

1 **Title**

2 A more selective and accurate method than indigo method for measuring  
3 dissolved ozone concentration (*p*-vinylbenzoic acid method)

4 **Author names and affiliations**

5 Kohei Kawaguchi <sup>a</sup>, Taira Hidaka <sup>a</sup> and Taku Fujiwara <sup>a</sup> \*

6 <sup>a</sup> Department of Environmental Engineering, Graduate School of Engineering, Kyoto University,  
7 Kyoto 6158540, Japan.

8 **Corresponding author**

9 Taku Fujiwara\*

10 \*fujiwara.taku.3v@kyoto-u.ac.jp

11 **Highlights**

12 Ozone reacts *p*-vinylbenzoic acid (*p*-VBA) to form *p*-formylbenzoic acid (*p*-FBA).

13 The ozone concentration can be determined from the *p*-FBA concentration.

14 *p*-VBA method is more accurate than the indigo method.

15 *p*-VBA method can be used in the presence of oxidants other than ozone.

16 *p*-VBA method does not require calibration of the quantitative reagent.

17 **ABSTRACT**

18 Dissolved ozone concentration measurement is crucial for ozone treatment. In the most used  
19 conventional indigo method, the ozone concentration is measured by the decrease in absorbance

20 due to the loss of the C–C double bond of indigo trisulfonic acid. However, measurement of ozone  
21 concentration is difficult when water contains substances that react with C–C double bonds other  
22 than ozone. To address this concern, we developed a novel breakthrough method to measure ozone  
23 concentration by measuring the *p*-formylbenzoic acid (*p*-FBA) produced after the reaction of *p*-  
24 vinylbenzoic acid and ozone. The formation of *p*-FBA was almost not caused by other substances  
25 (hypochlorous acid, hypobromous acid, permanganate ion and hydrogen peroxide), and its yield  
26 to ozone was maintained at 1 in river water, treated wastewater and seawater. In addition, the  
27 experimental error is smaller with the new method than with indigo. Furthermore, the new method  
28 does not require cumbersome calibration unlike indigo method because highly pure forms of *p*-  
29 FBA are commercially available. *p*-FBA can be separated by liquid chromatography and detected  
30 with highly sensitive ultraviolet and mass spectrometric detectors, and hence easily analyzed  
31 simultaneously with other substances. Our new method contributes to extensive ozone treatment  
32 and ozonation management.

### 33 **KEYWORDS**

34 ozonolysis, *p*-formylbenzoic acid, hypochlorous acid, hypobromous acid, permanganate ion,  
35 seawater

### 36 **1. Introduction**

37 Ozone treatment is applied in various fields, such as water and wastewater, seawater, food,  
38 medical care, sewage sludge, and pulp bleaching (Holah et al., 2016; Loeb et al., 2012; Paul and  
39 Liu, 2012; Tripathi et al., 2020). In practice and research in these fields, the measurement of  
40 dissolved ozone concentration is crucial.

41 In the indigo method (Bader and Hoigné, 1981), which is currently used for measuring dissolved  
42 ozone concentration, water containing ozone is mixed with an indigo-trisulfonic acid (indigo)  
43 solution adjusted to pH 2 using phosphate buffer. Ozone reacts with the C–C double bond in indigo  
44 in a 1:1 molar ratio. The concentration of indigo before and after the reaction is measured using  
45 absorption spectrophotometry at a wavelength of 600 nm, and the ozone concentration is  
46 calculated. The indigo method has long supported ozone treatment, but has five limitations.

47 Firstly, the reaction of substances other than ozone, such as hypochlorous acid (Yoshizawa,  
48 2019), hypobromous acid (Penru et al., 2013) and permanganate ion (Gregory and Carlson, 2001),  
49 hydrogen peroxide (Bader and Hoigné, 1981), with the C–C double bond can result in an  
50 overestimation of ozone concentration. Considering that a large amount of hypobromous acid is  
51 produced during the ozonation of seawater, there is currently no method to selectively detect ozone  
52 concentration (Buchan et al., 2005) in seawater. Secondly, the purity of commercially available  
53 indigo is less than 80% and contains many impurities. It is reported that even with the same  
54 manufacturer, the measured molar absorption coefficients of indigo are different on different dates  
55 of manufacture (Gordon and Bubnis, 2002). Thus, calibration of molar absorption coefficients is  
56 needed, but it is cumbersome and beset with uncertainties because of the high volatility of  
57 dissolved ozone. In addition, some ozone treatment facilities do not have the capacity to perform  
58 this calibration. Thirdly, the stoichiometry of ozone and indigo is dependent on pH. If the pH of  
59 the solution is above 4 in any fraction during mixing, the amino group in indigo reacts with ozone  
60 and a fixed stoichiometry is not maintained (Bader and Hoigné, 1981). Fourth, the concentration  
61 of the indigo reagent and mixing ratio of the indigo solution to ozone water should be optimized  
62 according to the ozone concentration (Bader and Hoigné, 1981). Fifth is the effect of the  
63 absorbance of the sample measured at a wavelength of 600 nm. As ozone is effective in removing

64 color, it is often applied to highly colored water. Therefore, the dissolved ozone concentration is  
65 often measured with a high background absorbance at a wavelength of 600 nm. It may be possible  
66 to remove ozone separately with a reducing agent that does not interfere with absorbance at a  
67 wavelength of 600 nm and subtract it as background noise. However, this operation is cumbersome  
68 and introduces additional measurement errors.

69 To summarize, it is desirable to develop a novel measurement that overcomes these drawbacks  
70 to selectively, accurately, and conveniently measure dissolved ozone concentrations in various  
71 fields. Therefore, we propose a new method to measure ozone concentration using the reaction of  
72 *p*-vinylbenzoic acid (*p*-VBA) with ozone to produce *p*-formylbenzoic acid (*p*-FBA) in equal  
73 amounts to the reacted ozone.

74 It is expected that oxidants other than ozone react with *p*-VBA to produce little or no *p*-FBA. *p*-  
75 VBA (FUJIFILM Wako Pure Chemical Corporation, date accessed Jul 20 2021, n.d.) and *p*-FBA  
76 (Apollo Scientific Ltd, date accessed Jul 20 2021) are commercially available with guaranteed  
77 purities of 97% or higher and are suitable for quantification. *p*-VBA has no sites that react quickly  
78 with ozone, other than vinyl groups, regardless of the pH. *p*-FBA can be separated by liquid, ion,  
79 or gas (Santander et al., 2013) chromatography and detected with high sensitivity by ultraviolet  
80 (UV) and mass spectrometry (MS) detectors. *p*-FBA measurement, carried out using column  
81 separation, is not affected by coexisting substances. The *p*-VBA method differs significantly from  
82 the indigo method, as stated above, and its measurement error can also be significantly lower. In  
83 addition, it is common to measure the concentration not only of dissolved ozone but also of  
84 pollutants at any moment in ozone treatment experiments. According to the Globally Harmonized  
85 System of Classification and Labelling of Chemicals, *p*-VBA and *p*-FBA are relatively safe  
86 reagents. On the aspect that indigo method can be performed with a spectrophotometer, it is

87 superior to the *p*-VBA method. However, based on the price at the purchase for this paper  
88 (FUJIFILM Wako Pure Chemical Corporation; Tokyo Chemical Industry Co. , Ltd.), *p*-VBA is  
89 7,392 JPY per 5 g and indigo is 28,215 JPY per 5 g. Taking molar mass and reagent purity into  
90 account, the price of *p*-VBA is 1/22<sup>nd</sup> that of indigo. In the aspect of reagent cost, *p*-VBA method  
91 is less expensive than indigo method, even considering the purchase of *p*-FBA.

