307 considerably less than the variation in N resorption efficiency. MRT could be increased
308 with leafing duration and N resorption efficiency, with a greater contribution of
309 changing N resorption efficiency relative to leafing duration in deciduous tree species.

310 Furthermore, the variations in leaf longevity and leafing duration in deciduous
311 trees were not accounted for in the calculation of N uptake or re-translocation on an
312 annual basis. More precisely, with even earlier or later leaf shedding, the N
313 re-translocated before winter will be utilized for next year's production. In this sense,
314 the total amount of re-translocated N is more important for temperate deciduous trees,
315 considering annual-based production and N cycling. Several studies have suggested that
316 leaf longevities are more important in explaining interspecific variation in the MRT,
317 especially in evergreen species (Escudero et al. 1992; Eckstein et al. 1999); however,
318 the differences in N resorption efficiency could be relatively more important for the

319 effective utilization of N in a cool temperate deciduous forest.

320 The inverse of the N concentration in senescent leaves has been used as an
321 index of NUE (e.g., Vitousek 1982). In the present study, both N concentration in
322 green leaves and N resorption efficiency were correlated with the N concentration in
323 senescent leaves, as previously reported (Aerts and Chapin 2000; Tateno and
324 Kawaguchi 2002). However, the N concentration of green leaves was not correlated
325 with N resorption efficiency. Killingbeck (1996) proposed that the N resorption
326 proficiency, the level to which leaf N is reduced during senescence, is for acclimation
327 to infertile soil conditions. N resorption proficiency is basically the same as leaf-level
328 NUE, which can be defined as the inverse of the N concentration in senescent leaves
329 (Vitousek 1982). If plants re-translocate N more proficiently at the time of leaf
330 abscission, the plant leaf-level NUE will increase (Killingbeck 1996). However, the
331 low N resorption proficiency of some species in this study could be related to early leaf
332 fall, probably due to abrupt water stress, wind and cold. Niinemets and Tamm (2005)
333 suggested that large N losses are associated with earlier leaf fall. Furthermore,
334 Yasumura et al. (2002) suggested that abrupt early leaf fall leads to low N resorption
335 efficiency in *Fagus crenata* in temperate deciduous forests in northern Japan.
336 Conversely, an extended delay in the senescence period due to climate change can
337 result in more proficient resorption (Estiarte and Peñuelas 2015). Further study is
338 needed to investigate the interrelationship between N utilization traits and leaf
339 phenology under climate change and increasing N deposition.

340

341 **Conclusions**

342

343 Leaf emergence patterns and habitat preference of deciduous tree species affected N
344 concentrations in green and senescent leaves and N resorption efficiencies in a cool
345 temperate natural forest in central Japan. The N concentration in senescent leaves,
346 which is an index of leaf-level NUE, was correlated significantly with N
347 concentrations in green leaves and N resorption efficiency, whereas the concentration
348 in green leaves was not correlated with N resorption efficiency. These results suggest
349 that interspecific variation in the N concentration in green leaves was not a major
350 determinate of N resorption efficiency at this study site. Rather, N resorption efficiency
351 was correlated positively with the start date of leaf fall, suggesting that species that
352 experience early leaf fall, which tend to have intermediate leaf flush types (with some
353 exceptions), were not proficient in N resorption. Early leaf fall, probably due to abrupt
354 water stress, wind, and cold, could lead to lower N resorption efficiency and NUE.
355 Thus, the patterns observed in N utilization, such as the NUE and N resorption
356 efficiencies, were affected not only by the distribution pattern along the slopes, but also
357 by leaf phenology, such as the leaf emergence type and timing of leaf fall, in this cool
358 temperate deciduous forest.