92 The structure of *p*-VBA was chosen, as the C–C double bond reacts quickly and  
93 stoichiometrically with ozone and a benzene ring is required for sensitive UV detection. It may  
94 seem possible to use a reaction in which vinylbenzene reacts with ozone to form benzaldehyde.  
95 However, benzaldehyde is a known byproduct of ozone (Richardson et al., 1999), and its structure  
96 needs to be specialized. Carboxy groups are easily ionized, enabling analysis with ion  
97 chromatography (IC) and contributing to higher sensitivity in liquid chromatography (LC)/MS  
98 analysis. Furthermore, carboxy groups improve hydrophilicity, reduce adsorption on experimental  
99 instruments and improve analytical precision. Therefore, the carboxyl group is considered a strong  
100 candidate for the specificity of vinylbenzene.

101 To make this study more comprehensive, vinylbenzenes other than *p*-VBA, *m*-vinylbenzoic acid  
102 (*m*-VBA), *o*-vinylbenzoic acid (*o*-VBA), *p*-chlorocinnamic acid, and *p*-bromocinnamic acid were  
103 examined for their suitability as reagents for ozone determination. The effect of persulfuric acid  
104 on the indigo and *p*-VBA methods was also examined (Zhou et al., 2019).

105 Although it is widely known that carbonyl is formed by the reaction of ozone with a C–C double  
106 bond, it has not been established as a quantitative method for ozone. This study demonstrates the  
107 robustness and advantages of the *p*-VBA method by various experiments and analyses.

## 108 **2. Methodology**

### 109 ***2.1 Environmental Water Samples***

110 As a representative of river water, a sample from the Katsura River in Kyoto Prefecture, Japan  
111 was used. As a representative example of wastewater, the secondary effluent from a sewage  
112 treatment plant in Kyoto prefecture, Japan was used. As a representative example of seawater,  
113 Osaka Bay coastal water in Osaka prefecture, Japan was used. The samples were filtered through  
114 1- $\mu\text{m}$  pore size glass microfiber filters (GF/B from Cytiva, Japan). The basic water qualities of the  
115 samples are listed in Table S1.

## 116 **2.2.Reagents**

117 The purity of all reagents in this study was determined by the supplier. The following were  
118 acquired from different sources: *p*-VBA (purity: 99%), *m*-formylbenzoic acid (*m*-FBA, purity:  
119 99.2%), *p*-chlorobenzaldehyde (purity: 99.8%), *p*-bromobenzaldehyde (purity: 99.2%) from  
120 FUJIFILM Wako Chemical Corporation, Japan; *p*-FBA (purity: 97%) from Apollo Scientific Ltd.,  
121 UK; indigo potassium salt (purity: 70.2%), *o*-formylbenzoic acid (*o*-FBA; purity: 99.8%), and *p*-  
122 bromocinnamic acid (purity: 99.9%) from Tokyo Chemical Industry Co., Ltd., Japan; *m*-VBA  
123 (purity: 95% < regarded as 97.5%) and *o*-VBA (purity: 99.7%) from Thermo Fisher Scientific Inc.,  
124 USA; *p*-Chlorocinnamic acid (purity: 99.8%) from Merck KGaA, Germany; the rest of the  
125 reagents were of analytical grade.

126 The hypobromous acid solution was adjusted by mixing the sodium bromide and hypochlorous  
127 acid solutions (Tokunaga et al., 2007). Ozone solution was prepared by blowing ozone gas into  
128 ice-cold or room temperature (20–25°C) ultrapure water or a 1 mM phosphate buffer solution (pH  
129 3). Ozone gas from an ozone generator (SGA-01-PAS2, Sumitomo Precision Products Co., Ltd.,  
130 Japan) was blown into water to remove NO<sub>x</sub> and dehumidified. The *p*-VBA solution was prepared  
131 by dissolving *p*-VBA in pure water or environmental water samples reacted with 50  $\mu\text{M}$  ozone  
132 (initial concentration) for more than 1 h, post which, the ozone was removed by nitrogen aeration.

133 At this pre-ozone treatment, the environmental water samples were diluted 1.07 times. The  
134 concentrations of all *p*-VBA solutions were 0.10 g L<sup>-1</sup>.

### 135 **2.3. Chemical analysis**

136 Absorbance was measured using a spectrophotometer (UV-1900i, SHIMADZU  
137 CORPORATION, Japan).

138 Dissolved oxygen was measured using a dissolved oxygen meter (Multi 3510 IDS, Xylem  
139 Analytics Germany Sales GmbH & Co. KG., Germany).

140 The dissolved ozone concentration was measured by direct absorption spectrophotometry (von  
141 Sonntag and von Gunten, 2012), for the indigo (Bader and Hoigné, 1981) and the *p*-VBA methods.  
142 For direct absorption spectrophotometry, a quartz cell with an optical path length of 1 cm was  
143 sealed immediately after syringe injection of ozone water three times the cell volume and measured  
144 at a wavelength of 260 nm (molar absorptivity coefficient: 3200 M<sup>-1</sup> cm<sup>-1</sup>) (von Sonntag and von  
145 Gunten, 2012). By performing the direct absorption spectrophotometry and the indigo method or  
146 a reaction with vinylbenzenes almost simultaneously, the molar absorption coefficient of indigo  
147 or the yield of benzaldehydes were determined based on the direct absorption spectrophotometry.  
148 The ozone water at this time was adjusted to a pH of 3 using phosphoric acid at room temperature,  
149 and the ozone in the syringe decreased at a rate of 0.07% per minute, which can be regarded as no  
150 decrease in the syringe over the timescale of the experiment. The volume of ozone water, before  
151 and after addition to volume, was determined by converting the difference in mass. The density  
152 used in the conversion was 1.02 kg L<sup>-1</sup> for seawater and 1.00 kg L<sup>-1</sup> for others.

153 The hypochlorite ion, hypobromite ion, and hydrogen peroxide concentrations were measured  
154 by absorption spectrophotometry at a wavelength of 292 nm (molar absorptivity: 350 M<sup>-1</sup> cm<sup>-1</sup>)  
155 (Soulard et al., 1981), 329 nm (molar absorptivity: 345 M<sup>-1</sup> cm<sup>-1</sup>) (Soulard et al., 1981) and 240  
156 nm (molar absorptivity: 39.4 M<sup>-1</sup> cm<sup>-1</sup>) (Nelson and Kiesow, 1972).

157 The concentrations of benzaldehydes and vinylbenzenes were measured using UPLC/UV or  
158 UPLC/MS [ACQUITY UPLC system with PDA  $\epsilon\lambda$  detector (optical path length 1 cm) or Xevo  
159 TQ MS detector from Waters Corporation, United States]. The MS detector was used only for  $m/z$   
160 measurements of precursor ions, and all quantitation was performed using a UV detector.  
161 Detection wavelengths and examples of detailed analysis conditions are shown in Tables S2, S4,  
162 S5 and S7.

## 163 **2.4. Experimental Method**

### 164 **2.4.1. Reaction rate measurements**

165 The rate constants of the reactions of  $p$ -VBA and  $m$ -VBA with ozone were determined by  
166 competition kinetics (von Sonntag and von Gunten, 2012) with CA and the pH was adjusted to 7.1  
167 with phosphate buffer or 1.0 with hydrochloric acid. On the time scale, phosphoric acid and  
168 hydrochloric acid did not react with  $p$ -VBA,  $m$ -VBA, and CA. The reaction rate of ozone with  $p$ -  
169 VBA or  $m$ -VBA ( $k_{O_3,VBA}$ ,  $M^{-1} s^{-1}$ ) was calculated using the reaction rate of ozone with CA  
170 ( $k_{O_3,CA}$ ,  $M^{-1} s^{-1}$ ), as shown in Eq. (1):

$$171 \quad k_{O_3,VBA} = \frac{\ln \frac{[VBA]_{\infty}}{[VBA]_0}}{\ln \frac{[CA]_{\infty}}{[CA]_0}} k_{O_3,CA} \quad (1)$$

172 where the subscript "0" and " $\infty$ " represent the concentration before and after the reaction, and  
173 [VBA] represents the concentration of  $p$ -VBA or  $m$ -VBA (M). The reaction rate constants of ozone  
174 with CA are  $1.2 \times 10^6$  or  $3.8 \times 10^5 M^{-1} s^{-1}$  in the completely dissociated state and  $1 \times 10^5$  or  $5 \times$   
175  $10^4 M^{-1} s^{-1}$  in the completely non-dissociated state (Leitzke et al., 2001; Lung, 1996). In addition,  
176 the  $pK_a$  values of  $p$ -VBA and CA were measured by potentiometric titration.