359

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361

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370

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467 **Figure legends**

468

469 **Fig. 1** Relationship between the N concentration of green leaves (Ng) and N resorption
470 efficiency (NRE) (a), N concentration of senescent leaves (Ns) and NRE (b), and Ng
471 and Ns (c) of 13 tree species. Closed symbols indicate determinate flush type and open
472 symbols indicate indeterminate flush type. Squares, circles, and triangles indicate
473 Valley, Uniform and Ridge species, respectively. Solid lines indicate the significant
474 linear regression equations. (b) The regression was $y = -1.495 x + 55.353$ ($r^2 = 0.769$, p
475 < 0.001). (c) The regression was $y = 1.197 x - 10.324$ ($r^2 = 0.752$, $p < 0.001$).

476

477 **Fig. 2** Relationship between N resorption efficiency (NRE) and start date of leaf fall
478 (a), N concentration of green leaves (Ng) and start date of leaf fall (b), and N
479 concentration of senescent leaves (Ns) and start date of leaf fall (c) of 13 tree species.
480 Closed symbols indicate determinate flush type and open symbols indicate
481 indeterminate flush type. Squares, circles, and triangles indicate Valley, Uniform and
482 Ridge species, respectively. Solid lines indicate the significant linear regression
483 equation. (a) The regression was $y = 0.441 x - 47.371$ ($r^2 = 0.506$, $p < 0.01$).