177 Reaction rate constants of *p*-FBA with ozone, hypochlorous acid, hypobromous acid and  
178 permanganate ion were measured. Details of the experimental method are described in Supporting  
179 Information.

#### 180 ***2.4.2. False positive ozone measurements in *p*-VBA and indigo methods***

181 Hypochlorite, hypobromite, potassium permanganate, hydrogen peroxide and persulfuric acid  
182 solutions were added to the stirred *p*-VBA solution and the amount of *p*-FBA formed after the  
183 reaction was measured. These oxidants were also added to the indigo solution and the amount of  
184 absorbance loss at a wavelength of 600 nm was measured. The molar ratio of each oxidant to *p*-  
185 VBA or indigo was 1:10. Indigo solutions containing malonic acid have also been used, as reports  
186 reveal that the use of malonic acid can mask false positives of ozone due to hypochlorous acid and  
187 hypobromous acid (Bader and Hoigné, 1981).

#### 188 ***2.4.3. Absorbance measurements of benzaldehydes and vinylbenzenes***

189 Molar absorption coefficients of *p*-FBA, *m*-FBA, and terephthalic acid (TA) were measured.  
190 The terephthalic acid is thought to be produced when *p*-FBA is oxidized. Absorbance was  
191 measured in a cell with 1 mm optical path length. In addition, the *p*-FBA solution containing 8 mg  
192 L<sup>-1</sup> dissolved oxygen was divided into glass tubes and stored in a dark room at 4°C, and their  
193 absorbances were measured at several days interval.

#### 194 ***2.4.4. Comparison of ozone concentrations in environmental water samples measured by the*** 195 ****p*-VBA and indigo methods***

196 Changes in the dissolved ozone concentration in the environmental water were measured using  
197 the *p*-VBA and indigo methods. Ozone water was injected into the environmental water  
198 maintained at 20°C and initial ozone concentration of the mixture was 50 µM. Ozone water was  
199 made from pure water and cooled on ice, and the environmental water samples were diluted 1.07

200 times by injection of ozone water. The stability of the *p*-FBA generated during the application of  
201 the *p*-VBA method to environmental water was evaluated.

#### 202 **2.4.5. Measurement error comparison method**

203 The formulae for calculating the ozone concentration ( $q_{\text{ind}}$  or  $q_{p\text{-VBA}}$ ,  $\mu\text{M}$ ) using the indigo  
204 method [Eq. (2)] and the *p*-VBA method [Eq. (3)] is as follows: For simplicity, it does not consider  
205 errors due to the background noise at 600 nm wavelength for the indigo method,

$$206 \quad q_{\text{ind}} = \frac{A_{\text{ind}}V_{\text{ind}} - A_{\text{mix}}V_{\text{mix}}}{\varepsilon(V_{\text{mix}} - V_{\text{ind}})} \times 10^6 \quad (2)$$

$$207 \quad q_{p\text{-VBA}} = \frac{C_{p\text{-FBA}}PV_{\text{mix}}}{Y(V_{\text{mix}} - V_{p\text{VBA}})} \quad (3)$$

208 where  $A_{\text{ind}}$  and  $A_{\text{mix}}$  are the absorbance (measured in a cell with a 1 cm optical path length) for  
209 the indigo solution and a mixture of indigo solution and water containing ozone (no unit).  $V_{\text{ind}}$ ,  
210  $V_{p\text{-VBA}}$ , and  $V_{\text{mix}}$  are the volumes of the indigo solution, *p*-VBA solution, and a mixture of ozone  
211 water and indigo solution or *p*-VBA solution (mL), respectively.  $\varepsilon$  is the molar absorption  
212 coefficient ( $\text{M}^{-1} \text{cm}^{-1}$ ). For this error comparison, a common  $\varepsilon$  of  $20000 \text{ M}^{-1} \text{cm}^{-1}$  was set  
213 (Bridgewater et al., 2012). The absorbance of the indigo solution was set (Bridgewater et al., 2012)  
214 to 2 and 0.2.  $C_{p\text{-FBA}}$  is the concentration of *p*-FBA ( $\mu\text{M}$ ) in the mixture of ozone water and *p*-VBA  
215 solution when the purity of the *p*-FBA reagent is 1.  $P$  is the purity of the *p*-FBA reagent and  $Y$  is  
216 the yield of *p*-FBA to ozone in the reaction between ozone and *p*-VBA.

217 When each measurement error ( $\Delta q_{\text{ind}}$  or  $\Delta q_{p\text{-VBA}}$ ,  $\mu\text{M}$ ) is assumed to be random and  
218 independent of each other, the errors of the indigo and *p*-VBA methods are as shown in Eqs. (4)  
219 and (5) (Taylor, 1996);  $\Delta$  represents the measurement error,  $\Delta V_{\text{ind}}$  and  $\Delta V_{p\text{-VBA}}$ , and were  
220 determined to be 0.08% using actual measurements of the pipettes (MPA-10000 from A&D  
221 COMPANY, LIMITED, Japan),

$$\Delta q_{\text{ind}} = \sqrt{\left(\frac{\partial q_{\text{ind}}}{\partial A_{\text{ind}}} \Delta A_{\text{ind}}\right)^2 + \left(\frac{\partial q_{\text{ind}}}{\partial A_{\text{mix}}} \Delta A_{\text{mix}}\right)^2 + \left(\frac{\partial q_{\text{ind}}}{\partial V_{\text{ind}}} \Delta V_{\text{ind}}\right)^2 + \left(\frac{\partial q_{\text{ind}}}{\partial V_{\text{mix}}} \Delta V_{\text{mix}}\right)^2 + \left(\frac{\partial q_{\text{ind}}}{\partial \varepsilon} \Delta \varepsilon\right)^2} \quad (4)$$

$$\Delta q_{p\text{-VBA}} = \sqrt{\left(\frac{\partial q_{p\text{-VBA}}}{\partial C_{p\text{-VBA}}} \Delta C_{p\text{FBA}}\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial P} \Delta P\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial V_{p\text{-VBA}}} \Delta V_{p\text{VBA}}\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial V_{\text{mix}}} \Delta V_{\text{mix}}\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial Y} \Delta Y\right)^2}$$

$\Delta V_{\text{mix}}$  was determined to be 1  $\mu\text{L}$  using actual balance measurements (HR-100A from A&D COMPANY, LIMITED, Japan).  $\Delta A_{\text{ind}}$  and  $\Delta A_{\text{mix}}$  were set because the transmittance had an error of 0.02% due to stray light (Sooväli et al., 2006) and the actual errors appeared larger than this because of other factors.  $\Delta \varepsilon$  was set (Gordon and Bubnis, 2002) to 5%.  $\Delta P$  was set (Malz and Jancke, 2005) at 1.5%.  $\Delta C_{p\text{FBA}}$  and  $\Delta Y$  values were determined from the experimental results of this study.

The measurement error was determined for three mixing ratios used for both the indigo and *p*-VBA methods. Five analyses were performed for each mixing ratio. The volume of the indigo and *p*-VBA solutions was set to 3 mL. Indigo solution with an absorbance of 2.05 and about 1.4 mg  $\text{L}^{-1}$  (Maximum:1.66, Minimum:1.04) ozone water was used; ozone concentration within each mixing ratio is the same. Experiments using ozone water at lower concentrations were not performed because ozone is unstable. The errors of the indigo and *p*-VBA methods were determined using Eqs. (6) and (7);

$$\Delta q_{\text{ind}} = \sqrt{\left(\frac{\partial q_{\text{ind}}}{\partial E_{\text{ind}}} \Delta E_{\text{ind}}\right)^2 + \left(\frac{\partial q_{\text{ind}}}{\partial \varepsilon} \Delta \varepsilon\right)^2}, \text{ where } E_{\text{ind}} = \frac{A_{\text{ind}}V_{\text{ind}} - A_{\text{mix}}V_{\text{mix}}}{V_{\text{mix}} - V_{\text{ind}}} \quad (6)$$

$$\Delta q_{p\text{-VBA}} = \sqrt{\left(\frac{\partial q_{p\text{-VBA}}}{\partial E_{p\text{-VBA}}} \Delta E_{p\text{-VBA}}\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial P} \Delta P\right)^2 + \left(\frac{\partial q_{p\text{-VBA}}}{\partial Y} \Delta Y\right)^2}, \text{ where } E_{p\text{-VBA}} = \frac{C_{p\text{-FBA}}V_{\text{mix}}}{V_{\text{mix}} - V_{p\text{VBA}}} \quad (7)$$

$\Delta E$  was set as half of the 95% confidence interval of the five repeated experiments divided by the mean, and the other error terms were the same as those described above. In the evaluation by

241 calculation, the errors of the indigo and *p*-VBA methods were compared using three representative  
242 examples of ozone concentrations of 1.4, 0.5, and 0.005 mg L<sup>-1</sup> with 1.4 mg L<sup>-1</sup> as the  
243 experimental setting.

### 244 **3. Results and discussion**

#### 245 **3.1. Reaction rate of ozone with *p*-VBA and *m*-VBA**

246 The p*K*<sub>a</sub> values of *p*-VBA and CA were determined to be 4.2 and 4.9 at 19–21°C. Thus, *p*-VBA  
247 and CA are regarded as completely dissociated at pH 7.1 and as completely non-dissociated at pH  
248 1.0. The reaction rate constant of ozone with *p*-VBA was 1.42 (95% CI: 1.41–1.44) times greater  
249 than that of ozone with CA at pH 7.1. The reaction rate constant of ozone with *p*-VBA was 8.2  
250 (95% CI: 7.8–8.7) times greater at pH 1.0. Therefore, the reaction rate constant for ozone and *p*-  
251 VBA were  $1.7 \times 10^6$  or  $5.4 \times 10^5$  M<sup>-1</sup> s<sup>-1</sup> in completely dissociated state and  $8 \times 10^5$  or  $4 \times 10^5$   
252 M<sup>-1</sup> s<sup>-1</sup> in completely non-dissociated state (Leitzke et al., 2001; Lung, 1996). In the subsequent  
253 experiment, the reaction rate constant was set to an average value of  $1.1 \times 10^6$  M<sup>-1</sup> s<sup>-1</sup> in the  
254 completely dissociated state. In addition, the following assumptions were made: the mixing ratio  
255 (volume ratio of *p*-VBA or indigo solution to ozone-containing water and *p*-VBA or indigo  
256 solution) of the *p*-VBA method is 0.3 (refer to discussion below), and the mixing is instantaneous.  
257 The ozone concentration drops to 1/10,000<sup>th</sup> within 40 ms.

258 To the best of our knowledge, the earliest concentration measurement after mixing ozone and  
259 sample water is after 115 ms in Buffle et al (Buffle et al., 2006). In this study, the sample water  
260 was the secondary treatment wastewater and had a relatively high DOC. In the study, the ozone  
261 concentration was approximately halved from 0 to 115 ms, so the average first-order reaction rate  
262 constant for ozone was 6 s<sup>-1</sup> if the decrease of ozone in this period is expressed as a first-order  
263 reaction. Since the rate of ozone disappearance decreases after the reaction, the reaction rate was  
264 less than 6 s<sup>-1</sup> at 115 ms sampling point. If the mixing ratio is 0.3, the ozone disappearance rate in

265 the *p*-VBA method is  $740 \text{ s}^{-1}$ . Thus, the consumption of ozone by the sample-derived material is  
266 less than 0.8%, which is an acceptable value for most experiments. Therefore, even though the  
267 reaction rate of *p*-VBA with ozone is slower than that of indigo ( $9.4 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$ ) (Muñoz and  
268 von Sonntag, 2000), it is sufficiently fast. The reaction rate constant of ozone with *m*-VBA was  
269 1.3 (95% CI: 1.2-1.4) times greater than ozone with CA at pH 7.1 and was sufficiently fast.

270 The second order reaction rate constants of *p*-FBA with ozone, hypochlorite ion, hypobromite  
271 ion and permanganate were less than 7, 0.2, 0.2 and  $0.7 \text{ M}^{-1} \text{ s}^{-1}$  at pH 7. Therefore, *p*-FBA is stable  
272 against these oxidants.

### 273 **3.2. Yields of benzaldehydes to ozone**

274 The ratio of benzaldehyde produced to consumed ozone is shown in Figure 1. The yield of *p*-  
275 FBA is 1 whether *p*-VBA is fully dissociated (pH 7) or fully non-dissociated (pH 1.7). The yield  
276 of *p*-FBA was maintained at 1, even in the presence of coexisting substances in environmental  
277 water samples. Within experiments using *p*-VBA solution at pH 7 with phosphate buffer, it has  
278 been confirmed that 1 mol of ozone reacts with 1.01 (95% CI: 0.99–1.03) mol of *p*-VBA in the  
279 stoichiometry. The data presented in Figure S1 show that the photodiode array-detected peak  
280 which increased after the reaction was the only *p*-FBA peak observed. Different researchers have  
281 reported different values ( $2900\text{--}3600 \text{ M}^{-1} \text{ cm}^{-1}$  at 258–260 nm) for the molar absorption  
282 coefficient of ozone in direct spectroscopy (Bader, 1982; Hart et al., 1983; Kilpatrick et al., 1956;  
283 Taube, 1957). Conversely, based on the assumption that 1 mol of ozone reacts with *p*-VBA to  
284 produce 1 mol of *p*-FBA, this study supports the molar absorption coefficient of  $3200 \text{ M}^{-1} \text{ cm}^{-1}$   
285 used here. Therefore,  $\Delta Y$  in Eq. (5) was set to 0. Unlike phosphoric acid, ammonium carbonate is  
286 a volatile buffer solution suitable for LC/MS analysis. However, ammonia reacted with ozone at a  
287 moderate rate (Garland et al., 1980; Hoigné and Bader, 1983) (reaction rate constant: 20 or  $44 \text{ M}^{-1}$   
288  $\text{s}^{-1}$ ). If the reaction rate between ammonia and ozone is  $44 \text{ M}^{-1} \text{ s}^{-1}$ , and with the use of  $0.1 \text{ g L}^{-1}$

289 *p*-VBA solution, the maximum concentration of ammonium carbonate is 9 mM when the amount  
290 of ammonia that reacts with ozone is less than 1/1000<sup>th</sup> of the amount of *p*-VBA. Unlike the indigo  
291 method, the *p*-VBA method is independent of pH. The recommended ammonium carbonate  
292 concentration is 9 mM or less in the *p*-VBA method.