Fig. 1 Tateno and Takeda

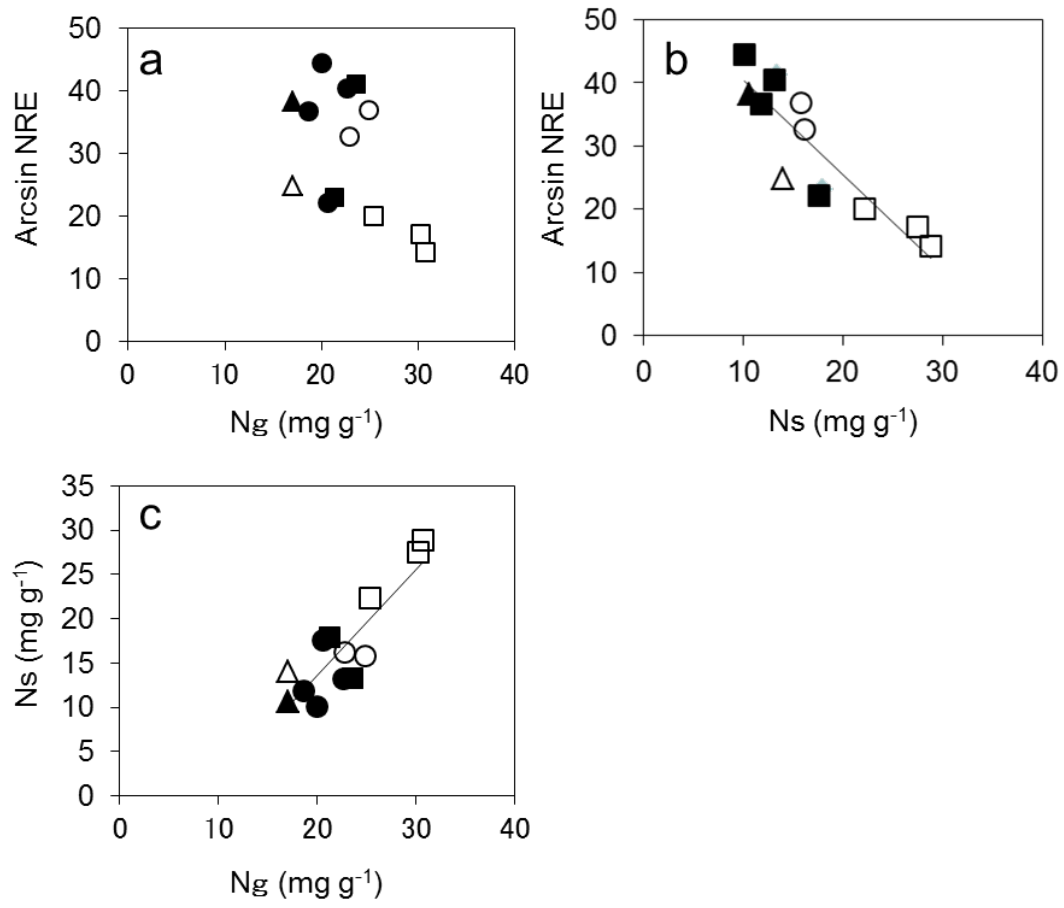


Fig. 2 Tateno and Takeda

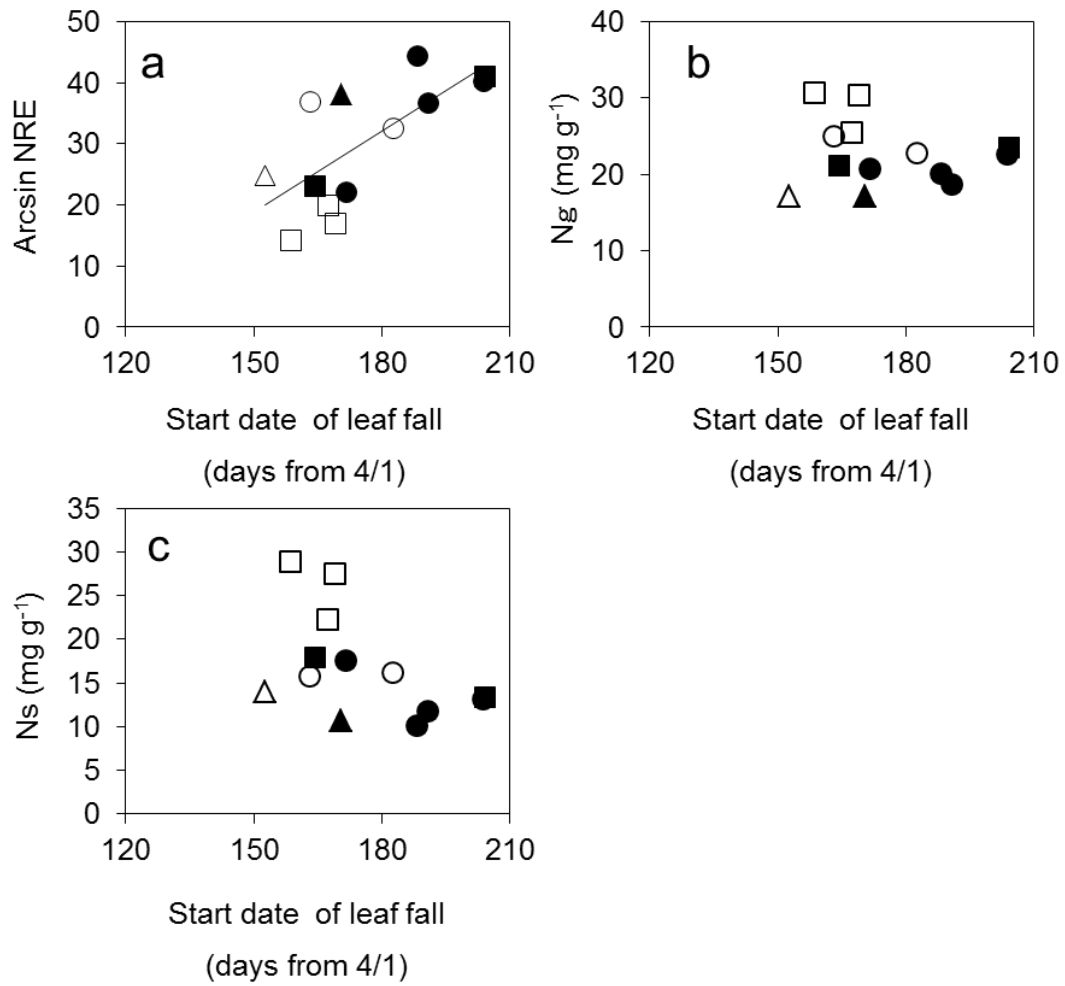


Table 1 Tateno and Takeda

Table 1 N concentration of green leaves (Ng), N concentration of senescent leaves (Ns), and N resorption efficiency (NRE) of the 13 study species. Different letter means there were no significant differences ($P < 0.05$) according to Tukey's HSD test for Ng and Ns and Games-Howell test for NRE.