293 The data presented in Figure S1 show no new peaks other than benzaldehyde from solutions in  
294 which vinylbenzenes other than *o*-VBA reacted with ozone and the yield of benzaldehydes from  
295 vinylbenzenes other than *o*-VBA is considered 1. In addition, Figure S1 depicts a new peak  
296 different from *o*-FBA in the solution in which *o*-VBA and ozone reacted. Thus, *o*-FBA formation  
297 competes with at least one other substance and is likely to be sensitive to reaction conditions  
298 (temperature, pH, etc.). Therefore, in terms of benzaldehydes yield, vinylbenzenes other than *o*-  
299 VBA are suitable reagents for ozone determination.

### 300 **3.3. False positive ozone in *p*-VBA and indigo methods**

301 The false-positive ozone (mol) to hypochlorous acid, hypobromous acid, hydrogen peroxide and  
302 persulfuric acid (mol) in the indigo method and hypochlorous acid in the *p*-VBA method are shown  
303 in Figure 2. False-positive ozone (mol) to hydrogen peroxide in the *p*-VBA method was below the  
304 lower limit of quantification and less than 10<sup>-4</sup> mol in 4 h.

305 Formation of *p*-FBA by hypobromous acid and potassium permanganate in *p*-VBA method were  
306 completed within minutes after the reaction. When the *p*-VBA solution at pH 7 contained 50 mM  
307 phosphoric acid and 7 mM ammonium carbonate, false-positive ozone (mol) to hypobromous acid  
308 and potassium permanganate was 0.013 (95%CI: 0.006-0.014) and 0.017 (95%CI: 0.016-0.018)  
309 mol. When the *p*-VBA solution at pH 7 only contained 50 mM phosphoric acid, false-positive  
310 ozone (mol) to hypobromous acid and potassium permanganate was 0.0007 (95%CI: 0.0005–  
311 0.0008) and 0.012 (95%CI: 0.011–0.013) mol. The false-positive ozone was higher in the presence

312 of ammonium carbonate. Therefore, the use of ammonium carbonate is undesirable when  
313 hypobromous acid or permanganate ions are present.

314 The false positive ozone (mol) to permanganate ion in the indigo method was 1.5 mol and  
315 decolorization of indigo was completed within minutes after the reaction. In a previous study, 1  
316 mol of permanganate ion decolorized 1.8 mol indigo (Rakness et al., 2010), which is different from  
317 the findings of our study. This difference may be attributed to the quality and quantity of the  
318 impurities in the indigo reagent. The stoichiometry of permanganate ions, as well as other oxidants  
319 and indigo, may vary depending on the source and date of acquisition of the indigo reagent.  
320 However, the fact that the *p*-VBA method is less likely to falsely detect ozone than the indigo for  
321 all oxidants is considered irrefutable.

322 The false positive ozone (mol) to persulfuric acid in the indigo method was below the lower  
323 limit of quantitation. For persulfuric acid, the indigo method had the fewer false positive than the  
324 *p*-VBA method. However, the amount of false positives in the *p*-VBA method is not large if  
325 analyzed immediately. The method for measuring ozone concentration when persulfuric acid is  
326 coexisting should be determined by comprehensively considering the other advantages of the *p*-  
327 VBA method.

328 The mass scan (range: 30–300 *m/z*) of a solution of *p*-VBA reacting with the  
329 permanganate ion presented a main peak of 179 *m/z*, which is two OH added to the C–C double  
330 bond of *p*-VBA. The mass scan (range: 30–300 *m/z*) of a solution of *p*-VBA reacted with  
331 hypobromous acid and exhibited a main peak of 163, 243, and 245 *m/z*, which is OH and Br  
332 added to the C–C double bond of *p*-VBA. The chromatogram data are shown in Figure S2. The  
333 product of reaction between *p*-VBA and hypochlorous acid was also not identified.

### 334 **3.4. Molar absorption coefficients**

335 The molar absorption coefficients of *p*-FBA, *p*-VBA, *m*-FBA, *m*-VBA, and TA are shown in  
336 Figure S3 and listed in Table S6. Ozone dissolved in ultrapure water reacted with the *p*-VBA  
337 solution at pH 7, and the absorbance of the solution after the reaction at wavelengths of 233 and  
338 279 nm was measured by direct absorption spectrophotometry. Further, the concentration of *p*-  
339 FBA (M) was calculated according to Eqs. (8) and (9), and the concentration of ozone was  
340 determined; the ratio of the value measured by this method to that measured by direct absorption  
341 spectrophotometry was 99% (95% CI: 95–104),

$$\varepsilon_{p\text{-FBA},233} [p\text{-FBA}] + \varepsilon_{p\text{-VBA},233} [p\text{-VBA}] = \frac{\text{Abs}_{233}}{c}, \quad (8)$$

$$\varepsilon_{p\text{-FBA},279} [p\text{-FBA}] + \varepsilon_{p\text{-VBA},279} [p\text{-VBA}] = \frac{\text{Abs}_{279}}{c}, \quad (9)$$

343 Where  $\varepsilon$  is the molar absorption coefficient ( $\text{M}^{-1} \text{cm}^{-1}$ ) at wavelength of *p*-VBA or *p*-FBA, Abs  
344 is the absorbance measured in a 1 cm cell at a wavelength (no unit), and  $c$  is the optical path length  
345 (cm). The wavelength of 233 nm is the wavelength at which the ratio of the molar absorption  
346 coefficients of *p*-FBA to those of *p*-VBA are large and wavelength of 279 nm is the opposite.  
347 Cooled highly-concentration ozone water adjusted with pure water or a phosphoric acid solution  
348 is often used in ozone treatment research. The ozone concentration in such water can be measured  
349 easily with a spectrophotometer using this method.

350 The largest difference ( $1.39 \times 10^4 \text{M}^{-1} \text{cm}^{-1}$ ) between the molar absorption coefficients of *p*-  
351 FBA and TA was observed at a wavelength of 262 nm in the non-dissociated state. The change in  
352 absorbance at 262 nm of the *p*-FBA solution stored in a dark room at 4°C was within a narrow  
353 range of 0.5767–0.5788 over 29-d (Figure S3); the absorbance measurements were conducted at a  
354 temperature of 25°C and pH of 1. If 1% *p*-FBA was changed to TA, the absorbance decreased by  
355 0.0048. In addition, no change in the absorbance was observed at other wavelengths (205–315 nm),  
356 suggesting that *p*-FBA did not change to terephthalic acid or other compounds within a month,

357 even when the *p*-FBA solution contained 8 mg L<sup>-1</sup> of dissolved oxygen. Therefore, *p*-FBA is an  
358 excellent quantitative reagent that is stable over a long period of time. On the contrary, *p*-FBA of  
359 about 5 µg L<sup>-1</sup> was detected in the *p*-VBA solution about a week after adjustment. Therefore, it  
360 is needed to measure the *p*-FBA concentration in the *p*-VBA solution each time and correct the  
361 amount of *p*-FBA produced, particularly when measuring low ozone concentrations.

362 The maximum absorption wavelengths of *p*-VBA and *m*-VBA are very different. When UV  
363 detectors are used for analyses, it will be possible to improve the measurement sensitivity by proper  
364 use of *p*-VBA and *m*-VBA.

### 365 ***3.5. Comparison of ozone concentrations in environmental water samples measured by the p-*** 366 ***VBA and indigo methods***

367 The molar absorption coefficient of indigo at 600 nm was determined to be  $2.4 \times 10^4$  (95% CI:  
368  $2.24\text{--}2.55 \times 10^4$ ) M<sup>-1</sup> cm<sup>-1</sup>. Figure 3 shows the dissolved ozone concentration measured by the  
369 indigo and *p*-VBA methods in environmental water. In the river water, the measurements of the *p*-  
370 VBA method tended to fall below those of the indigo method over time. In treated wastewater, the  
371 measurements of the *p*-VBA and indigo methods were almost the same. In seawater, there was a  
372 significant difference between the measurements obtained using the *p*-VBA and indigo methods.  
373 The reaction rate constant between bromide ions and ozone at 20°C has been reported (Haag and  
374 Hoigne, 1983) to be 160 M<sup>-1</sup> s<sup>-1</sup>. Thus, even considering ozone consumption by only bromide ions,  
375 the ozone concentration dropped to 1.4 µM after 30 seconds, the first measurement point. This  
376 condition was satisfied by the measurements obtained by the *p*-VBA method but not by the indigo  
377 method. *p*-FBA above 1 µg L<sup>-1</sup> was not detected in solutions of river water, treated wastewater,  
378 or seawater after ozonation. If the ozonated water sample contains *p*-FBA, the ozone concentration  
379 measurements should be corrected.

380 Measurements of the *p*-FBA concentration produced when the *p*-VBA solution was mixed with  
381 environmental water containing ozone over an 18-h period showed no change in *p*-FBA  
382 concentration (Figure S4). Thus, *p*-FBA concentrations were shown to be stable in environmental  
383 water after ozone treatment. Therefore, the *p*-VBA method is also appropriate for situations in  
384 which the analysis cannot be performed immediately.

385 The calibration curve for *p*-FBA had good linearity ( $R^2 = 0.998$ ,  $1/x$  weighting) in the range of  
386  $1 \mu\text{g L}^{-1}$  to  $5 \text{mg L}^{-1}$ , with each concentration analyzed eight times. The details of the analytical  
387 conditions are shown (Table S7). Half of the 95% CI divided by the mean was 5.9% for  $1.020 \mu\text{g}$   
388  $\text{L}^{-1}$ , 2.9% for  $5.262 \mu\text{g L}^{-1}$ , 0.75% for  $51.71 \mu\text{g L}^{-1}$ , 0.43% for  $509.6 \mu\text{g L}^{-1}$ , and 0.61% for  $4.974$   
389  $\text{mg L}^{-1}$ . These results were plotted on a graph with the *p*-FBA concentration on the horizontal axis  
390 and the errors on the vertical axis, with a straight line connecting each plot.  $\Delta C_{pVBA}$  is defined  
391 using these straight lines. The linearity of the calibration curve was lost at about  $5 \mu\text{g L}^{-1}$  when  
392 methanol was not added to the sample for determination of chlorobenzaldehyde and  
393 bromobenzaldehyde. However, the use of methanol is undesirable because it is cumbersome and  
394 costly. As there is linearity in the calibration curve of *p*-VBA up to  $1.020 \mu\text{g L}^{-1}$  in an aqueous  
395 solution, the enhancement of water solubility by the carboxyl group plays an important role in  
396 improving the accuracy and convenience of the analysis. *p*-VBA is a better reagent for ozone  
397 concentration determination than *p*-chlorocinnamic acid or *p*-bromocinnamic acid.

### 398 ***3.6. Comparison of measurement error between p-VBA and indigo methods***

399 The molar absorption coefficient of indigo determined in this study was 20% larger than what  
400 was considered common ( $20000 \text{M}^{-1} \text{cm}^{-1}$ ), confirming that the indigo method needed calibration.  
401 In addition, the relative error of the measurement was 7%, which was larger than the setting value

402 of 5% for comparison of measurement error. This result demonstrates that the error by the indigo  
403 method is underestimated for the following comparison.

404 The comparison of the relative errors between the indigo and *p*-VBA methods is shown in Figure  
405 4 (a). The *p*-VBA method presented a smaller relative error than the indigo method for any mixing  
406 ratio for both the calculated and measured values. When the measurement error of indigo's molar  
407 absorption coefficient is larger than 5% or when there is background noise at 600 nm wavelength,  
408 the error of the indigo method is bigger. In the indigo method, when the absorbance and mixing  
409 ratio were not properly adjusted according to the ozone concentration, the error was large. Hence,  
410 reduction of error in the indigo method involves determination of the mixing ratio and absorbance  
411 of the indigo solution in the pre-measurement, which is complicated. On the contrary, *p*-VBA  
412 method is convenient because a small measurement error was achieved with one *p*-VBA solution  
413 (for example, 0.1 g L<sup>-1</sup>) and one mixing ratio (for example, 0.3), regardless of the ozone  
414 concentration to be measured. In addition, when measuring substances (e.g. pollutants) in a  
415 mixture of indigo or *p*-VBA solution and ozone-containing environmental water by LC or IC, the  
416 *p*-VBA method, which can fix the reagent concentration and mixing ratio, provides a more  
417 consistent baseline among different samples.

418 The error in volume measurement ( $\Delta V_{\text{ind}}$ ,  $\Delta V_{\text{p-VBA}}$  and  $\Delta V_{\text{mix}}$ ) in this study was very small, and  
419 such a small error may not be achievable in more complex experiments or outdoor surveys. The  
420 results of the relative error, for the error in the volume 10 times larger ( $\Delta V_{\text{ind}}$  and  $\Delta V_{\text{p-VBA}} = 0.8\%$   
421 and  $\Delta V_{\text{mix}} = 1 \mu\text{L}$ ), are shown in Figure 4 (b). The *p*-VBA method is always more accurate than  
422 the indigo method. Regarding the *p*-VBA method, the tendency of the error to increase with  
423 increasing mixing ratio becomes more pronounced. The relative error is almost the minimum  
424 below a mixing ratio of about 0.3 at any ozone concentration. However, a mixing ratio that is too

425 low may result in *p*-VBA not being distributed throughout the mixture, especially with insufficient  
426 agitation. Therefore, it is recommended that the *p*-VBA method be performed with a mixing ratio  
427 of about 0.3. When the mixing ratio is fixed at 0.3, the change in relative error with respect to  
428 changes in dissolved ozone concentration is shown in Figure S5.

429 In the *p*-VBA method, the measurements impacting the error in the measured ozone  
430 concentration at a mixing ratio below 0.75 consisted of purity of the *p*-FBA reagent or the  
431 concentration of *p*-FBA. Uncertainty of the purity is dominant when ozone concentration is high  
432 and uncertainty of the *p*-FBA concentration becomes dominant as ozone concentration decreases,  
433 and under the study conditions, the magnitude of  $\frac{\partial q_{p-VBA}}{\partial P} \Delta P$  and  $\frac{\partial q_{p-VBA}}{\partial C_{p-FBA}} \Delta C_{p-FBA}$  in Eq. (5) matched  
434 at  $0.02 \text{ mg L}^{-1}$ . Therefore, when ozone concentrations are low, multiple repetitions of the *p*-FBA  
435 concentration analysis are effective in reducing measurement errors. In the indigo method, the  
436 measurements mainly impacting the error are not volumetric measurements of mixtures of indigo  
437 solution and ozone water. Detailed figure of the error factors and their magnitude comparisons are  
438 shown in Figure S6. Assuming a relative error of 10% as the lower limit, the lower limit of  
439 quantitation for ozone in this calculation was  $4 \text{ } \mu\text{g L}^{-1}$  for the indigo method (using indigo solution  
440 with absorbance of 0.2) and  $0.4 \text{ } \mu\text{g L}^{-1}$  for the *p*-VBA method, with the *p*-VBA method being  
441 smaller.

#### 442 **4. Conclusion**

443 The yield of *p*-FBA from the reaction of *p*-VBA and ozone was maintained at 1, even in the  
444 presence of coexisting substances in environmental water, and the *p*-VBA method can be used for  
445 the determination of dissolved ozone concentration.

446 The *p*-VBA method is more selective than the indigo method, and is more resistant to  
447 hypochlorous acid, hypobromous acid, permanganate and hydrogen peroxide.