Species	Distribution pattern ¹	Leaf emergence type ²	Ng (mg g ⁻¹)	Ns (mg g ⁻¹)	NRE (%)	Start date of leaf fall ³	Relative dominance (%) ⁴
<i>Aesculus turbinata</i>	Valley	Determinant	23.6 ±1.1 _{bcd}	13.4 ±0.9 _{cde}	43.2 ±3.1 _{cd}	10/22	0.1
<i>Acer mono var. marmoratum f. dissectum</i>	Valley	Determinant	21.3 ±0.9 _{bcdef}	18.0 ±3.2 _{bc}	15.4 ±13.7 _{abcd}	9/12	4.9
<i>Fagus crenata</i>	Uniform	Determinant	20.7 ±2.6 _{cdef}	17.6 ±1.6 _{bc}	14.1 ±6.1 _{ab}	9/19	23.4
<i>Quercus crispula</i>	Uniform	Determinant	22.7 ±2.0 _{bcde}	13.2 ±1.7 _{cde}	42.0 ±5.3 _{cd}	10/21	18.0
<i>Acer sieboldianum</i>	Uniform	Determinant	18.7 ±1.7 _{ef}	11.8 ±1.5 _{de}	35.7 ±13.3 _{abcd}	10/8	4.3
<i>Sorbus alnifolia</i>	Uniform	Determinant	20.0 ±1.7 _{def}	10.1 ±2.0 _e	49.1 ±12.0 _d	10/6	4.8
<i>Lyonia ovalifolia var. elliptica</i>	Ridge	Determinant	17.1 ±1.5 _f	10.6 ±1.7 _e	38.2 ±6.1 _{cd}	9/18	5.6
<i>Pterostyrax hispida</i>	Valley	Indeterminant	30.7 ±1.3 _a	28.9 ±0.8 _a	6.0 ±2.7 _a	9/6	0.3
<i>Swida controversa</i>	Valley	Indeterminant	30.3 ±3.2 _a	27.5 ±2.6 _a	8.6 ±8.3 _{abc}	9/17	1.6
<i>Carpinus tschonoskii</i>	Valley	Indeterminant	25.4 ±2.0 _b	22.3 ±2.6 _b	11.7 ±13.6 _{abcd}	9/15	3.5
<i>Carpinus laxiflora</i>	Uniform	Indeterminant	22.8 ±1.7 _{bcde}	16.2 ±2.9 _{cd}	29.2 ±9.9 _{abcd}	9/30	7.2
<i>Acanthopanax sciadophylloides</i>	Uniform	Indeterminant	24.9 ±2.4 _{bc}	15.8 ±1.5 _{cd}	36.1 ±7.8 _{bcd}	9/11	1.9
<i>Clethra barvinervis</i>	Ridge	Indeterminant	17.1 ±1.0 _f	14.0 ±1.6 _{cde}	17.6 ±11.3 _{abcd}	8/31	5.9

mean ±SD

1: Categories are according to Tateno and Takeda (2003) and Kaneko et al. (1999).

2: Categories are according to Kikuzawa (1982), Hirayama and Sakimoto (1999) and S Morozumi personal observations.

3: Day of 10% of leaf fall occurred. Data from Tateno et al. (2005).

4: Relative dominance of each tree species at 0.6 ha plot based on basal area. Data from Tateno and Takeda (2003)