448 To reduce the error, the indigo method requires adjustment of the concentration of the indigo  
449 solution and the mixing ratio according to the ozone concentration. On the contrary, the *p*-VBA  
450 method can measure a wide range of ozone concentrations with one concentration of the *p*-VBA  
451 solution and mixing ratio with less error than the indigo method.

452 As a reagent for ozone concentration determination, although fewer evaluations have been  
453 performed on *m*-VBA than on *p*-VBA in this study, *m*-VBA is considered a possible substitute for  
454 *p*-VBA. The most significant difference between *p*-VBA and *m*-VBA is the UV spectrum. It will  
455 be possible to improve the measurement sensitivity by proper use of *p*-VBA and *m*-VBA when  
456 UV detectors are used for analyses. *o*-VBA is inferior to *p*-VBA, because the yield of *o*-FBA to  
457 ozone is not 1. *p*-chlorocinnamic acid and *p*-bromocinnamic acid are inferior to *p*-VBA because  
458 the calibration curves for *p*-chlorobenzaldehyde and *p*-bromobenzaldehyde produced in the  
459 reaction with ozone show poor linearity.

460 Compared to the indigo method, the *p*-VBA method is more selective, more accurate, and can  
461 be performed simultaneously with measurement of other analytes and does not require  
462 cumbersome calibration. *p*-VBA can contribute widely to ozone treatment research and  
463 management.

464

## 465 **Author Contributions**

466 Kohei Kawaguchi conceived the study, analyzed the data, and wrote the manuscript. Taira  
467 Hidaka made important revisions to the manuscript. Taku Fujiwara supervised the study and  
468 made important revisions to the manuscript. The manuscript is contributed by all the authors. All  
469 the authors approved the final version of the manuscript.

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## 479 **ABBREVIATIONS**

480 *p*-FBA, *p*-formylbenzoic acid; *p*-VBA, *p*-vinylbenzoic acid; *m*-FBA, *m*-formylbenzoic acid; *m*-  
481 VBA, *m*-vinylbenzoic acid; *o*-FBA, *o*-formylbenzoic acid; *o*-VBA, *o*-vinylbenzoic acid; TA,  
482 terephthalic acid; CA, cinnamic acid; indigo, indigotrisulfonic acid; mixing ratio, volume ratio of  
483 *p*-VBA or indigo solution to the sum of ozone-containing water and *p*-VBA or indigo solution;  
484 UV, ultraviolet; MS, mass spectrometry; IC, ion chromatography; LC, liquid chromatography.

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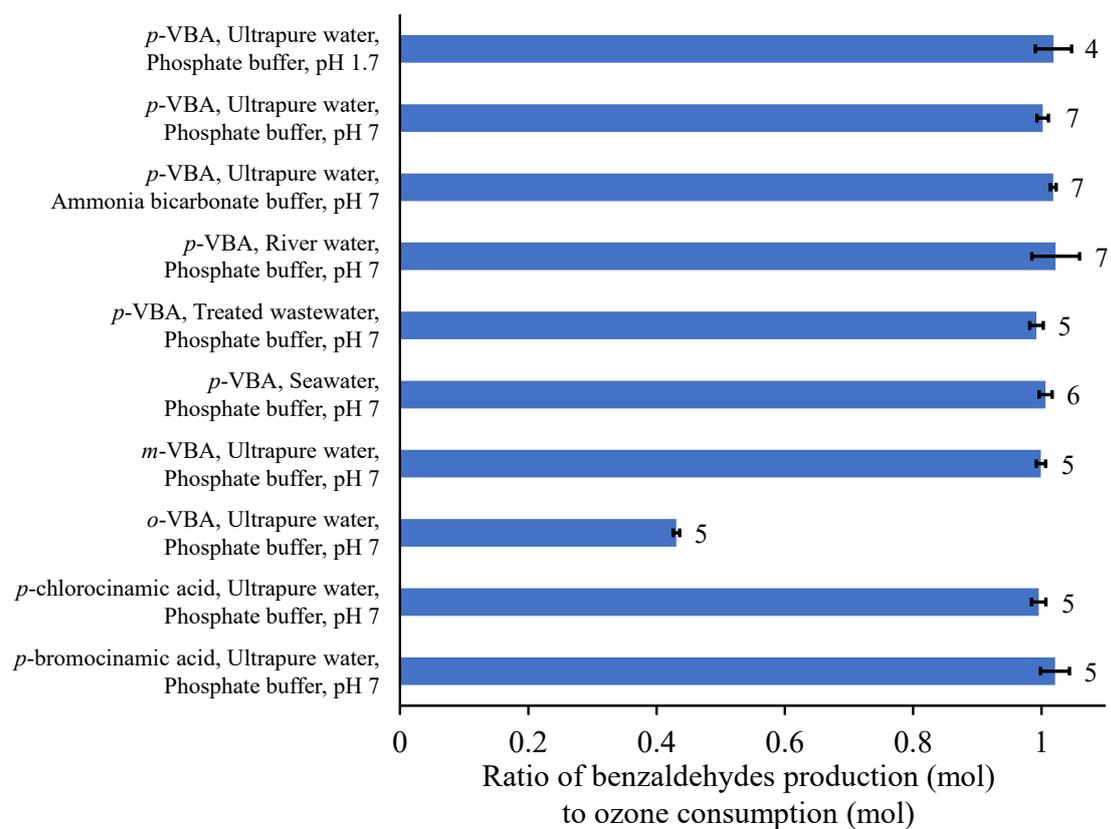
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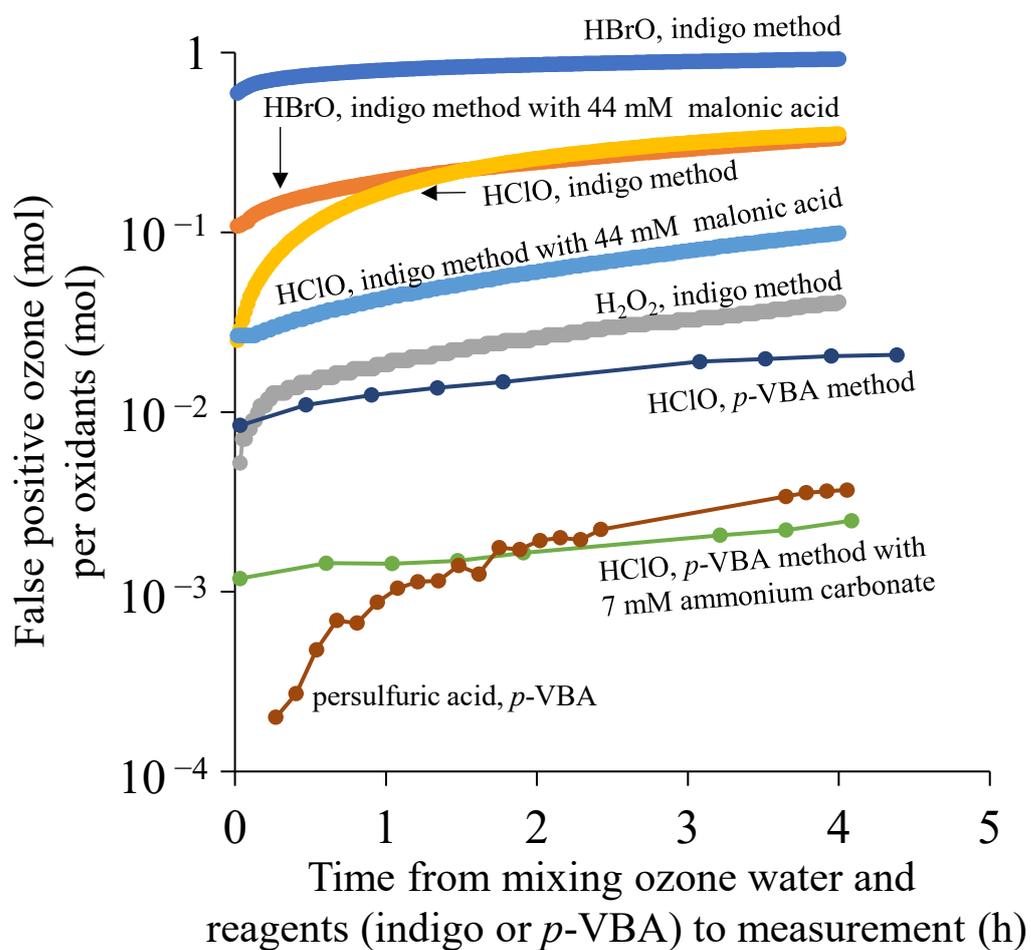
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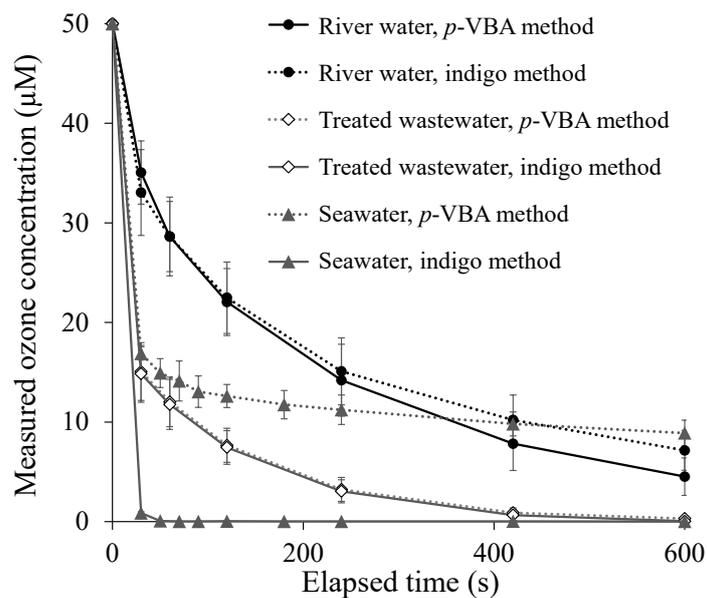
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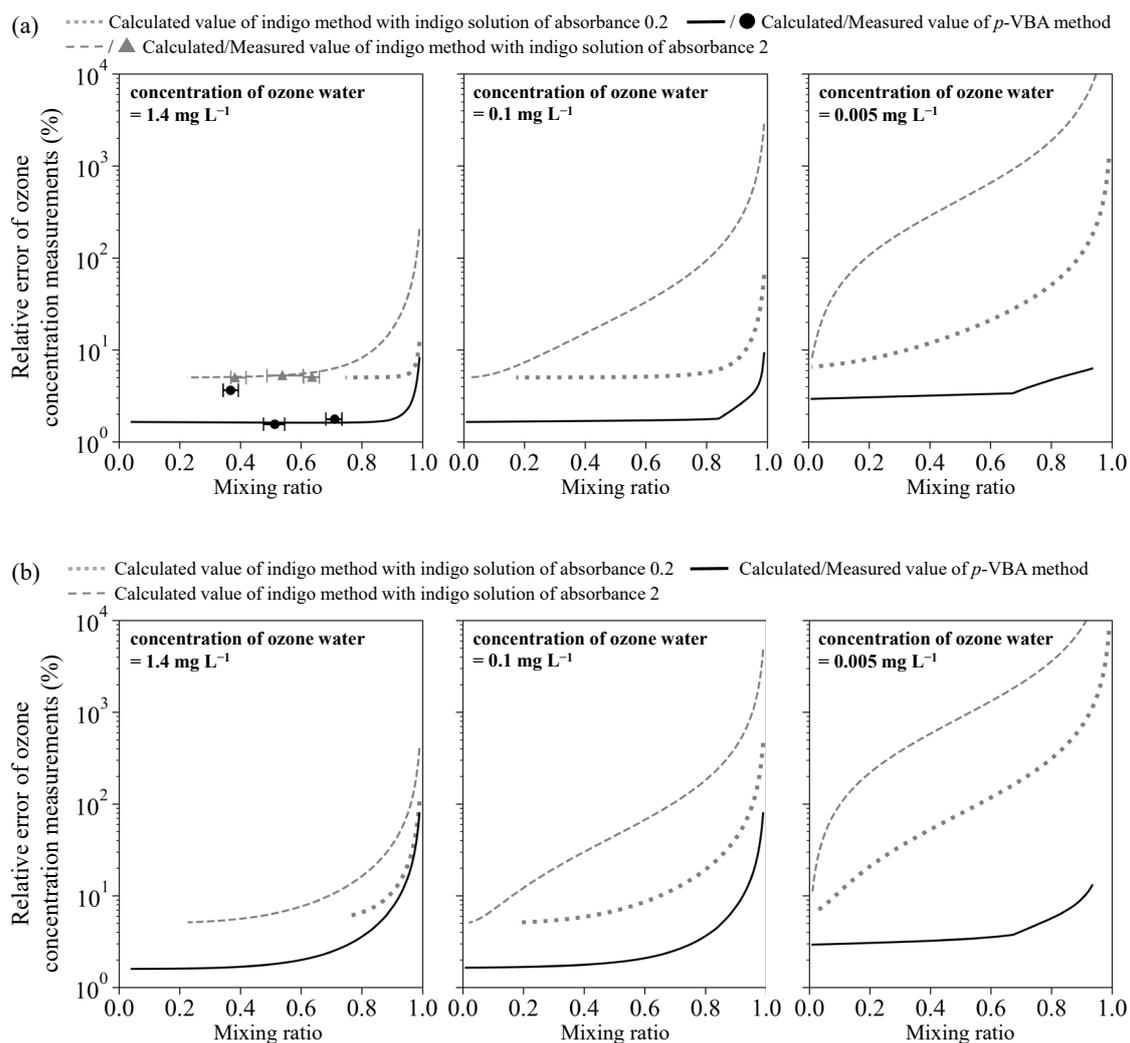
**Figure 1.** Yields of benzaldehydes from the reaction of ozone with vinylbenzenes solutions. The legend is written in the order of vinylbenzenes, solvent, buffer, and pH. Error bars indicate 95% confidence intervals. The number to the right of the error bars is the number of repetitions of the experiment



**Figure 2.** False positive ozone (mol) in *p*-VBA and indigo methods for hypochlorous acid, hypobromous acid, hydrogen peroxide and persulfuric acid (mol). All *p*-VBA solutions contained 50 mM phosphate buffer and were adjusted to pH 7.



**Figure 3.** Measurements of the dissolved ozone concentration measured by the indigo and *p*-VBA methods in environmental water. The initial value is not an actual measured value, but a set value determined by the dilution rate and the concentration of ozone water to be injected. Measurements were taken twice, and the average value was plotted. Error bars indicate maximum and minimum values for two measurements



**Figure 4.** Comparison of the relative errors between the indigo method and the *p*-VBA method. Relative errors were calculated by Eq. (4)–(7). For calculated values, the following were excluded: Amount of substance in indigo or *p*-VBA is less than the amount of substance in ozone or the concentration of *p*-FBA exceeds the error-evaluated range ( $1.020 \mu\text{g L}^{-1}$  to  $4.974 \text{mg L}^{-1}$ ). The absorbance of indigo was the value when measured in a 1 cm cell. (a) Result of analysis to reproduce experimental conditions. Error bars indicate maximum and minimum values in the measurement. (b) Result of analysis when the measurement error of the volume is 10 times larger than (a)